Appendix 4: Green Island Resource Recovery Park Precinct - Groundwater Technical Assessment



DCC – Resource Recovery Park Precinct

Groundwater Technical Assessment

Dunedin City Council 29 February 2024

→ The Power of Commitment





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Appendices

Appendix A Technical groundwater assessment for landfill closure consent

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1. Introduction

1.1 Waste Futures Programme

As part of Dunedin's wider commitment to reducing carbon emissions and reducing waste going to landfill, the Dunedin City Council (DCC) has embarked on the Waste Futures Programme to develop an improved comprehensive waste management and diverted material system for Ōtepoti Dunedin. The programme aligns with DCC's responsibility under the Waste Minimisation Act 2008 to 'promote effective and efficient waste management and minimisation within its district'.

Improving Dunedin's whole waste system includes enhancing collection services for reuse and recycling, and safe disposal of residual waste to landfill.

The Waste Futures Programme includes provision of an enhanced kerbside recycling and waste collection service for Dunedin from July 2024. The new kerbside collection service will include collection of food and green (organic) waste.

To support the implementation of the new kerbside collection service, the DCC is planning to make changes to the use of Green Island landfill site (Figure 1.1) in coming years including:

- Developing an improved Resource Recovery Park Precinct (RRPP) for food and green waste and to process recycling; and
- Providing new waste transfer facilities to enable the safe disposal of any residual waste to landfill.



Figure 1.1 Green Island Landfill and Resource Recovery Park Precinct Site (Designation D658)

In addition, DCC is planning for the ongoing operation and closure of the Green Island landfill, which is coming to the end of its operational life. The existing Otago Regional Council (ORC) resource consents, required to operate a landfill at Green Island, expired in October 2023. In March 2023, DCC applied to ORC for replacement resource consents to continue to use the landfill until it closes completely, and waste

disposal can be transferred to a new landfill facility. These consent applications are in the process of being considered by ORC.

1.2 Green Island Resource Recovery Park Precinct

To meet the requirements of the new kerbside collection service the DCC is investing in improvements and expansion to the existing resource recovery area at Green Island landfill site. Proposed new facilities are shown on Figure 1.2 and include:

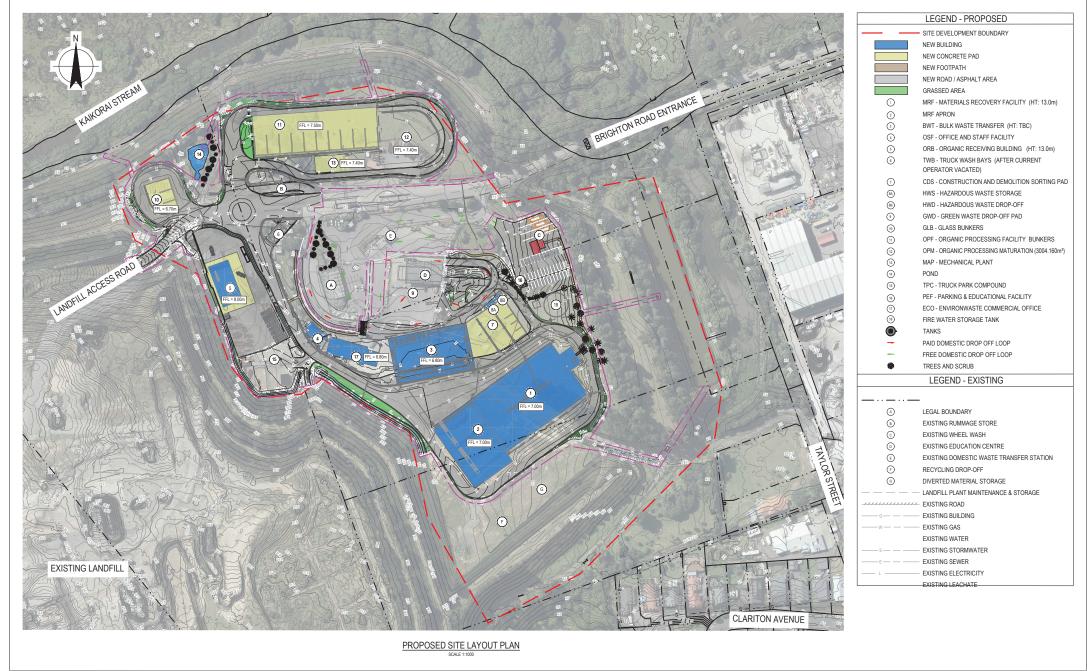
- organic receivals building (ORB) and processing facilities to support the organic waste kerbside collection;
- materials recovery facility (MRF) to sort and bale items collected from kerbside mixed recycling bins;
 and
- bulk waste transfer station (BWTS) to facilitate the compaction and trucking of waste to landfill.

Additional facilities also include new glass bunkers, staff offices, parking, and breakrooms and associated access roads and truck parking areas. Several existing facilities are to be retained including the rummage shop, public drop-off areas and the education centre. The development will result in an increase in the area covered by hardstand or buildings compared to the current facility.

The resource consents for the development and operation of the new facilities relate to ground disturbance, and discharges to land and air. The Green Island landfill site is subject to an operative designation (D658) in the Proposed Second-Generation Dunedin City District Plan (2GP) for the purpose of Landfilling and Associated Refuse Processing Operations and Activities.

The RRPP will be run by EnviroNZ on behalf of DCC and will start operating in July 2024 following construction of the ORB, which is currently underway. Resource consent to operate the ORB was granted by ORC in September 2023 under the existing landfill consents.

The other new RRPP facilities are planned to start operating from mid to late 2025.



Author R. GERONIMO Drafting Check G. DOUGHERTY* Designer K. ATKINS Design Check M. ROUJON* Plot Date: 20 October 2023 - 12:51 pm Plotted by: Greg Dougherty













Status PRELIMINARY

Drawing PROPOSED SITE LAYOUT PLAN

Status S3

1.3 Purpose of this report

The purpose of this report is as follows:

- Provide a description of the groundwater environment, existing management of groundwater and leachate discharges, and environmental monitoring data;
- Provide a technical assessment of the potential effects on groundwater and connected surface water flows from the proposed RRPP;
- Provide an assessment of the effects of leachate on groundwater quality; and
- Provide recommendations for mitigation measures to minimise the effects on the environment and monitoring conditions to confirm the effectiveness of the recommended mitigation measures.

Note: the potential effects of site stormwater on downstream surface water flows and quality are discussed in the GHD Waste Futures – Green Island – Resource Recovery Park Precinct Stormwater Management and Assessment of Effects (GHD 2024A) and referenced in this report.

A groundwater technical assessment was completed to support the consent application for the extension and closure of the Green Island landfill (GHD 2023A). Much of the information presented in the landfill groundwater assessment is applicable to the RRPP area. Rather than replicate the information presented in the previous assessment, only key information is summarised in this RRPP technical report with the landfill closure groundwater assessment included in Appendix A.

This report should be read in conjunction with the following reports:

- Green Island Resource Recovery Park Precinct Stormwater Management and Assessment of Effects (GHD, 2024A).
- Waste Futures Green Island Landfill Closure Groundwater Technical Assessment (GHD 2023A and included as Appendix A)
- Green Island Resource Recovery Park Precinct Design and Operations Report (GHD, 2024B)
- Green Island Resource Recovery Park Precinct DRAFT Contaminated Land Management Plan (GHD, 2024C)

These reports provide supporting information and context which the groundwater assessment relies upon. Where appropriate, a summary of critical information is provided in this report with cross- references to the relevant technical report.

In addition, the assessment undertaken in this report is based on a review of previous investigations, including those undertaken as part of the 1994 resource consent application. This information has been supplemented with additional site investigations undertaken by GHD to support the design and consenting process (documented in the *Waste Futures – Green Island Landfill Extension – 2022 Geotechnical Investigation Factual Report* (GHD 2023B).

1.4 Current Consents

The operation of the Green Island landfill, including associated waste processing operations and facilities, is currently subject to 14 existing resource consents granted by the ORC. The consents cover landfill operation activities relating to discharges to land, water, and air, taking and/or diverting water, and disturbance of a contaminated site. All consents expire on 1 October 2023.

The current consents limit the extent of landfilling through the combination of a maximum 38 ha landfill footprint, conditions limiting the deposit of waste to 270 m³/day and 100,000 m³/year¹, and the end date of 1 October 2023. The proposed RRPP reorganises and relocates existing waste management facilities and improves on existing processes. Some RRPP activities will be covered by existing consents, but the remainder will require new consents.

2. Environmental Setting

2.1 Introduction

A review of the environmental setting was undertaken to inform the technical assessment (Section 3). As noted in Section 1.2 only key information is summarised in the following sections, with further information provided in the Green Island landfill closure groundwater technical assessment included in Appendix A.

2.2 Site Description

The Green Island landfill site is located in the suburb of Green Island, approximately 8.8 km by road southwest of Central Dunedin. The landfill site comprises a total area of 75.6 ha. The site is generally bound by State Highway 1 to the north, the Kaikorai Stream and Estuary to the north and west, the Green Island Wastewater Treatment Plant (GIWWTP) to the southwest, rural land to the south, and the Clariton Ave residential area and Brighton Road industrial area to the east. The landfill is located within a designation for landfill and waste management purposes.

The margins of the Kaikorai Stream and Estuary bordering the landfill to the north and west are identified as a Regionally Significant Wetland in the Regional Plan: Water; and an Area of Significant Biodiversity Value, and a Wāhi Tupuna of cultural significance to mana whenua in the DCC Second Generation District Plan (2GP). Low lying areas around the stream and estuary are also identified as being within a Hazard 2 Flood overlay at risk of flooding in the 2GP.

2.2.1 History

The site history and distribution of waste is described in Appendix A. Landfilling commenced at the south-east corner of the landfill site and has continued north and west over the decades. The RRPP area is estimated to have a waste thickness of between 3 and 10m.

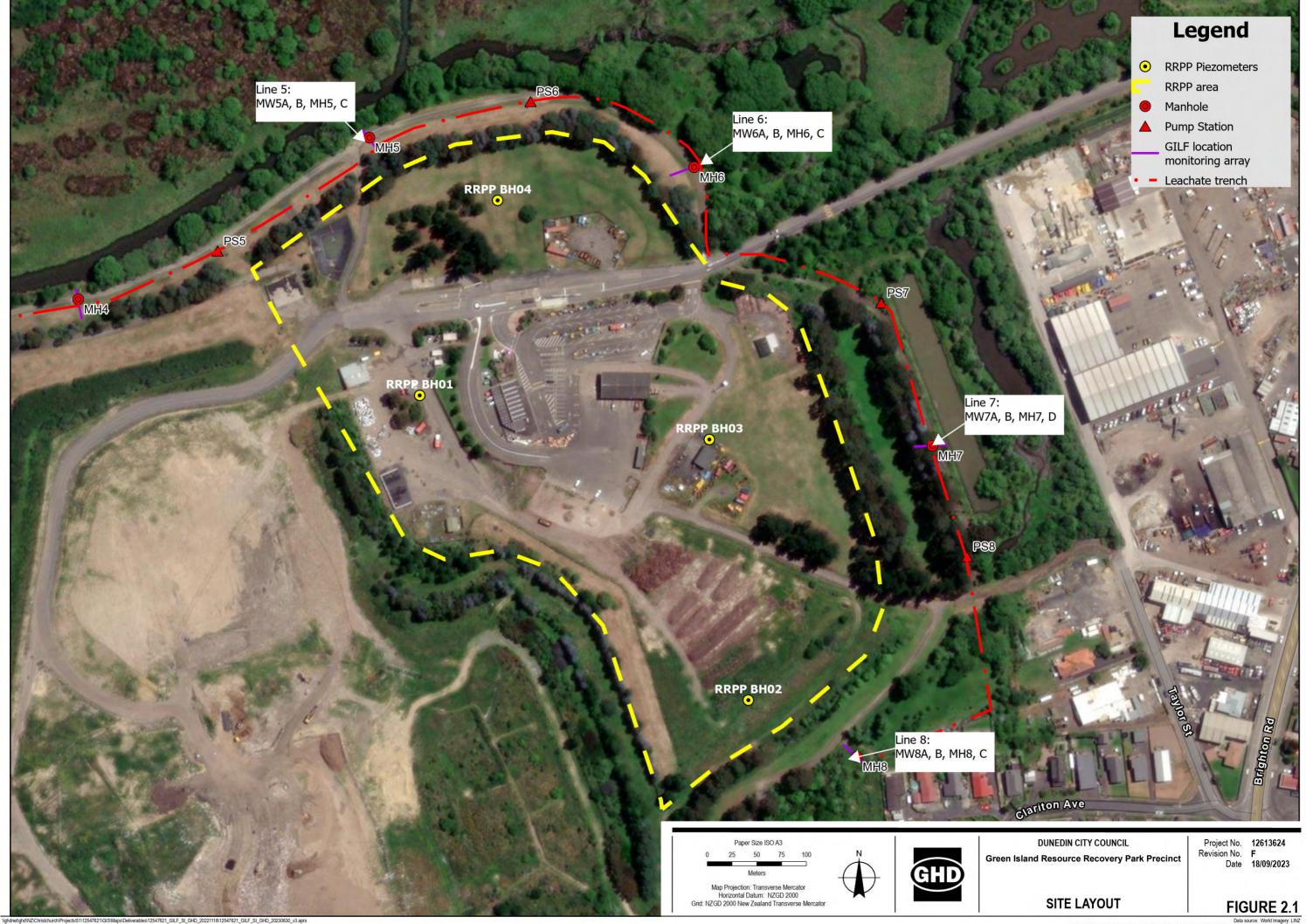
The pre-existing landform for the Green Island landfill was tidal estuary associated with the upper reaches of the Kaikorai Estuary. Abbotts Creek flows into Kaikorai Stream to the north of the site with the Kaikorai Stream flowing to the east then south into the Kaikorai Estuary. Waste was originally end dumped directly onto the estuarine muds and up against the southeastern estuary edge where the pre-existing landform rises gently to the southeast. A soil bund has been constructed around the edges of the landfill to constrain waste placement. The soil bund separates the main landfill area from the RRPP.

A leachate interception trench was installed along most of the perimeter of the site in the 1990s. In the landfill area, the leachate interception trench is located on the outside of the soil bund, between the landfill and Kaikorai Stream and estuary. The leachate interception trench surrounds the RRPP area and is located between the RRPP, the sedimentation ponds and Kaikorai Stream (see Figure 2.1).

2.2.2 Leachate management

The leachate interception trench was commissioned in 1994 and installed in 1995. Details on the construction and operation of the trench are provided in Appendix A. The trench creates a hydraulic barrier for groundwater and leachate migration offsite. A HDPE liner was placed against the outside face of the trench during construction and aids in reducing the volume of water entering the trench from the Kaikorai Stream but does not completely prevent inflows. The continuous dewatering of the trench is required to maintain this barrier, with the pump stations set to maintain water levels at low levels to create the hydraulic gradient which directs flow to the trench.

In the RRPP area the leachate collection trench was installed in historic waste along the eastern boundary (between MH8 and MH6), shown in Figure 2.1. There is a 90 m gap in the trench between MH8 and PS9 on the southeastern side of the RRPP area. However, this gap aligns with a short ridge of land that extended into the estuary based on historical maps and photos. Based on the geological map (Figure 2.2) this ridge is inferred to be mudstone and the trench is inferred to have been butted into either side of this ridge.



Pump flow rates and pump hours are recorded within each pump station continuously. Median flow rates for each pump station are presented in Table 2.1. The leachate collection system receives stormwater and diverted flows from water held in the northern leachate pond, with higher flows (up to 7 L/s) following rainfall events. Lower flow rates in the 2020-2021 and 2021-2022 reporting years are likely to be related to below average rainfall during this period but may also partly reflect the installation of final capping to part of the landfill which will have reduced rainfall seepage into the waste. The four pump stations in the RRPP area (PS5, 6, 7 and 8) intercept ~0.2 L/s of leachate each.

Table 2.1 Pump station flows – Median flow rate (reporting year July to June). Pump stations PS6-PS8 (shaded) are located adjacent to RRPP

Pump station	2019 – 2020 (L/s)	2021 – 2022 (L/s)
PS1	0.76	0.13
PS2	0.04	0.03
PS3	0.04	0.08
PS4	0.15	0.10
PS5	0.31	0.24
PS6	0.25	0.21
PS7	0.20	0.19
PS8	0.16	0.10
PS9	<0.01	<0.01
Combined median flow rate	2.0	1.3

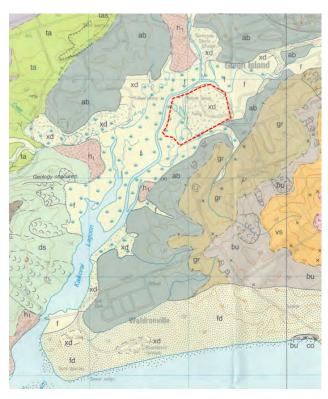


Figure 2.2 Geological map of Green Island area (McKellar, 1990). Approximate site boundary shown by red dashed line. ab: Abbotsford mudstone, f: alluvium, xd:fill

2.3 Hydrology

The Kaikorai Stream borders the RRPP area to the north. Summary flow statistics for the Kaikorai Stream are provided below in Table 2.2 (NIWA, 2023¹). Further details on the surface water environment are provided in the Waste Futures – *Green Island Landfill Closure* – *Surface Water Report (GHD, 2023C)*. From time to time, the estuary/ river mouth is blocked as a result of sand/debris accumulation from storm events. When this occurs, water backs up in the stream and results in higher stream levels adjacent to the site. Elevated water levels are maintained until the mouth is opened up again, either by natural or mechanical methods.

Table 2.2 Kaikorai Stream Flow Statistics (source NIWA)

Location	Mean flow	Mean Annual Low Flow
Upstream of Abbots Creek confluence	227 L/s	49 L/s
Downstream of Abbots Creek confluence	368 L/s	81 L/s

2.4 Geology

A description of the geological and hydrogeological setting is provided in Appendix A. In the RRPP area the historic landfill waste is sitting on sedimentary deposits associated with the Kaikorai Estuary. The geological profile is summarised in Table 2.3 and based on published reports and the results of the GHD (2023B) site investigation.

The site is situated on estuarine sedimentary deposits associated with the Kaikorai Estuary. These have been characterised as the Kaikorai Estuary Formation (KEF) (BDGC, 2002). KEF is subdivided into two members, Upper Kaikorai Estuary Member (UKEM, predominantly sand and silt) and Lower Kaikorai Estuary Member (LKEM, predominantly silt and clay). The UKEM appears to be absent from the RRPP area (GHD Geotech report 2023B). In several of the bore holes there was a coarse-grained layer (sand and/or gravel) at the contact of the Lower Kaikorai Estuary Member (LKEM) and mudstone.

Table 2.3 Summary geological profile – RRPP area

Geology	Description	Layer thickness
Fill	Variable landfill waste and soil	~7 m
UKEM	Silty fine to medium sand, sandy silt	Mostly absent from RRPP area
LKEM	Organic silt, silty clay	2 – 3 m
Coarse sediments	Sands/gravel	0.5 – 1 (only some bore holes)
Abbotsford Mudstone	Grey-brown mudstone, very weak	-

2.5 Hydrogeology

2.5.1 Overview

Prior to the landfill, groundwater within the estuarine deposits (KEF) is likely to have been hydraulically connected to the Kaikorai Stream and other surface water features. As discussed in Section 2.2.2, pumping from the perimeter leachate trench creates a hydraulic barrier between surface water and the shallow aquifer underlying the landfill. The underlying Abbotsford Formation mudstone is inferred to be an aquitard due to the very low permeability of the mudstone and effectively an impermeable barrier for downward seepage. Adams Geotechnical (2019) reported permeability for the Abbotsford Formation mudstone associated with the capping material borrow pit located to the east of the landfill, with permeabilities as low as 1 x 10⁻¹⁰ m/s. Hydraulic testing of monitoring wells (Appendix A) indicated a range of hydraulic conductivity values, this reflects heterogeneity created by the depositional environment. Summary results are included in Table 2.4.

¹NIWA River Maps online view: https://shiny.niwa.co.nz/nzrivermaps/ (accessed 09/02/2023)

The Abbotsford Formation mudstone also forms the spur of ground in the gap between MH8 and PS9 (see Figure 2.1). As the trench butts into either side of this spur the low permeability nature of the mudstone is an effective barrier to leachate seepage around the ends of the leachate trench.

Table 2.4 Hydraulic conductivity summary

Geology	Monitoring wells*	Hydraulic Conductivity Range (m/s)	Adopted Hydraulic Conductivity (m/s)
Upper Kaikorai Estuary Member (UKEM)	MW1C, MW5C, MW6C	8.4 x 10 ⁻⁷ to 2.8 x 10 ⁻⁶	1 x 10 ⁻⁶
Lower Kaikorai Estuary Member (LKEM)	BH100, BH101, BH103, BH108, MW2D, MW8C	6.2 x 10 ⁻¹⁰ to 3.3 x 10 ⁻⁶	1 x 10 ⁻⁷
Abbotsford Mudstone (AM)	BH104	< 1 x 10 ⁻⁹	1 x 10 ⁻⁹

2.5.2 Groundwater monitoring

Monitoring of water levels and water quality is undertaken on a routine basis in accordance with the conditions of the current consents. The monitoring programme includes sampling of surface water, groundwater and leachate. A review of the groundwater data and surface water data relevant to the groundwater assessment is included in Appendix A.

The monitoring well network around the RRPP is shown in Figure 2.1. The monitoring well network comprises lines of monitoring well transects along the leachate collection trench. Monitoring well lines 5, 6, 7 and 8 are located around the perimeter of the RRPP area. Each well line is located at mid-distance between two pump stations and each line comprises three shallow wells, MWA through to MWC², with the exception of Line 7, where MWC is absent (noting that the majority of these are founded in the UKEM geological unit). Deep wells are located at well lines 2, 4, and 7. These deep wells are screened in the LKEM.

Additional monitoring wells were installed in 2021 within the RRPP, the locations of these wells are shown in Figure 2.1 (RRPPBH 01-04). Unfortunately, one of the wells, RRPPBH02 located in the current composting area, has recently been destroyed or buried. Bore logs are included in Appendix B.

2.5.2.1 Groundwater levels

The leachate collection trench was designed to create a hydraulic barrier by reducing groundwater levels in the trench and drawing groundwater from both sides of the trench (i.e. landfill and stream side, although the stream side inflows were designed to be retarded by the installation of a HDPE liner on the stream side of the trench). Groundwater levels are checked on a monthly basis to monitor the hydraulic gradient. The water level monitoring consistently shows that the lowest groundwater levels (in monitoring wells) occur adjacent to the trench (MWC wells) and the leachate trench is performing as designed. These results are shown in the Green Island annual monitoring report (GHD, 2023D).

Groundwater level data recorded in the RRPP monitoring wells is summarised below in Table 2.5ll of the wells are screened over landfill waste. Bore logs for these wells are included in Appendix B. Groundwater levels recorded in RRPPBH03 (Figure 2.3) indicate that groundwater levels generally respond to rainfall events.

Table 2.5 Water levels recorded in RRPP monitoring wells

	Water / leachate level m below ground level			
Date	RRPPBH01	RRPPBH02	RRPPBH03	RRPPBH04
20/02/2023	2.32	2.52	2.78	3.41
13/04/2023	2.16	2.20	2.55	3.27

² "MWA relates to the monitoring bore code. For example, MWA1A denotes well line (1) and position of the well (A being closest to the landfill). See Appendix A for more details.

	Water / leachate level m below ground level			
Date	RRPPBH01	RRPPBH02	RRPPBH03	RRPPBH04
26/06/2023	1.87	-	2.22	2.92
20/7/2023	2.02	-	2.31	3.05

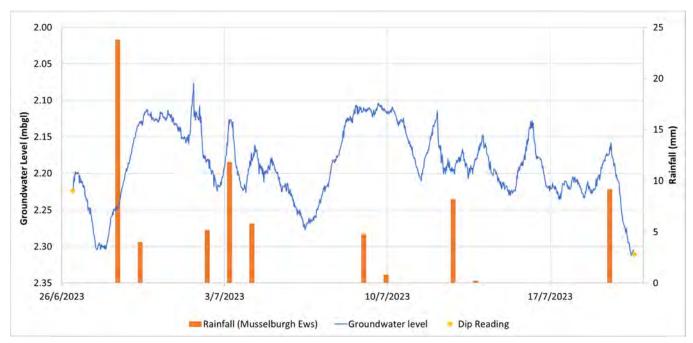


Figure 2.3 Groundwater level data at RRPP BH03 from June 26, 2023 to July 20 2023

2.5.2.2 Groundwater quality

Groundwater quality is monitored on a quarterly basis in accordance with the consent conditions, and results are reported in annual monitoring reports, the most recent being the 2022-2023 report (GHD, 2023D). The annual report provides a full summary of water quality trends and comparison against relevant standards/guidelines. A brief overview of the data is provided below with selected monitoring plots included in Appendix A.

In general, the monitoring wells inside the leachate trench (A and B series) are likely to be impacted by leachate. However, some of the wells on the outside of the trench also show the influence of landfill waste. A review of the site history and waste distribution was undertaken as part of the landfill technical assessment (GHD, 2023A) which identified areas where waste has been placed outside of the leachate trench. The likely extent of waste on the outside of the trench is shown on Figure 2.11 of Appendix A. In the RRPP area, all wells, with the possible exception of 5C, are expected to be impacted by historical waste. Monitoring well 7D is only 7 m deep and is likely to be screened within waste materials.

The impact of waste outside of the trench is managed through operation of the leachate collection system, which pulls groundwater (and any leachate) from both sides of the trench.

Groundwater quality trends, major ion composition and isotopic analysis are discussed further in Appendix A. The major ion composition shows that the samples from monitoring wells 6C, 7D and 8C have an intermediate composition between groundwater and landfill leachate (Figure 2.4).

No water quality data is available for the monitoring wells installed in the centre of the RRPP, however given that they are screened in waste materials it is expected that the water quality will reflect leachate chemistry.

2.6 Other groundwater users

Bore records and water take consents listed on ORC webmaps³ were reviewed for a 2 km radius surrounding the site. Two water takes were identified, one upgradient of the site at Blackhead quarry and the other for Maxwells

³ https://maps.orc.govt.nz/OtagoMaps/

Landfill (previously operated by Waste Management) (Figure 2.5). Two bore consents were identified, the status for these consents are described as "proposed". Forty-eight bore records were identified in the area of interest, as summarised in Table 2.6. The site is not located within a mapped aquifer zone⁴.

Table 2.6 ORC Bore records within 2 km radius of the site

Well type	Count	Comments
Monitoring / Investigation	13	
Dewatering	1	Green Island WWTP
Domestic / stockwater	1	Proposed – RM14.355.01
Other	33	Green Island Landfill investigation bores /monitoring wells

⁴ https://www.orc.govt.nz/plans-policies-reports/regional-plans-and-policies/water

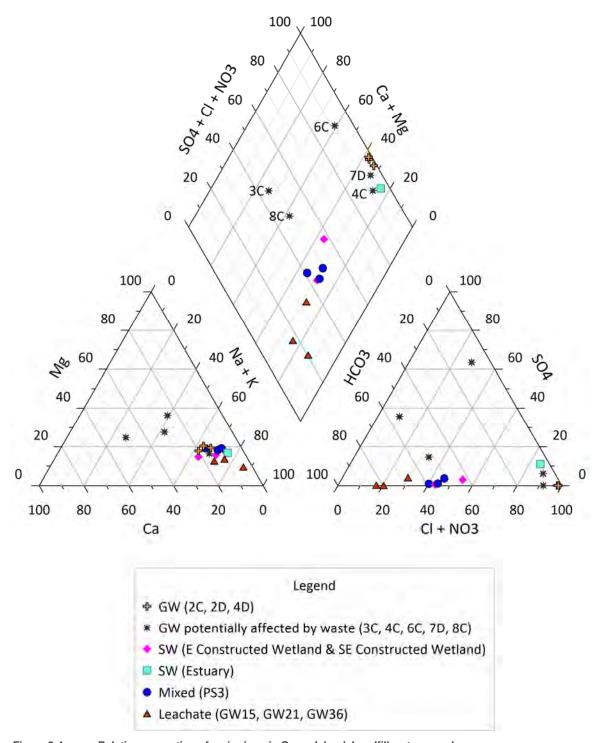
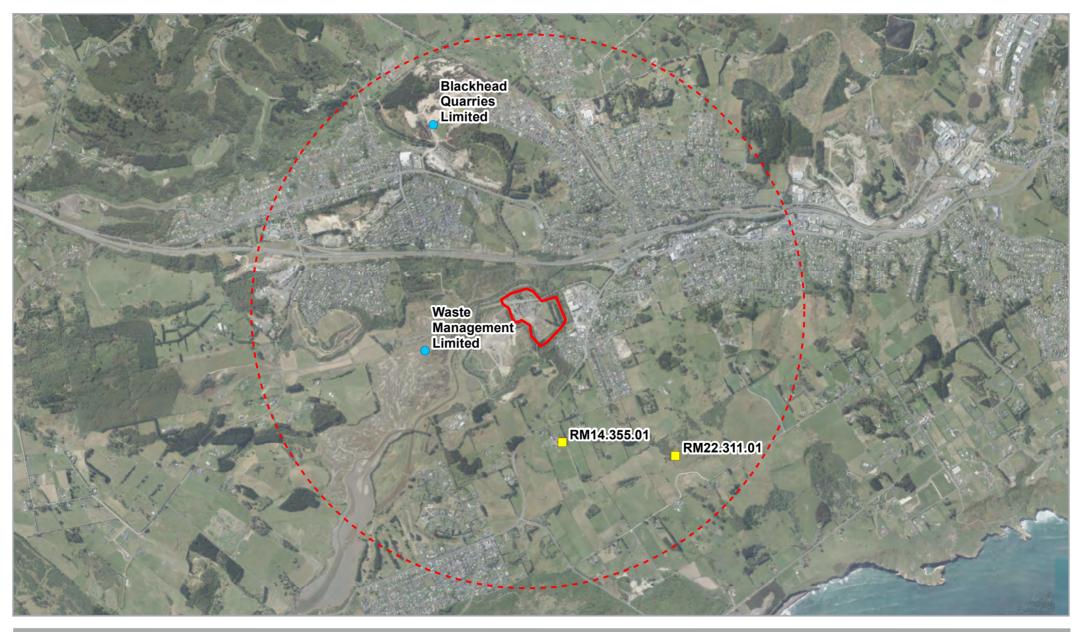


Figure 2.4 Relative proportion of major ions in Green Island Landfill water samples





Map Projection: Transverse Mercator Horizontal Datum: NZGD 2000 Grid: NZGD 2000 New Zealand Transverse Mercato



Legend

Water takes





Dunedin City Council Waste Futures - Green Island Resource Recovery Park Precinct

WATER TAKES AND BORE CONSENTS WITHIN 2 KM RADIUS

Project No. 12613624 Revision No. A

Date 03/07/2023

FIGURE 2.6

3. Technical assessment

3.1 Introduction

The following section documents the conceptual understanding of the groundwater system, interaction of groundwater with the landfill leachate and the leachate collection system, and interaction with surface water. This conceptual understanding, site data and investigation results, have been used to assess the effect of the proposed activities.

A technical assessment for the wider site, including the effectiveness of the leachate interception trench is detailed in Appendix A and includes:

- An assessment of rainfall infiltration through the current and proposed cap (HELP modelling);
- An assessment of leachate head within the landfill (2D SEEP/W groundwater modelling);
- An assessment of the effectiveness of the leachate trench in intercepting leachate (SEEP/W); and
- Groundwater surface water interaction.

The findings of this assessment can be applied to the RRPP area with rainfall infiltration into the waste in areas not covered by hardstand, generation of leachate, and interception of leachate in the perimeter leachate trench.

A key aspect of the RRPP development is an increase in the area covered by hardstand and/or buildings. Thereby reducing the area available for rainfall infiltration into the landfill material underlying the RRPP. In addition, the existing green waste/organics area (~1 ha), where leachate and stormwater can currently infiltrate the cap will be replaced by a hardstand organics facility with leachate and stormwater management controls.

An additional modelling assessment was undertaken to model the potential dewatering required during the construction of buildings as part of the RRPP development (Section 3.3). This assessment has used the design information presented in GHD 2024B) and assumed that an excavation up to 2.5 m deep will be required for the foundations (gravel raft) of the following RRPP buildings and facilities (see Figure 1.2):

- MRF building and apron;
- BWTS building;

It is understood that excavation for the MRF building and apron (buildings 1 and 2, Figure 1.2) will be undertaken at the same time and will require the largest excavation of the proposed buildings. Therefore, for the purposes of this assessment, the technical assessment (dewatering modelling) has been undertaken on this excavation to represent the largest potential dewatering effects.

3.2 Conceptual model

The conceptual understanding of the groundwater system for the RRPP area is summarised below (Figure 3.2). The model is based on published information, routine monitoring, and previous site investigations. In summary:

- Infiltration of rainfall into the old landfill material beneath the RRPP site (in areas not covered by hardstand/concrete pad or buildings) results in generation of leachate as water comes into contact with the underlying waste;
- Migration of leachate occurs downwards and outwards towards the edge of the landfill driven by pumping and maintenance of a low ground water/leachate level barrier in the leachate collection trench that surrounds the site:
- The pumped groundwater/leachate is directed to the Green Island Wastewater Treatment Plant (GIWTP).
- Water chemistry for the abstracted groundwater/leachate indicates mixing of water types. Surrounding groundwater levels indicate the abstracted water is a mixture of groundwater/leachate from the waste materials underlying the RRPP and a limited amount of surface water drawn in from the nearby Kaikorai Stream;

- Stream depletion effects are limited by the presence of a HDPE liner on stream side of trench;
- Upward gradient in LKEM beneath the landfill waste restricting migration of leachate into deeper layers under the trench collection system; and
- Underlying mudstone forms an aquitard limiting deeper flow paths.

RRPP Design and Operation

As noted above, the RRPP development will result in increased hardstand areas, and changes to the organics/composting practices. In relation to the groundwater environment, the following is expected and shown on Figure 3.1.

- The RRPP development is expected to result in reduced leachate generation volumes from rainfall seepage into the underlying waste due to the increase in the area of hardstand and/or buildings and associated increase in stormwater runoff.
- The current green waste/organics processing area (~ 1 ha), which discharges stormwater and leachate to ground will cease and be replaced by the ORB and organics process bunkers/maturation area. Leachate at these facilities will be managed as follows (see GHD 2024A and GHD 2024B for further details on leachate and stormwater management):
 - The ORB will generate limited amounts of leachate as materials are contained within an enclosed building. The building has a concrete floor and all wash down water/leachate will be collected and stored in a tank adjacent to the building. Collected leachate will be re-used to add moisture to processed materials/wash down. If disposal is required, leachate will be directed to the existing leachate collection system as described below;
 - The organic processing facility bunkers will have a concrete floor and will include leachate collection facilities. Leachate will be managed as described below; and
 - The maturation pad will have a hardfill base and from a stormwater perspective be effectively impermeable. Very little seepage of leachate will occur through the compacted hardfill base. The pad will be graded to allow leachate runoff from the maturation piles to be collected and managed as described below.
- For these new facilities, leachate will be collected and directed to the existing leachate collection system pump stations for direct piping to the GIWTP.
- For the balance of the RRPP site the increase in roof areas and hard standing will result in a reduction stormwater seepage into the underlying waste and the generation of leachate.
- Therefore, compared to the existing operation, the proposed collection and management of leachate and stormwater will result in an overall reduction in seepage to ground and reduction in the generation of leachate within the underlying waste materials.
- In addition, no changes are proposed to the operation of the leachate collection trench as part of the RRPP development that would impact its functionality, and hence create a potential adverse effect on surface water quality.

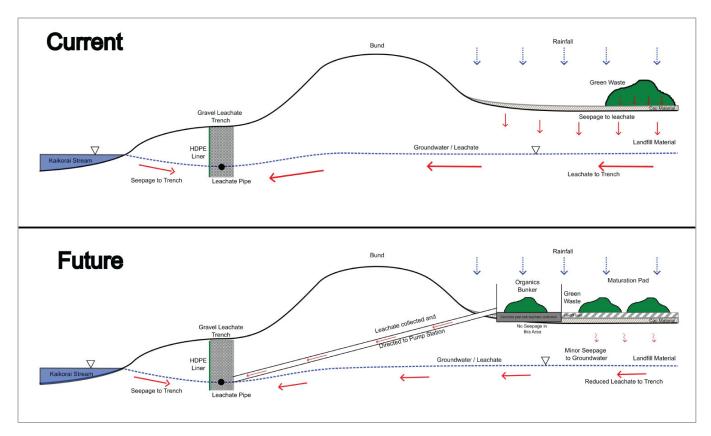


Figure 3.1 Conceptual model – RRPP area

RRPP Construction and Dewatering

The conceptual model shows the proposed excavation for the MRF buildings and effect of dewatering on the groundwater/leachate level. This conceptualisation is considered to be conservative as it exaggerates the dewatering required and shows the RRPP to be at a similar elevation to the leachate trench (when much of the area is higher).

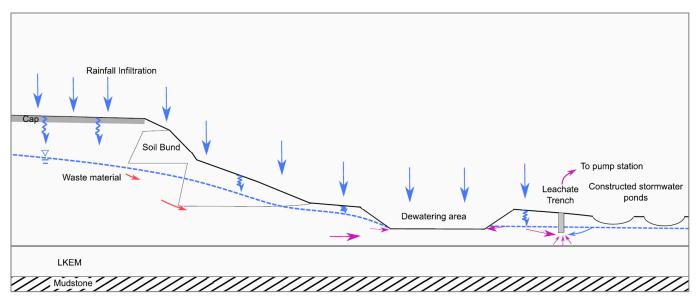


Figure 3.2 Conceptual model - construction dewatering

3.3 Construction Dewatering assessment

The proposed design of the RRPP facilities involves the installation of gravel raft building foundations to a (maximum) depth of 2.5 m below ground level. It is likely that dewatering will be required to manage leachate inflows into the excavations for the MRF. Modelling of the proposed dewatering was undertaken using 2D Geostudio 2021 SEEP/W finite element modelling software. Modelling was undertaken to estimate the seepage into the proposed excavation and simulate drawdown of leachate levels within the RRPP area. This assessment adopted the material properties and boundary conditions used in the modelling assessment undertaken for the landfill closure assessment (Appendix A), except as outlined below.

The model focused on the excavation that will be undertaken for the MRF building and apron (Buildings 1 and 2 on Figure 1.2) as this represents the largest potential for dewatering effects. No sheet piling will be used for the excavation, the excavation will be stepped back at an approximate 3:1 ratio. The model reflects this excavation method and geometry. Key model inputs for the dewatering assessment are shown in Table 3.1. The location of the model cross section line is shown in Figure 3.3. The dewatering model was initially set up as steady state to simulate the measured groundwater level (~2.0-2.5 m bgl) and then run as a transient model to simulate the proposed dewatering over a 30 day period. The model cross section is shown in Figure 3.4.

Table 3.1 Dewatering model set up

Parameter	Value
Model time period	30 days
Model time step	0.25 days
Length of excavation base	195 m
Length of excavation at surface	210 m
Depth of excavation	2.5 m



Figure 3.3 Site layout, model cross section shown in red

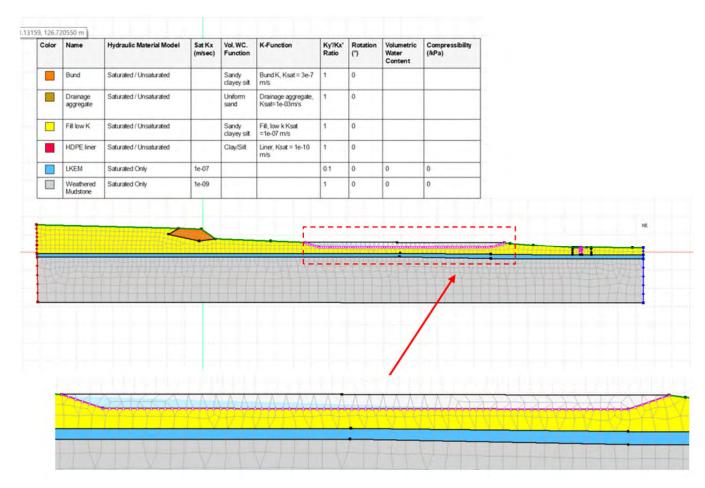


Figure 3.4 SEEP/W model set up, excavation trench shown in detail

The excavation is expected to be into historic waste materials. It is likely that the waste material is highly heterogeneous, with some areas having a higher hydraulic conductivity than other areas depending on the type of waste materials and the degree of compaction undertaken. Hydraulic conductivities of municipal solid waste (MSW) reported in the literature vary between 1 x 10⁻³ and 1 x 10⁻⁹ m/s, although most values are in the range of 10⁻⁵ to 10⁻⁶ m/s (LANDSS⁵). It is understood that soil makes up a significant proportion of the waste material at Green Island landfill, which is likely to reduce the bulk hydraulic conductivity of the landfill material. The best calibration for the closure modelling assessment (Appendix A) was achieved using a waste hydraulic conductivity of 6 x 10⁻⁷ m/s. It is not known if soil was widely deposited in the waste underlying the RRPP area. However, given the heterogeneity of the waste material, three model scenarios were run to represent the expected range in bulk hydraulic conductivity (Table 3.2).

Table 3.2 Model scenarios

Scenario	Hydraulic conductivity of waste
Low	1 x 10 ⁻⁷ m/s
Mid	1 x 10 ⁻⁶ m/s
High	1 x 10 ⁻⁵ m/s

⁵ Landfill (Aftercare) Decision Support System, University of Southampton https://landss.soton.ac.uk/

3.3.1 Results

The results of the modelling are presented in Table 3.3 below. The 2D modelling approach calculates flow for a 1 metre slice of aquifer. The results generated from dewatering model scenarios are scaled up by the width of the excavation (50 m) to provide the estimated inflow rates presented in Table 3.3. The applicant proposes that leachate collected during dewatering of the excavation could be piped to one of the Pump Stations associated with the existing leachate collection system or recirculated and discharged to the landfill. Given the variability in the waste material, there are likely to be pockets of slightly higher inflows compared to other areas of the excavation. Nevertheless, the estimated flow rates are well within the operating range of the leachate collection system.

Table 3.3 Modelled dewatering flow rates

Scenario	Peak flow rate (L/s)	Average flow rate (L/s)	30 day volume (m³)
Low (1 x 10 ⁻⁷ m/s)	0.006	0.003	7.5
Mid (1 x 10 ⁻⁶ m/s)	0.03	0.02	45
High (1 x 10 ⁻⁵ m/s)	0.2	0.1	330

Assessment of effects on the environment 4_

4.1 **Operational Effects**

An assessment of effects for the wider Green Island landfill, including capping and operation of the leachate system, is included in Appendix A and provides context for the site including the RRPP area. Effects addressed in the Green Island landfill assessment associated with the placement of waste at the site include:

- Effects to groundwater levels and flow;
- Effects to surface water flows;
- Effect of climate change and sea level rise; and
- Effects to water quality.

The landfill assessment considered the effectiveness of the leachate interception trench in preventing leachate migration off site into groundwater and surface water. Water quality and water level measurements supported the conclusions of the modelling assessment, that the leachate collection trench is effective at intercepting landfill leachate.

These conclusions apply to the RRPP area in relation to the existing waste that underlies the site. However, as described in Section 3.2, the RRPP development is expected to reduce the overall impacts on groundwater as:

- The increase in hardstanding and building areas along with additional stormwater control measures will result in more runoff and less infiltration to the underlying groundwater/leachate. The Surface Water Report for the RRPP application (GHD 2024A) indicates an increase in surface runoff for the site during a 50-year 30 mins rainfall event of approximately 125 litres/second. This will result in a corresponding decrease in infiltration to groundwater, leachate generation, and the requirement for pumping and treating leachate via the leachate collection system and GIWWTP. This is a positive impact.
- The current green waste/organics processing area (~ 1 ha) has no hard standing or formalised stormwater control measures. Any rainfall or seepage from the current operation will mostly seep into the underlying groundwater/leachate system. This will be replaced by the ORB and organics processing bunkers/maturation area. For these new facilities, leachate will be collected and directed to the existing leachate collection system for treatment and stormwater runoff will be managed through the existing stormwater facilities (as described in the above bullet point). The proposed collection and management of leachate will result in an overall reduction in contaminated seepage to ground - albeit that the existing green waste operation is primarily garden waste and similar. This is also considered a positive impact.
- No changes are proposed to the existing design of the leachate collection system. As described in GHD 2024A where leachate is generated on site (including the ORB and organics processing facility) it will be directed to three of the existing leachate collection system pumps (PS5, 6 and 7) for subsequent pumping to the GIWWTP. GHD 2024A, describes how this transfer will be managed to ensure any leachate from the RRPP is pumped directly from the pump stations and will not backflow into the adjacent sections of leachate collection trench during periods of high rainfall and leachate generation. Management of the system in this way ensures the RRPP facilities have no impact on the operation of the leachate collection trench.

4.2 **Construction Effects**

An assessment of effects from the proposed dewatering activities during construction is provided in the following

Surface water quality is primarily addressed in GHD 2024A and the Waste Futures - Green Island Landfill Closure - Surface Water Report (GHD, 2023C).

4.2.1 Dewatering effects

Monitoring of leachate levels in the waste materials underlying the RRPP area indicates that dewatering may be required during the excavation and installation of a gravel raft foundation proposed for the RRPP buildings. Modelling was undertaken to estimate the required dewatering rate for the largest of the proposed excavations. The modelling assessment indicated inflow rates up to 0.2 L/s for the MRF building and apron excavation. Due to the presence of water materials, it is assumed that all inflows will be landfill leachate. The applicant proposes to collect all leachate inflows and direct these inflows to the leachate collection system or recirculate and discharge to the landfill for disposal. The estimated flow rates are well within the usual operating range of the leachate system.

The leachate trench operates as a hydraulic barrier by drawing groundwater from both the landfill side and outside of trench. The dewatering activities are unlikely to result in any additional groundwater drawdown outside of the landfill and RRPP footprint.

4.3 Effects on other groundwater users

The surrounding area is not used for groundwater supply due to the low permeability geology, with the only two consented groundwater takes (within a 2 km radius of the site) located upgradient of the site (quarry and landfill). As noted above, the effect of the proposed dewatering activities of groundwater are expected to be limited to the area within the leachate perimeter trench with no effect on any other groundwater users.

Cumulative Effects 4.4

As described above in Section 4.1 the overall impact of the RRPP development on groundwater is expected to be a reduction in leachate generation at the site and collection/treatment requirements. This is considered a positive effect on the wider management of leachate at the site.

Recommendations 5.

Monitoring 5.1

Recommendations for monitoring are provided in the Green Island Landfill Groundwater Technical Assessment (Appendix A). No additional monitoring is recommended relating to groundwater for the RRPP area. Surface water monitoring is recommended in the Waste Futures - Green Island Landfill Closure - Surface Water Report (GHD 2023C).

Conclusions 6.

This technical assessment provides an overview of the potential effects on groundwater associated with activities for the construction and operation of the RRPP at Green Island landfill. An assessment of effects for the wider landfill, including capping and operation of the leachate system is included in Appendix A. and is considered applicable to the RRPP area. In summary:

- The leachate collection trench is designed to manage leachate seepage associated with waste material placed at the site. It has been assessed previously as being effective in intercepting landfill leachate and drawing groundwater from the limited areas of buried waste outside the trench;
- In regard to the RRPP construction and operation the overall impact on the underlying groundwater/leachate system is considered positive or neutral as:
 - The increase in areas of hardstanding and buildings along with improved stormwater management systems will result in an increase in stormwater runoff from the site and a reduction in seepage to the underlying groundwater system and generation of leachate.
 - The current green waste/organics area with associated uncontrolled seepage of any associated leachate to groundwater will be replaced by the ORB and organics processing facility where all leachate will be collected and discharged directly to the existing leachate collection system.
 - No changes will be made to the existing leachate collection systems and all new discharges to the system will be managed to ensure leachate is pumped directly from one of the existing pump stations to the GIWWTP for treatment.
 - Limited dewatering may be required for the construction of gravel raft foundations for the MRF.
 - Leachate abstracted during the dewatering of excavations will be piped to the site leachate collection system or discharged to the landfill with no impact on surface water and groundwater environment outside of the site.

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Limitations 8.

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GHD has relied on information from a number of sources, including but not limited to the following:

Environmental Impact Assessment (Beca. 1992)

Green Island Landfill Leachate Collection Trench Geological Report (Barry J Douglas Geological Consultants, 2002 – herein referred to as BDGC, 2002)

Green Island Landfill Gas Management Project. Stage 1: Investigation Works. URS, 2007.

Bund stability report (T&T, 2020)

Leachate collection summary (DCC) and associated plans

Claymine Assessment (DCC) and associated plans

GILF & RRRP master plan (Stantec 2020)

Green Island Annual monitoring reports prepared by GHD, 2022

Appendices

Appendix A

Technical groundwater assessment for landfill closure consent





Waste Futures – Green Island Landfill Closure

Groundwater Technical Assessment

Dunedin City Council
09 March 2023

→ The Power of Commitment



Project n	ame	GILF Closure Consents						
Documer	nt title	Waste Futures – Green Island Landfill Closure Groundwater Technical Assessment						
Project number		12547621						
File name)	GHD Technical A	ssessment - Gro	undwater_230309_	Rev02.docx			
Status	Revision	Author Reviewer		Approved for issue				
Code			Name	Signature	Name	Signature	Date	
S3	RevA	Dusk Mains	Stephen Douglass	Joy	Stephen Douglass	Just	12 Feb 2023	
S4	Rev01	Dusk Mains	Stephen Douglass	Short	Stephen Douglass	floor -	9 March 2023	

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1. Introduction

1.1 Background

As part of Dunedin's wider commitment to reducing carbon emissions and reducing waste going to landfill, the Dunedin City Council (Council) has embarked on the Waste Futures Programme to develop an improved comprehensive waste management and diverted material system for Ōtepoti Dunedin. The Waste Futures Programme includes the roll out of an enhanced kerbside recycling and waste collection service for the city from July 2024. The new service will include collection of food and green waste.

To support the implementation of the new kerbside collection service, the DCC are planning to make changes to the use of Green Island landfill site (Figure 1.1) in coming years.



Figure 1.1 Site location

The proposed changes include:

- planning for the closure of the Green Island landfill, which is coming to the end of its operational life
- developing an improved Resource Recovery Park (RRPP) to process recycling, and food and green waste
- providing new waste transfer facilities to service a new Class 1 landfill currently planned for a site south of Dunedin, at Smooth Hill.

The resource consents for the new Smooth Hill landfill are subject to appeal. Depending on the outcome of this appeal process, and the time needed to undertake baseline monitoring, preparation of management plans, landfill and supporting infrastructure design and construction, DCC anticipate that the new Class I landfill facility, won't be able to accept waste until 2027/2028 at the earliest.

In the interim, DCC therefore plans to continue to use Green Island landfill for waste disposal. Based on Dunedin's current waste disposal rates, it is likely that that the Green Island landfill can keep accepting waste for another six years (until about 2029). Between now and then, and as it continues to fill up, the landfill will be closed and capped in stages. When the landfill closes completely, there will be opportunities for environmental enhancements and public recreational use around the edge of the site. Examples could be planting restoration projects and new walking and biking tracks beside the Kaikorai Estuary. Long term use and public access to the landfill site post closure will be determined in consultation with Te Rūnanga o Ōtākou, the local community and key stakeholders.

As current Otago Regional Council resource consents needed to operate a landfill at Green Island expire in October 2023, the DCC are now applying to ORC for replacement resource consents to continue to use the landfill until it closes completely, and waste disposal can be transferred to a new landfill facility. The replacement consents relate to ground disturbance, flood defence and discharges to land, water, and air. The site is subject to an operative designation (D658) in the Proposed Second-Generation Dunedin City District Plan (2GP) for the purpose of Landfilling and Associated Refuse Processing Operations and Activities.

The development of the new RRPP and waste transfer facilities at Green Island does not form part of the replacement consent applications. Resource consents for the development and operation of the RRPP will be applied for following the completion of design work and technical assessments later in 2023.

This groundwater technical assessment has been prepared to support the resource consent application for the continued operation and closure of Green Island Landfill.

1.2 Purpose of this report

The purpose of this report is as follows:

- Provide a description of the groundwater environment, existing management of groundwater and leachate discharges, and environmental monitoring data.
- Provide a technical assessment of the potential effects on groundwater and connected surface water flows from the proposed extension of landfill operations and closure management.
- Provide an assessment of the effects of leachate leakage on groundwater quality.
- Provide recommendations for mitigation measures to minimise the effects on the environment and monitoring conditions to confirm the effectiveness of the recommended mitigation measures.

Note: the potential effects of site stormwater on downstream surface water flows and quality are discussed in the *Surface Water Technical Report* (GHD 2023) and referenced in this report.

This report should be read in conjunction with the following reports:

- Green Island Landfill Design Report (GHD, 2023)
- Surface Water Report (GHD, 2023), including Appendix B, Annual Monitoring Plan 2021-2022.
- Green Island Landfill Geotechnical Factual Report (GHD, 2023)
- Green Island Landfill Liquification and Stability Assessment (GHD, 2023)

These reports provide supporting information and context which the surface water assessment relies upon. Where appropriate, a summary of critical information is summarised in this report with crossed references to the relevant technical report.

In addition, the assessment undertaken in this report is based on a review of previous investigations, including those undertaken as part of the 1994 resource consent application. This information has been supplemented with additional site investigations undertaken by GHD to support the design and consenting process (documented in the *Geotechnical Factual Report* (GHD 2023). The information obtained from the desktop review and site investigations has been used to undertake an assessment of potential impacts to groundwater and connected surface water, associated with the proposed extension and ultimate closure of Green Island Landfill.

1.3 Current landfill operation and management

1.3.1 Current Consents

The operation of the Green Island Landfill, including associated waste processing operations and facilities, is currently subject to 14 existing resource consents granted by Otago Regional Council (ORC). The consents cover landfill operation activities relating to discharges to land, water, and air, taking and/or diverting water, and disturbance of a contaminated site. All consents expire on 1 October 2023.

The current consents limit the extent of landfilling through the combination of a maximum 38 ha landfill footprint, conditions limiting the deposit of waste to 270 m³/day and 100,000 m³/year¹, and the 2023 term of the consents. The consent conditions do not impose any specific limit on the overall finished height, shape, or contour of the landfill. However, the plans included in the 1994 resource consent applications show a finished landfill surface rising to a maximum height of 25 m above mean sea level (amsl).

The consent conditions also require the consents are exercised in accordance with a Landfill Work Programme (LWP) prepared by the consent holder, which is to be reviewed annually or at such lesser frequency as the consent authority may approve. Among other matters, the LWP is required to describe present projections and intentions for landfill operations, and the sequencing of works.

1.3.2 Landfill Development and Management Plan

A Landfill Development and Management Plan (LDMP) was developed following the issuing of the consents to serve the purpose of the LWP. The LDMP is to document site-specific procedures, including monitoring and contingency actions to be implemented to ensure the landfill achieves the conditions set out in the resource consents. The LDMP is structured into the sections set out below:

- 1. **Introduction –** the existing resource consents, designation, and status and review of the LDMP.
- 2. **Site Management –** management structure, responsibilities, requirements for staff training, and community liaison.
- 3. **Landfill Development** including design principles, landfill capacity, and the filling programme and sequence.
- 4. **Site Operations** including controls and procedures for access control, stormwater management, leachate management, LFG management, greenwaste mulching and composting, salvage and management of diverted materials, roading and traffic management, waste acceptance and placement, waste cover, and control of nuisances.
- 5. **Environmental Monitoring** including monitoring, recording, and reporting for surface water, groundwater, LFG, leachate, odour, and weather.
- 6. **Emergency Management** including procedures for management of fires, hazardous waste/materials, leachate and LFG escape, extreme weather/flooding, machinery failure, accidents, and earthquakes.
- 7. **Closure, Reinstatement, and Aftercare** including final capping, continued operation and maintenance of landfill infrastructure, and ongoing monitoring.

The LDMP was first provided to ORC in 1994 following the issuing of the consents and was subsequently updated in 2004, and 2007. The most recent LDMP, which reflects the current approach to landfill operation and management was provided to ORC in February 2023.

1.3.3 Landfill Operations Plan

The landfill is currently operated by Waste Management NZ Ltd. under contract to the Council. Waste Management NZ Ltd. are required to maintain a Landfill Operations Plan (LOP) which reflects the LDMP and more specifically addresses day-to-day management landfill operational matters.

The LDMP (February 2023) and LOP (October 2018) will be updated after the granting of any replacement resource consents to ensure that they align with the final approved consent documentation, and any resource consent conditions.

2. Environmental Setting

2.1 Introduction

A review of the environmental setting was undertaken to inform the technical assessment (Section 3). This review includes a factual summary of information relevant to the conceptual understanding of the groundwater system and baseline parameters used in the modelling assessment. Assessment and interpretation of the site data is included in the following sections.

2.2 Site Description

The Green Island Landfill site is located in the suburb of Green Island, approximately 8.8 km by road southwest of Central Dunedin. The landfill site comprises a total area of 75.6 Ha. The site is generally bound by State Highway 1 to the north, the Kaikorai Stream and Estuary to the west, the Green Island Wastewater Treatment Plant (GIWWTP) to the southwest, Brighton Road to the south, and the Clariton Ave residential area and Brighton Road industrial area to the east.



Figure 2.1 Site Layout

The margins of the Kaikorai Stream and Estuary bordering the landfill to the north and west are identified as a Regionally Significant Wetland in the Regional Plan: Water; and an Area of Significant Biodiversity Value, and a Wāhi Tupuna of cultural significance to mana whenua in the 2GP. Low lying areas around the stream and estuary are also identified as being within a Hazard 2 Flood overlay at risk of flooding in the 2GP.

2.2.1 History

The historical placement of waste and its distribution across the site is described in detail in Appendix D. The following provides a summary of the waste filling history that is relevant to the future engineering design and closure management presented herein.

Waste disposal first occurred at the Green Island site in 1954 with the disposal of industrial waste and the site has been used for waste disposal since that time. A number of other sites have been used over the decades across the Dunedin region including the "Maxwell" landfill on the opposite side of the estuary to Green Island landfill. The Maxwell landfill was formally closed to waste disposal in mid-2017 and the Green Island landfill continued as the sole municipal solid waste disposal facility in the Dunedin region after that time. The existing operational consents were granted in 1994.

Landfilling commenced at the south-east corner of the landfill site and has continued north and west over the decades. The eastern portion of the landfill has a relatively shallow depth of waste at around 3 to 6 m thickness and is currently used for facilities and waste transfer station operations. This area is proposed to be developed in the near future to establish the Resource Recovery Park Precinct (RRPP). No further waste disposal will occur in this area.

The main landfill area is located immediately to the west of the facilities area. Waste placement in this area has been confined over recent decades within a constructed soil bund that encircles the landfill on the eastern, northern and western sides adjacent to the estuary. However, prior to berm construction waste had been placed across the whole extent of the area (see Appendix D). In recent years significant waste disposal has progressed north to south. In the northern and eastern areas waste has been placed up to the 1994 design contours (see *Design Report (GHD 2023))* and final capping has been completed. The south western half of the landfill has up to approximately 6 m - 8 m depth of waste placed during the 1990's, and a further 10 to 15 m of waste can be placed in this area. This is the primary area where future waste placement will occur through to closure of the landfill.

The pre-existing landform for the Green Island landfill was tidal estuary associated with the upper reaches of the Kaikorai Estuary. Abbotts Creek flows into Kaikorai Stream to the north of the site with the Kaikorai Stream flowing to the east then south in the Kaikorai Estuary. Waste was originally end dumped directly onto the estuarine muds and up against the south eastern estuary edge where the pre-existing landform rises gently to the southeast. As discussed above, a soil bund was constructed around the north, west and south-western sides of the landfill to confine the waste from the adjacent Kaikorai Stream. The current landfill has an access track on the outside of the bund along with a leachate interception trench and perimeter groundwater monitoring wells.

The leachate interception trench was installed on the outside of the soil bund in the mid-1990's. This perimeter control is not present along the southern side of the landfill against the rising ground of the hillside (see Drawing 12547621-01-G102, Appendix A Design Report (GHD 2023). The main wastewater trunk sewer follows the existing southern extent of existing landfill, flowing to the GIWWTP located 200 m southwest of the landfill site. A surface water drain follows the alignment of the sewer creating a valley that intercepts runoff from the landfill and directs it to the leachate collection system.

2.2.2 Leachate management

The perimeter bund and leachate interception trench was installed on the outside of the soil bund in 1994 and commissioned in 1995. The leachate trench intersects contaminated groundwater (landfill leachate mixed with groundwater) seeping from the site. The leachate collection system comprises the gravel interception trench with the HDPE liner (on outer side) and the slotted PVC drainage pipe, together with a manhole and pump station configuration (Figure 2.2). Figure 2.3 shows details of the gravity leachate drain arrangement between pump stations and manholes. The manholes are located at the ends of each PVC collector pipe, allowing inspection and cleaning of the pipes to occur. Impacted groundwater is then conveyed by gravity to nine individual pump stations,

which then pump into a 125 mm dia. rising main, which conveys the leachate and groundwater to the main sewer and ultimately GIWWTP. The rising main has a discharge into the sewer at each end, one approximately 79 m south of PS1 and the other approximately 80 m south of PS8. Hence, the riser pumps in both clockwise and/or anti-clockwise direction before being discharged to the sewer line. PS9 discharges directly to the sewer main in the southern valley.

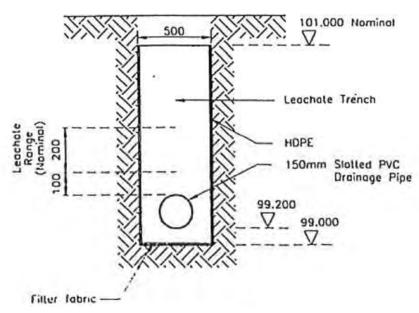


Figure 2.2 Leachate drain schematic (City Consultants, 1997)

The trench creates a hydraulic barrier for groundwater and leachate migration offsite. The HDPE liner aids in reducing the volume of water entering the trench from the Kaikorai Stream but does not completely prevent inflows. The continuous dewatering of the trench is required to maintain this barrier, with the pump stations set to maintain water levels at low levels to create the hydraulic gradient which directs flow to the trench. The leachate pumps are automated to start when the pumpwell reaches a "Pump ON" level, and then stop when the level reaches a "Pump OFF" level. The pumpstations also have alarms for the following conditions (high level alarm, pump running low level and power loss). Flow rate through the pump is measured continuously, as are the pump run hours. Consent conditions require regular monitoring of groundwater levels adjacent to the trench to confirm the hydraulic gradient (discussed in sections 2.3.3).

The trench is installed in the Upper Kaikorai Estuary Member (UKEM) (see section 2.2) comprising fine sands and silt. However, landfill refuse was recorded as overlying the UKEM in over half of the trench profiles during construction (Barry Douglas, 2022) with a maximum thickness of landfill recorded of 2.6 m. Borehole data from the recent drilling investigation (see section 2.3.2) confirmed similar profiles.

The leachate interception trench is absent along the southern edge of the landfill where waste is placed against the base of the slope that rises to the east. Management of leachate in this area is currently via a shallow surface drain which conveys the leachate (and any shallow groundwater seepage) to PS1. It is also noted that there is a 90 m gap in the trench between MH8 and PS9. This gap aligns with a short ridge of land that extended into the estuary based on historical maps and photos. Based on the geological map (section 2.2) this ridge is inferred to be mudstone.

A culvert located on the eastern side of the landfill between the South Eastern Constructed wetland and the Eastern Constructed Wetland has recently been identified as a pathway for leachate seepage, which has been confirmed from water quality monitoring and a culvert inspection (discussed in the *Surface Water Report*, 2023). These results and proposed mitigation are discussed further in the *Design Report* (GHD 2023).

Additional leachate drains have been installed over intermediate cover soils in the southern portion of the landfill and in the northern sector of waste placed in 2019-2022. These drains direct leachate to the perimeter leachate collection trench. These drains installed within the landfill are described in the *Design Report (GHD 2023)* and shown on Drawings 12547621-01-C204 (*Design Report – Appendix A*).

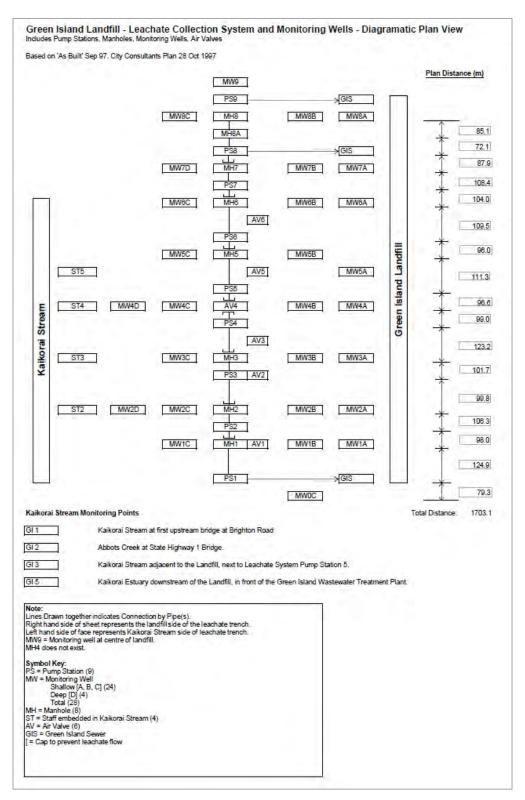
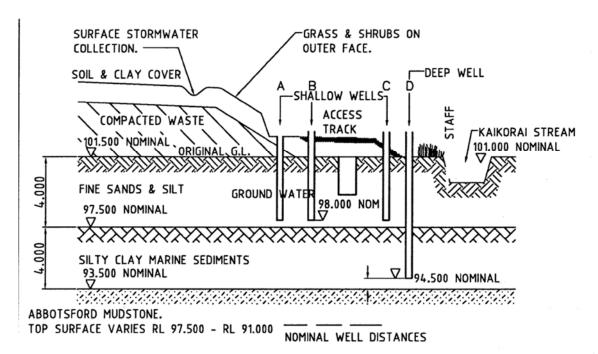


Figure 2.3 Well and pump station arrangement, leachate collection trench



TYPICAL CROSS SECTION LEACHATE TRENCH AND MONITORING WELLS

Figure 2.4 Typical cross section of leachate collection trench (MWH, 2004

2.2.3 Leachate volumes

Pump flow rates and pump hours are recorded within each pump station continuously. The volume of leachate pumped from each pump station to the rising main (weekly total) was reviewed as part of this assessment. Based on this record, a box and whisker plot of the combined flow rate (to all pump stations) averaged on a weekly basis is presented in Figure 2.5. Lower flow rates in the 2020-2021 and 2021-2022 reporting years are likely to be related to below average rainfall during this period (Table 2.1) but may also partly reflect the installation of final capping to part of the landfill which will reduce rainfall seepage into the waste. It is understood that PS1 has in the past received large quantities of stormwater flows from the landfill via open drains, with a recent improvement to the stormwater collection from the landfill (September 2021) resulting in some of these flows now going directly to PS3. Water held in the Northern Leachate Pond (see Drawing 12547621-01-C402) is also diverted to the leachate system (at PS5). Table 2.2 lists the median flow rates from the individual pump stations for an average rainfall year (2019-2020) and the past year (2021-2022, below average rainfall).

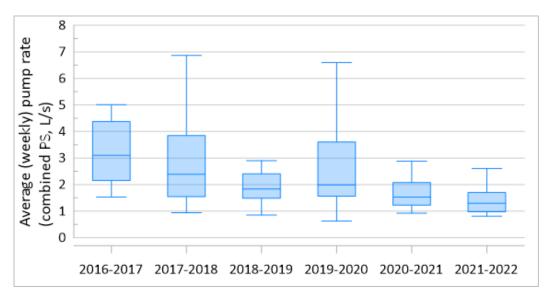


Figure 2.5 Combined flows from pump stations to rising main, reporting year July to June.

Table 2.1 Rainfall recorded at Musselburgh EWS

Year (July- June)	Rainfall total (mm) ¹
2016 – 2017	819.8
2017 – 2018	783.6
2018 – 2019	658.6
2019 – 2020	734.4
2020 -2021	556.6
2021- 2022	476

¹Average rainfall ~750 mm

Table 2.2 Pump station flows - Median flow rate (reporting year July to June)

Pump station	2019 – 2020 (L/s)	2021 – 2022 (L/s)
PS1	0.76	0.13
PS2	0.04	0.03
PS3	0.04	0.08
PS4	0.15	0.10
PS5	0.31	0.24
PS6	0.25	0.21
PS7	0.20	0.19
PS8	0.16	0.10
PS9	<0.01	<0.01
Combined median flow rate	2.0	1.3

2.3 **Hydrology**

As discussed above the Kaikorai Stream borders the landfill to the north and west. Summary flow statistics for the Kaikorai Stream are provided below in Table 2.3 (NIWA, 20231). Further details on the surface water environment

¹NIWA River Maps online view: https://shiny.niwa.co.nz/nzrivermaps/ (accessed 09/02/2023)

are provided in the Surface Water Technical Assessment (GHD, 2023). From time to time, the estuary/ river mouth is blocked as a result of sand/debris accumulation from storm events. When this occurs, water backs up in the stream and results in higher stream levels adjacent to the site. Elevated water levels are maintained until the mouth is opened up again, either by natural or mechanical methods.

Table 2.3 Kaikorai Stream Flow Statistics (source NIWA)

Location	Mean flow	Mean Annual Low Flow
Upstream of Abbots Creek confluence	227 L/s	49 L/s
Downstream of Abbots Creek confluence	368 L/s	81 L/s

Geology 2.4

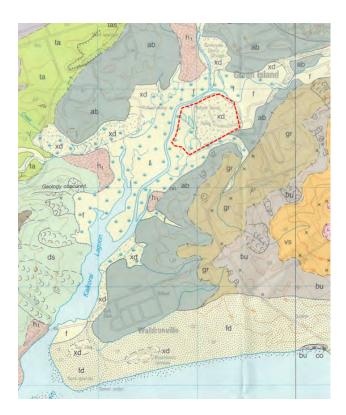
241 Overview

The site is situated on estuarine sedimentary deposits associated with the Kaikorai Estuary. These have been characterised as the Kaikorai Estuary Formation (KEF) (BDGC, 2002). KEF was considered to extend to a depth of approximately 11 m in the landfill area, and immediately overlies the Abbotsford Formation mudstone (BDGC, 2002). Mudstone occurs near the surface at the southeast corner of the landfill site (Figure 2.6).

The KEF was divided into an upper and lower layer (Member), with the upper member being further divided by BDGC (2020) into two subgroups as shown in Table 2.4. The estuarine sediments are underlain by Abbotsford Formation mudstone.

Table 2.4 Description of KEF lithological units (after BDGC, 2002)

Member	Description	Subgroup	Thickness
Upper Kaikorai Estuary Member	Variable thin beds of sand, silty sand, sandy silt, silt,	Subgroup A -mostly homogeneous fine grained	4.5 m
(UKEM)	clayey silt and silty clay	Subgroup B – heterogeneous, coarser grain size	
Lower Kaikorai Estuary Member (LKEM)	Massive homogeneous beds of clayey silt, silty clay and silt, and minor (possibly localised) beds of clay, very fine sandy silt and silty very fine sand.	-	6.5 m



Geological map of Green Island area (McKellar, 1990). Approximate site boundary shown by red dashed line. ab: Abbotsford mudstone, f: alluvium, xd:fill

2.4.2 GHD site investigation (2022)

Site investigations were undertaken by GHD to inform the geotechnical and hydrogeological assessments. The results of the site investigation are documented in the Geotechnical Factual Report (GHD 2023). A brief summary of the geology encountered is included here. The GHD investigation comprised twelve bore holes across the site, with piezometers installed in six of the bore holes. The general geological profile for boreholes on site perimeter is summarised in Table 2.5 below. There was no clear geological distinction between the two subgroups of the Upper Kaikorai Estuary Member (UKEM), therefore we have not adopted this division. In several of the bore holes there was a coarse grained layer (sand and/or gravel) at the contact of the Lower Kaikorai Estuary Member (LKEM) and mudstone. Depending on the location, variable amounts of fill were encountered. At BH104, drilled in the south-eastern part of the site, fill sits directly on mudstone.

Figure 2.6

Table 2.5 Summary geological profile – perimeter (near leachate trench) boreholes (Geotechnical Factual Report, GHD 2023)

Geology	Description	Layer thickness
Fill	Variable waste and soil	variable
UKEM	Silty fine to medium sand, sandy silt	1 – 3 m
LKEM	Organic silt, silty clay	6 – 8.5
Coarse sediments	Sands/gravel	0.5 -1.5
Abbotsford Mudstone	Grey-brown mudstone, very weak	-

2.5 **Hydrogeology**

2.5.1 Overview

Prior to the landfill, groundwater within the estuarine deposits (KEF) is likely to have been hydraulically connected to the Kaikorai Stream and other surface water features. As discussed in Section 2.2.1, pumping from the perimeter leachate trench creates a hydraulic barrier between surface water and the shallow aquifer underlying the landfill. The underlying Abbotsford Formation mudstone is inferred to be an aquitard due to the very low permeability of the mudstone and effectively an impermeable barrier for downward seepage. Adams Geotechnical (2019) reported permeability for the Abbortsford Formation mudstone associated with the capping material borrow pit located to the east of the landfill, with permeabilities as low as 1 x 10⁻¹⁰ m/s.

2.5.2 Hydraulic Conductivity

Groundwater investigations undertaken as part of the 1992 EIA (Beca, 1992), included permeability testing in three boreholes at various depths. The results indicated an average hydraulic conductivity for the estuarine sediments of 4 x 10⁻⁶ m/s above 4 m depth and 6 x 10⁻⁷ m/s for greater than 4 m depth.

Hydraulic conductivity estimates of the differing geological units at the site, determined from hydraulic testing during the GHD site investigation (Appendix A) are presented in Table 2.6. The testing resulted in a range of hydraulic conductivity values, this reflects the heterogeneity created by the depositional environment. There are likely to be discrete channels of higher permeability materials (where active stream channels were located), with both horizontal (due to the direction of deposition) and vertical anisotropy² (due to the layering of sediments) likely.

Table 2.6	Hydraulic conductivity summary
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Geology	Monitoring wells*	Hydraulic Conductivity Range (m/s)	Adopted Hydraulic Conductivity (m/s)
Upper Kaikorai Estuary Member (UKEM)	MW1C, MW5C, MW6C	8.4 x 10 ⁻⁷ to 2.8 x 10 ⁻⁶	1 x 10 ⁻⁶
Lower Kaikorai Estuary Member (LKEM)	BH100, BH101, BH103, BH108, MW2D, MW8C	6.2 x 10 ⁻¹⁰ to 3.3 x 10 ⁻⁶	1 x 10 ⁻⁷
Abbotsford Mudstone (AM)	BH104	< 1 x 10 ⁻⁹	1 x 10 ⁻⁹

2.5.3 Groundwater monitoring

2.5.3.1 **Monitoring network**

Monitoring of water levels and water quality is undertaken on a routine basis in accordance with the conditions of the current consents. The monitoring programme includes sampling of surface water, groundwater and leachate. A review of the surface water quality data is presented in the Surface Water Technical Assessment (GHD, 2023), and surface water quality will only be discussed here where relevant to the groundwater assessment. The groundwater monitoring network comprises:

- Eight lines of groundwater monitoring wells transecting the leachate collection trench, as shown in Figure 2.7.
- Each Well Line is located at mid-distance between two pump stations and each line comprises three shallow wells, MWA through to MWC, with the exception of Line 7, where MWC is absent (noting that the majority of these are founded in the UKEM geological unit).
- At each Well Line, monitoring wells MWA and MWB are located on the landfill side of the leachate trench, approximately 20 m and 5 m from the trench respectively.
- Monitoring well MWC is located between the trench and the Kaikorai Stream / eastern sedimentation pond / eastern boundary.
- Along each Well Line, an inspection manhole is located at the point the Well Line intersects the leachate trench, between monitoring wells MWB and MWC.
- On three of the Well Lines (Well Line 2, 4 and 7), deep wells are also present and monitored, located between the leachate collection trench and the stream. They are described as MWD (and founded in the LKEM geological unit).
- An additional bore, MW0C located at the end of the leachate trench collection system at Well Line 0, to the south of PS1 is also monitored.

² Where the hydraulic conductivity of sediments is different in two directions. le. in the horizontal direction compared to the vertical direction, in layered sedimentary units the vertical to horizontal hydraulic conductivity ratio is commonly 0.1.

A further monitoring well, MW9D, had been located towards the centre of the landfill, but was lost due to landfilling activities in 2015.

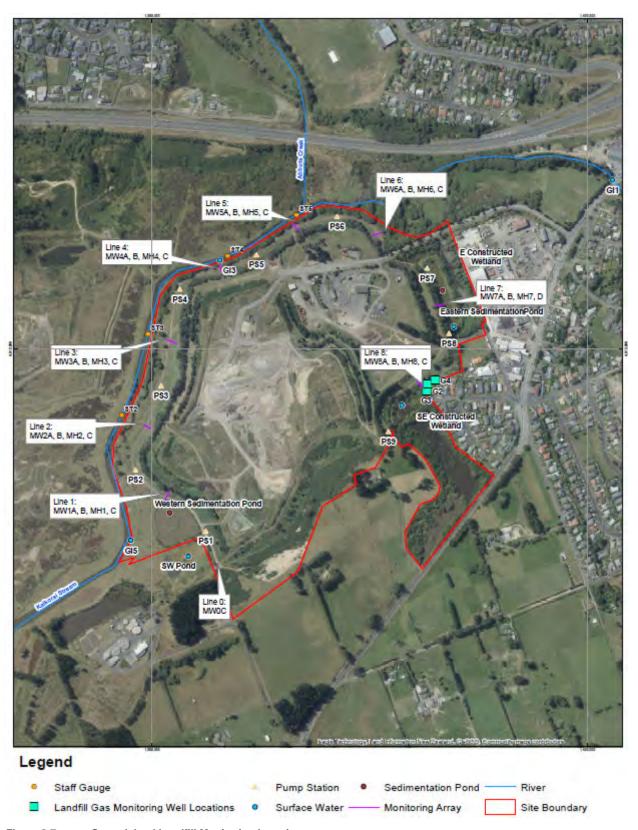


Figure 2.7 Green Island Landfill Monitoring Locations

A schematic of the monitoring Wells Lines is shown in Figure 2.8.

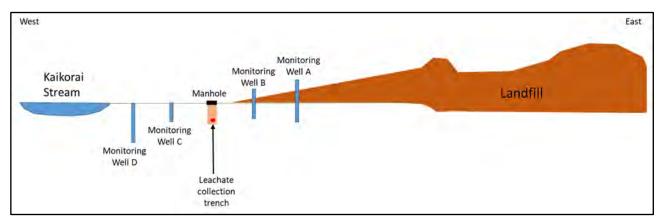


Figure 2.8 Schematic of monitoring well transect

2.5.3.2 **Groundwater levels**

The leachate collection trench was designed to create a hydraulic barrier by reducing groundwater levels in the trench and drawing groundwater from both sides of the trench (i.e. landfill and stream side, although the stream side inflows were designed to be retarded by the installation of a HDPE liner on the stream side of the trench). Groundwater levels are checked on a monthly basis to monitor the hydraulic gradient. The water level monitoring consistently shows that the lowest groundwater levels (in monitoring wells) occur adjacent to the trench (MWC wells). These results are shown in the Green Island annual monitoring report (GHD, 2022 and attached to the Surface Water Report (GHD 2023)) with a selection (representing winter, spring, summer, and autumn, included in Appendix B. These groundwater levels have been monitored monthly and reported annually to the ORC since the commencement of these consents in 1994.

Long term groundwater level records were reviewed as part of this assessment. Groundwater level data is available for the period from 1995 to 2003 and from 2015 to present. In general:

- Groundwater levels fluctuate within a range for each well, with no long-term trend in groundwater levels evident (with the exception of MW4D discussed below, Figure 2.9); and
- Seasonal variation is evident in the record for some wells (e.g. MW3C and MW6C, Figure 2.10). Groundwater levels are generally lowest in drier periods (summer-autumn), with groundwater highs occurring winter/spring and large rainfall events (e.g. January 2021).

Groundwater levels in MW4D showed an increasing trend from 1997 to 2003 (Figure 2.9). Due to the increasing groundwater head, the monitoring well casing was extended higher above ground to ensure no leakage of groundwater from the top of the casing (artesian conditions). There is some uncertainty about the datum used for 2015-2016 measurements, however the data shows that groundwater levels appear to have stabilised around 102-102.5 m RL³ in recent years. MW4D was drilled to a depth of 10.5 m depth. There is no bore log available for the well, but the geology encountered in a nearby borehole, BH100 (Geotechnical Factual Report GHD,2023), suggests that MW4D is screened over a sand and gravel layer present between the base of the LKEM and mudstone.

³ Note throughout this report two datums are used. On older figures/drawings a DCC Design Datum of AMSL +100m is used (hence a 1994 flood level of 103.3m). More recent data and the design drawings for this study use NZVD2016 as the datum and are referred to through this report as "amsl"

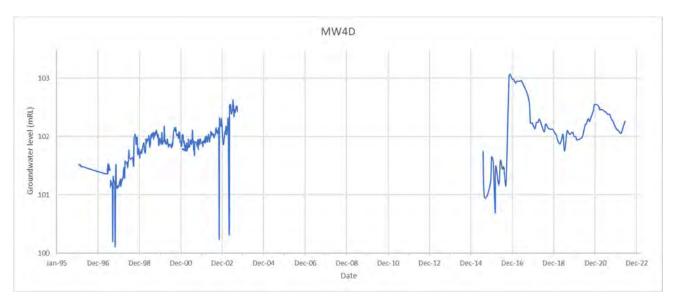


Figure 2.9 Groundwater levels measured in MW4D

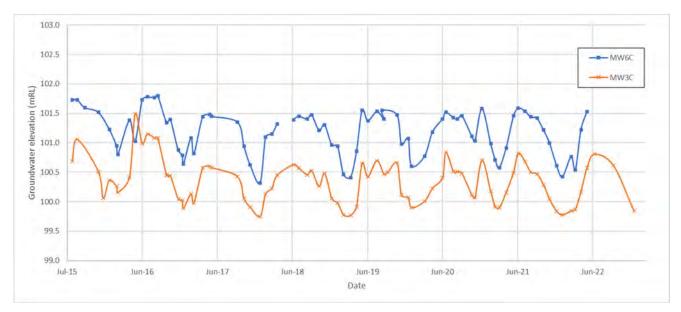


Figure 2.10 Groundwater levels measured in MW3C and MW6C - 2015 to 2022

2.5.3.3 Groundwater quality

Groundwater quality is monitoried on a quarterly basis in accordance with the consenting conditions, and results are reported in annual monitoring report, the most recent being the 2021-2022 report (GHD, 2022). The annual report provides a full summary of water quality trends and comparison against relevant standards/guidelines. A brief overview of the data is provided below with selected monitoring plots included in Appendix C.

In general, the monitoring wells inside the leachate trench (A and B series) are likely to be impacted by leachate. However, some of the wells on the outside of the trench also show the influence of landfill waste. A review of the site history and waste distribution was undertaken (included as Appendix D) which identified areas where waste has been placed outside of the leachate trench. The likely extent of waste on the outside of the trench is shown on Figure 2.11. Based on this review, the following "C" wells are likely to be within or influenced by historical waste materials; 8C, 7C, 6C, 4C, and 3C. Of the "D" wells, 7D is only 7 m deep and may be within waste. Monitoring well 4D, at 10.5 deep, is likely to be screened below the waste materials.

The impact of waste outside of the trench is managed through operation of the leachate collection system, which pulls groundwater (and any leachate) from both sides of the trench. This is discussed further in Section 4.



Water quality trends

A review of water quality data was undertaken to inform the conceptual understanding of the groundwater system and interactions with leachate/water. The water quality trends and patterns for relevant parameters are summarised below in Table 2.7. Water quality plots are provided in Appendix C.

Table 2.7 Water quality trends in groundwater

Parameter	Groundwater trends
Electrical Conductivity	Elevated in all monitoring wells relative to typical background groundwater Deep wells – highest in 2D and 4D (also higher chloride in these wells) Shallow wells – no clear pattern between A/B/C wells
Dissolved oxygen	Dissolved oxygen in groundwater is low with most samples <10% oxygen saturation, in contrast most surface samples are > 50%. The Eastern Constructed Wetland and South Eastern Constructed Wetland exhibit a wide variation in dissolved oxygen content.
Ammoniacal nitrogen	Generally elevated in groundwater relative to surface water with the exception of Eastern Constructed Wetland;
	Deep wells – highest in 2D (range of 14-21 mg/L in 2022) and 4D (0.5-10.7 mg/L), compared to 7D (<1.3 mg/L)
	C wells – elevated in 5C (14-21 mg/L), 2C (9-13 mg/L) and 4C (5.3-9.8 mg/L), the rest of the C monitoring wells recorded concentrations <5 mg/L
Chromium	Most groundwater concentrations < 0.002 mg/L, the exception is MW5C with chromium between 0.0052 – 0.012 mg/L in 2022
	Groundwater chromium concentration is generally lower than site surface water (such as W and E Sediment Ponds) but elevated compared to Kaikorai Stream
Boron	Boron concentration highest in 1C (~4 mg/L), 5C (~3 mg/L) and 4C (~2 mg/L).
	Deep well concentrations is highest in 7D (~1.4 mg/L), 4D and 2D <0.8 mg/l
	Boron elevated in E Constructed wetland (up to 9 mg/L), estuary concentration up to 1.8 mg/L, rest <1 mg/L (note boron analysis not undertaken in GI1, GI2, GI3, GI5)
Arsenic	Highest groundwater concentrations measured in 2D. Most results < 0.005 mg/L with the exception of 6C and 7D
	Groundwater and site surface water concentrations in similar range
Iron	Elevated in groundwater, in particular 4C, 2C, 6C, 5C, and deep monitoring wells.
	Highest concentrations recorded in 2D (116 mg/L)
	Iron concentration in groundwater an order of magnitude higher than site surface water and two orders of magnitude higher than Abbots Creek/Kaikorai Stream (GI1-GI5)

Major ion chemistry

The major ion chemistry for selected water samples is shown graphically in a piper (trilinear) plot (Figure 2.12). The piper plot shows relative proportions of the major anion and cation species within the water samples and is used to display the differences (or similarities) between different water types. The plot shows the following water types:

- Sodium bicarbonate type leachate
- Sodium chloride groundwater and estuary
- Mixed water types pump station and ponds
- Magnesium bicarbonate groundwater (MW3C and MW8C)

The piper plot shows that the samples from PS3 are an intermediate water type between leachate and groundwater. This is unsurprising given that the leachate trench intercepts both groundwater and landfill leachate. The Eastern and Western Sediment Pond samples also show an intermediate chemistry. These ponds are influenced by stormwater runoff from the landfill.

Groundwater from monitoring wells potentially influenced by waste (see above review of waste distribution) show a more varied water chemistry compared to the three wells unlikely to be influenced by landfill waste (MW2C, 2D, 4D). Major ion chemistry is not available for Kaikorai Stream (SW sample sites GI1-5), however a sample from the estuary at low tide is most similar to monitoring well MW4C. The estuary sample is likely to be influenced by activities in the wider catchment.

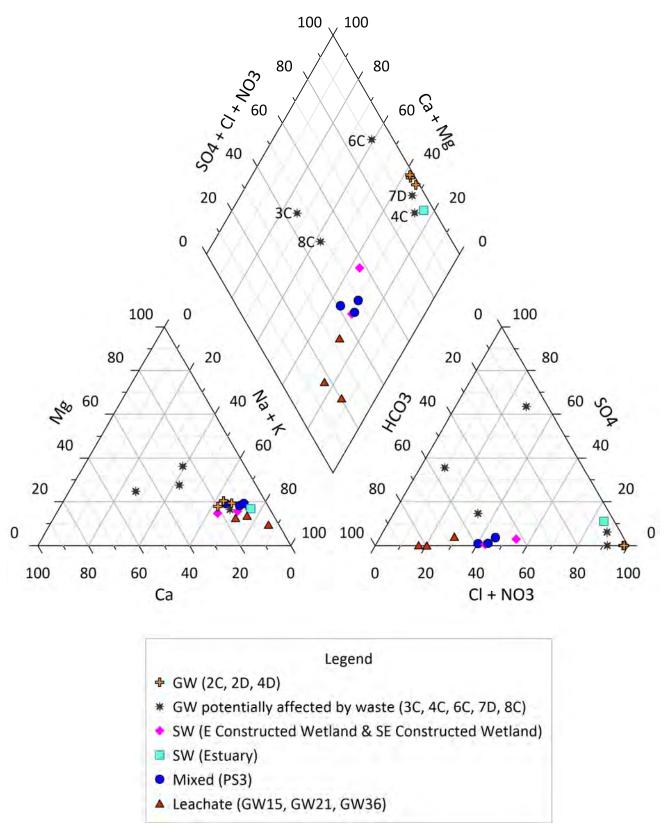


Figure 2.12 Relative proportion of major ions in GILF water samples

Isotopic Analysis

Isotopic analysis of samples from pump stations (leachate/groundwater mix), selected surface water and groundwater wells is undertaken in accordance with the consent, and the results and interpretive isotopes are included in the annual monitoring report. The relative proportions of stable isotopes can change through chemical processes, such as evaporation, diffusion, or chemical reactions, and can be used to understand mixing of different water types. The following isotopes are analysed:

- Oxygen-18 in water from leachate (δ18O-H2O), relative to Vienna standard mean ocean water.
- Hydrogen-2 in water from leachate (δD- H2O), relative to Vienna standard mean ocean water.
- Carbon-13 in dissolved inorganic carbon from leachate (δ13C- DIC), relative to Vienna Pee Dee Belemite.
- Nitrogen-15 in ammonium from leachate (δ15N-NH4+), relative to atmospheric nitrogen (from October 2019).

The data below are presented relative to the local Meteoric Water Line⁴. A number of biogeochemical and physical processes can result in waters plotting above or below the MWL.

The isotopic analysis indicates the following:

- ¹³C concentrations have been relatively stable with no one sampling location deviating greatly. A slight net increase in ¹³C concentrations can be observed in the surface water and monitoring well MW4D data sets.
 The enriched ¹³C data for leachate is a by-product of methane producing bacteria which use the lighter ¹²C to form CH₄ (Hackley & Liu, 1996).
- The record for ¹⁵N-NH₄⁺ isotope analysis is relatively short (data available from October 2019). There is some variability in the date but no clear pattern or trends in the available data set.

The isotopic signature for hydrogen and oxygen shows

- Groundwater from MW2D and 4D plot close together and generally within ±5% of the Dunedin Meteoric Water line.
- The majority of the surface water data plots with in MWL ±5% lines.
- The majority of the leachate data points sit as a cluster above the Dunedin MWL -5% line.
- Isotopic signature of pump station samples (leachate/groundwater) suggest a mature stage of leachate methanogenesis (interaction with older waste).

The isotope data indicates that leachate is not influencing deep groundwater or surface water. The oxygen/hydrogen isotopic signature of the leachate is relatively distinct, plotting well above the Dunedin MWL compared to other samples.

⁴ The Global Meteoric Water Line (MWL) describes the global annual average relationship between hydrogen and oxygen isotope (2H and 18O) ratios in natural meteoric waters (water derived from precipitation). The Dunedin MWL has been adapted for local conditions by North and Frew (2006).

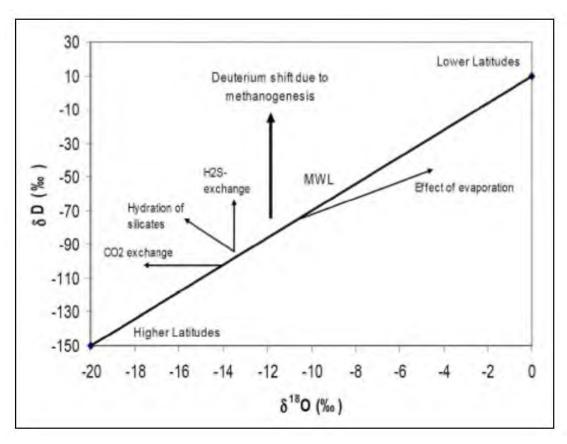


Figure 2.13 Deviations from the Meteoric Water Line caused by various biogeochemical and physical processes (from North and Frew , 2006)

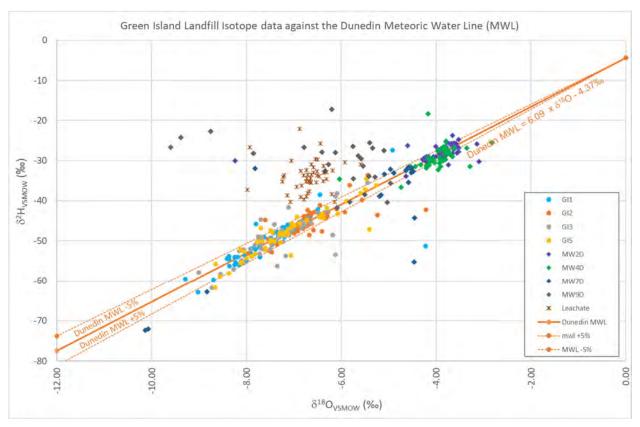


Figure 2.14 Green Island Landfill Isotope data against the Dunedin Meteoric Water Line (MWL)

PFAS

During late 2022 and early 2023 additional water sampling was undertaken for the presence of Persistent Organic Pollutants (POP), specifically PFOS and PFOA (i.e. perfluoroalkyl and polyfluoroalkyl substances⁵). The sampling was undertaken as landfills and industrial activities are a known sources of these contaminants. Water samples were collected from the perimeter groundwater monitoring wells, surface water monitoring sites, sedimentation ponds, and the leachate collection system. The results from the groundwater samples are provided in Table 2.8 below and the location of the sampling sites are shown in Figure 2.7, with the full set of results provided in Appendix C of the *Surface Water Report (2023)*.

Table 2.8 PFAS and PFOA Concentrations in GW monitoring wells

Site	October 2022		January 2023	
	Total PFAS (ug/L)	PFOA (ug/L)	Total PFAS (ug/L)	PFOA (ug/L
MW0C	0.003	<0.001	<0.001	<0.001
MW1C	<0.001	<0.001	<0.001	<0.001
MW2C	0.004	<0.001	<0.001	<0.001
MW2D	0.003	<0.001	<0.001	<0.001
MW3C	0.009	0.01	0.0073	0.011
MW4C	15	0.069	<0.001	<0.001
MW4D	0.006	<0.001	<0.001	<0.001
MW5C	0.008	0.007	<0.001	0.0088
MW6C	0.006	0.002	<0.001	<0.001
MW7D	0.005	<0.001	<0.001	<0.001
MW8C	0.012	<0.001	0.0019	<0.001
PS3 (leachate)	0.061	0.08	0.057	0.11

With the exception of the result reported for MW4C for October 2022, the concentrations of Total PFAS in the perimeter groundwater wells are at low concentrations and are below the 95% species protection limits of 0.13 ug/L, defined in the PFAS National Environmental Management Plan Version 2.0 – known as NEPM V2.0). With the exception of MW3C, the results from the January 2023 sampling round were below the laboratory detection limit of 0.001ug/L.

The concentration of Total PFAS and PFOA obtained from PS3, which is representative of leachate mixed with groundwater, was consistent for both sampling events, with concentrations recorded an order of magnitude above the groundwater samples (with the notable exception of MW4C, October 2022). If leachate was migrating beyond the leachate trench it would be reasonable to expect that the concentrations of PFAS in the groundwater perimeter wells would be similar to the leachate PS3 sample.

The concentration of Total PFOS in MW4C for October 2022 was significantly higher than other results obtained from the groundwater wells (i.e. four orders of magnitude greater than Kaikorai Estuary results). There was also a high concentration for Total PFAS reported for the Western Pond in October 2022, which is located outside the landfill site and not connected to landfill water or leachate management infrastructure. It is possible that these results are a laboratory error or sampling error. Further monitoring would assist in resolving this area of uncertainty.

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⁵ Synthetic chemicals found in many manufactured products

Summary of water quality data

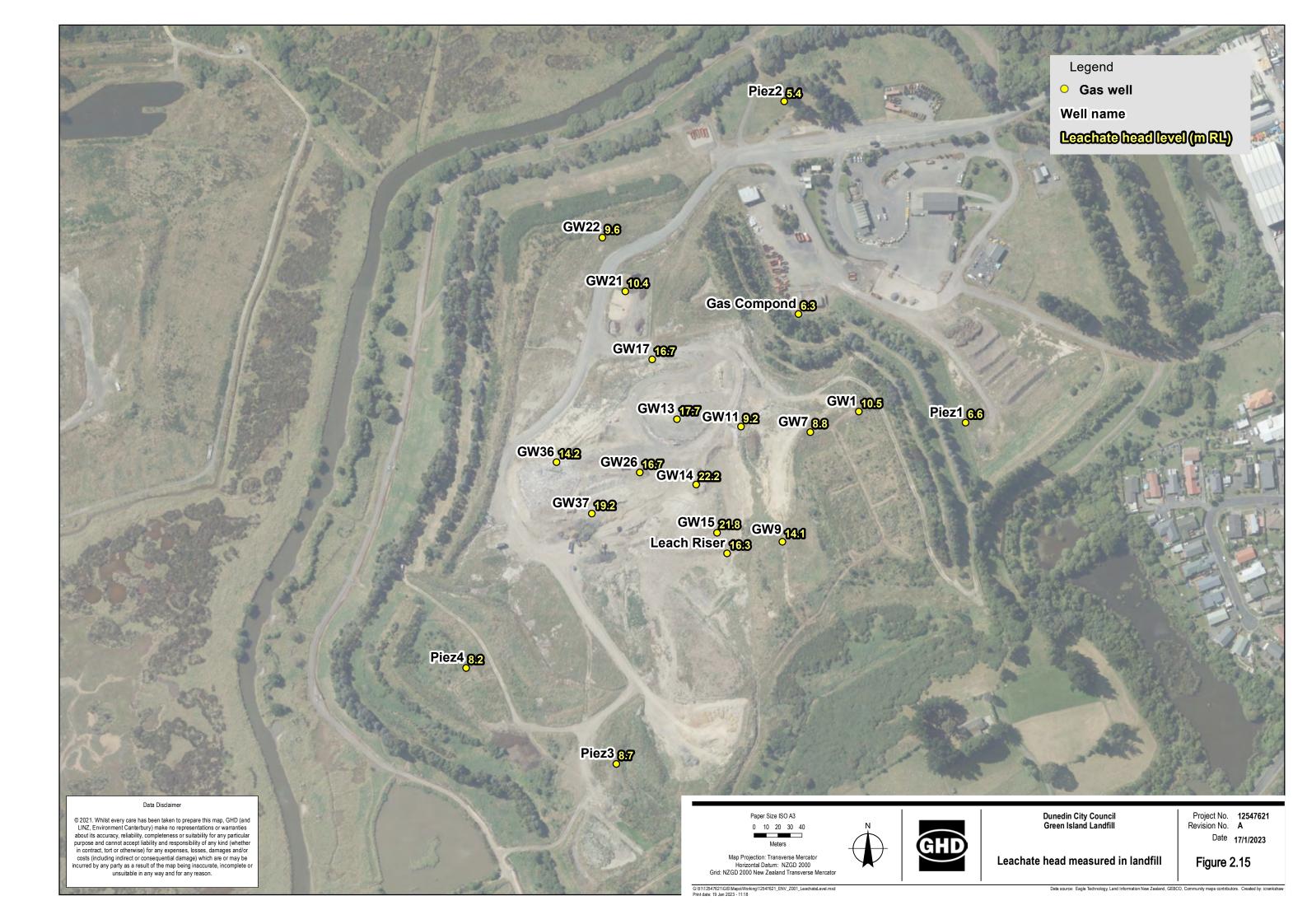
The water chemistry data shows the influence of landfill waste on groundwater quality. In areas where waste is present outside of the leachate trench, the groundwater quality shows a mixed major ion signature with elevated contaminants.

The major ion chemistry clearly shows mixing of groundwater and landfill leachate in water pumped from the leachate trench.

The depositional environment, and relatively recent (in geological terms) change from estuarine to a freshwater setting influences the groundwater chemistry. In estuarine and shallow marine sediments elevated chloride, alkalinity and boron is likely due to the influence of sea water. Ammoniacal nitrogen (Ammoniacal-N) and iron are elevated in many of the groundwater samples, including monitoring wells unlikely to be influenced by waste. The elevated ammoniacal-N and iron in background groundwater may reflect the influence of the organic material in the estuarine sediments (KEF) and reducing conditions in the aquifer. Electrical conductivity is elevated in all samples, reflecting the influence of leachate and/or brackish water in the Kaikorai estuary.

2.5.4 Leachate levels within the landfill

Leachate levels were measured in the landfill by dipping existing gas wells and groundwater monitoring wells within the landfill footprint. The leachate survey was completed in August 2022 by a site contractor. These data are included in Appendix D and shown in Figure 2.15. There is some variability in leachate measurements, this is likely due to the heterogeneity of the fill. As of August 2022, the leachate level in the centre of the landfill is average 16 or 17 amsl over completed areas of landfill and cap, with two wells over 20 m amsl.



3. Technical assessment

The following section documents the conceptual understanding of the groundwater system, interaction of groundwater with the landfill leachate and the leachate collection system, and interaction with surface water. This conceptual understanding, site data and investigation results, have been used to assess the effect of the proposed activities. The technical assessment includes:

- An assessment of rainfall infiltration through the current and proposed cap (HELP modelling)
- An assessment of leachate head within the fill (2D SEEP/W groundwater modelling)
- An assessment of the effectiveness of the leachate trench in intercepting leachate (SEEP/W)
- Groundwater surface water interaction

Detailed summaries of the modelling assessments are provided in and Appendix F (HELP modelling) Appendix G (Seep/W).

3.1 Conceptual model

The conceptual understanding of the groundwater system is summarised below and shown in Figure 3.1. Figure 2.2The model is based on published information, routine monitoring, and previous site investigations, in summary:

- Infiltration of rainfall into the landfill, generation of leachate as water comes into contact with waste;
- Migration of leachate down and outwards towards the edge of the landfill;
- Abstraction of groundwater and leachate from leachate trench, water chemistry confirms mixing of water types;
- Groundwater quality influenced by historical waste deposition; this includes areas outside of the trench (e.g MW4C). However, continuous pumping from the leachate trench maintains the hydraulic gradient and pulls the impacted groundwater towards the trench (and away from surface water);
- Stream depletion effects limited by presence of HDPE liner on stream side of trench. However, some abstraction of shallow groundwater from the outer edge of the site (and stream water) is likely to occur;
- Upward gradient in LKEM restricting migration of leachate into deeper layers under the trench collection system; and
- Underlying mudstone forms an aquitard limiting deeper flow paths and restricting flow to the south (southern valley) where it occurs at/near the surface.

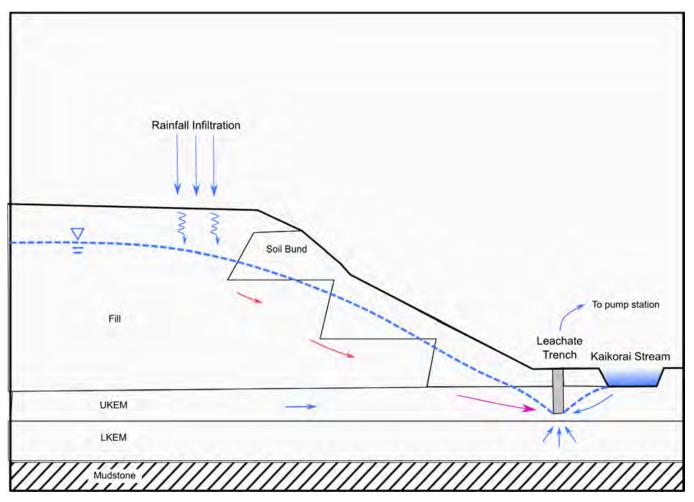


Figure 3.1 Conceptual model

3.2 Modelling assessment

The modelling assessment utilised two different methods:

- HELP (Hydrologic Evaluation of Landfill Performance)
- SEEP/W 2D groundwater model

Rainfall infiltration through the landfill cap was assessed using the Hydrologic Evaluation of Landfill Performance (HELP) software (Berger and Schroeder, 2013). HELP 3.95D is a quasi-two-dimensional hydrologic model for conducting water balance analysis of landfills and cover systems. The model utilises weather, soil and landfill design data to account for the effects of surface storage, runoff, infiltration, evapotranspiration, soil moisture, lateral subsurface drainage, vertical drainage and leakage through soil and liners. The modelled leakage (infiltration) through the current and future landfill cap was included in the SEEP/W modelling assessment. The HELP modelling assessment is included in Appendix F. HELP can also be used to calculate leachate leakage through the base of the landfill, however this is typically applied to lined landfills. Given the complex history of the Green Island Landfill site and absence of a modern landfill liner system, it was considered that the SEEP/W was more suitable for representing the interaction between the landfill, including the leachate volumes and head level within the landfill, and the receiving environment.

The groundwater assessment included 2D modelling using Geostudio 2021 SEEP/W finite element modelling software. Modelling was undertaken to estimate the seepage into the leachate collection drain and to simulate the leachate head within the landfill.

Two SEEP/W cross-sections were created to model the landfill. The location of the cross-section lines is shown in Figure G.1 (Appendix G). The models were created based on a drone survey provided by DCC. Each model was initially run under steady-state conditions to simulate the interpreted baseline groundwater conditions. Models were calibrated to:

- Measured leachate head within the landfill (average level); and
- Combined pump station flows (from leachate collection trench), under dry weather conditions.

Steady state model scenarios were then run to simulate future conditions at closure, in particular, capping of the landfill and installation of a leachate collection drain in the southern valley. The Line 2 future scenarios include additional filling as outlined in the *Design Report* (GHD 2023). Details of the model set up and assumptions are provided in Appendix G. Model scenarios are included in Table 3.1 below.

Table 3.1 Model scenarios

Scenario Number	Hydraulic conductivity of ca	Hydraulic conductivity of cap	
	Currently uncapped*	Capped areas (north/central, applies- Section Line 1 only)	
Current	-	-	Leachate trenches operational
Current (trench off)	-	-	Trench" turned off"
1A	1 x 10 ⁻⁷ m/s	1 x 10 ⁻⁸ m/s	Leachate trenches operational
1B	1 x 10 ⁻⁷ m/s	1 x 10 ⁻⁸ m/s	Trench" turned off"
1C	1 x 10 ⁻⁷ m/s	1 x 10 ⁻⁸ m/s	Trench operational, Sea level rise (+ 0.5 m) applied to river boundary.
2A	1 x 10 ⁻⁸ m/s (all areas)	1 x 10 ⁻⁸ m/s (all areas)	
2B	1 x 10 ⁻⁸ m/s (all areas)	1 x 10 ⁻⁸ m/s (all areas)	
2C	1 x 10 ⁻⁸ m/s (all areas)	1 x 10 ⁻⁸ m/s (all areas)	
3A	1 x 10 ⁻⁹ m/s (all areas)	1 x 10 ⁻⁹ m/s (all areas)	

3.2.1 Results

The results of the modelling for the base case (current) and future simulations are summarised in Table 3.2.

The modelling indicates flows to the leachate trench of approximately 1 L/s, this does not include the flows reporting to pump stations 5 to 8, which are expected to flow into the trench at a lower rate (per metre of trench) than the sections modelled. However, the results are in line with the current recorded flows (section 2.2.3).

The modelled leachate head for the base case scenarios are similar field measurements, although the very high leachate in the centre of the landfill could not be simulated. It is likely that the high leachate levels reflect the heterogeneity of the landfill materials resulting in pockets/areas of higher leachate.

Table 3.2 Summary modelling results – leachate trench flow rate and leachate head

Scenario	Model Section	Perimeter leachate drain flow rate for length of trench modelled* (L/s)	Southern Valley leachate drain flow rate for length of trench modelled* (L/s)	Modelling leachate head in fill (m RL)
Current	Line 1	0.5		16
	Line 2	0.5		10.2
1A	Line 1	0.5	0.1	15.4
	Line 2	0.4	0.2	9.8
1C	Line 1	0.6	0.1	15.4

Scenario	Model Section	Perimeter leachate drain flow rate for length of trench modelled* (L/s)	Southern Valley leachate drain flow rate for length of trench modelled* (L/s)	Modelling leachate head in fill (m RL)
	Line 2	0.5	0.2	9.9
2A	Line 1	0.5	0.1	14
	Line 2	0.4	0.1	9.3
2C	Line 1	0.6	0.1	14
	Line 2	0.4	0.1	9.3
3A	Line 1	0.3	0.01	4
	Line 2	0.2	0.02	3

^{*}Flow rate (per metre) multiplied by length of trench represented by each section:

Line 1 – Perimeter trench, 400 m, Southern Valley trench (150 m)

Line 2 – Perimeter trench, 620 m, Southern Valley trench (300 m)

The relative proportion of flows from each side of the trench was calculated from the model. For Section 1, the modelling results indicate that approximately 70% of the flow to the trench comes from the landfill (and underlying groundwater) and 30% from the direction of the stream. Stream depletion rates are estimated to be < 0.1 L/s.

If the leachate trench pumps were to fail for an extended period (i.e. several weeks), the modelling shows a reversal of flow over time with flow from the landfill into the stream eventually at an estimated rate of 0.5-0.8 L/s (combined rate). This equates to approximately 0.2% of the mean flow in the Kaikorai Stream, downstream of Abbotts Creek confluence.

By undertaking a simple mass mixing model approach the potential impact on water quality in the Kaikorai Stream from a prolonged loss of the hydraulic containment can be estimated (Table 3.3). At mean flow, using the mixing model approach, the Ammoniacal-N concentration would increase by an order of magnitude when compared to existing background concentrations, whilst at low flows the concentration would be higher still.

Table 3.3 Simple mass mixing model for GW Ammoniacal-N discharge to Kaikorai Stream

Leachate impacted groundwater	Value			
Groundwater discharge (mean) (m³/day)	69.1			
PS3 Ammoniacal -N (g/m³)	323			
Kaikorai Stream	Mean Flow Conditions	Low Flow Conditions		
Stream Flow (below Abbots Creek) (m³/day)	31,795	6,998		
Background Ammoniacal-N (g/m³)	0.05	0.05		
Mixed Concentrations if Discharge Occurred				
Mixed Ammoniacal concentration (g/m³)	0.75	3.2		

It is understood that when pump failure has occurred previously, it has taken weeks for leachate levels to rise within the trench (pers comm. L Coe). The monthly water level monitoring data from the perimeter wells and leachate manholes that has been collected over the years provides a good evidentiary basis for showing that the trench has been effective in maintaining a hydraulic barrier. It is designed to enable gravity flow of leachate along the trench to a pump station, with the ability of leachate to bypass a pump station in the event of an isolated pump shut down. In addition, it is unlikely that the entire system would fail at once for a prolonged period of time, where remedial actions were not able to be instated to maintain the hydraulic barrier. Therefore, the effects to surface water as shown by the model scenario are unlikely to be realised for short term failures.

The Section Line 2 model indicates a very small flow from groundwater to the stream (0.01 L/s) this is interpreted to be from the estuarine sediments adjacent to and under the stream bed. The stream is located ~70 m from the

trench, therefore while the trench appears be effective at intercepting leachate from the landfill, it does not result in stream depletion and does not influence groundwater levels immediately adjacent to the stream.

A summary of the key finding from the modelling assessment is provided below in Table 3.4.

Table 3.4 Modelling summary

Feature	Summary
Cap permeability	Decrease in leachate flow rate and leachate head level when a lower permeability cap is applied. This effect is most pronounced when a cap hydraulic conductivity (K) of 10 ⁻⁹ m/s is applied. The difference between leachate flow and head for the K 10 ⁻⁷ and 10 ⁻⁸ m/s scenarios is small.
River side leachate trench	The trench is effective at drawing down leachate levels and intercepting flows to the stream. When the trench drainage boundary condition is "turned off" for modelling purposes, the flow rate to stream is estimated to be between 0.5-0.8 L/s.
	Modelled flow rates are similar to dry weather trench pump rates
Southern valley trench	Estimated flow rates to the proposed trench are 0.2-0.3 L/s (0.03 L/s for low permeability cap scenario)
Surface water interaction	The modelling indicates that stream water (and shallow groundwater) is intercepted by the leachate trench. Stream depletion rates are estimated to be in the order of 0.1 L/s for Section 1. Stream depletion rates are higher in the sea level rise scenarios due to the higher water levels.
	However, the Section 2 model indicates flow (at a very low rate, 0.01 L/s) into the stream from shallow groundwater adjacent to the stream. This is much lower than the modelled flow to the stream (0.3-0.5 L/s) if the trench was "turned off" showing that the trench is effective at intercepting leachate. The small inflow (0.01 L/s) to the stream under operating conditions is interpreted to represent groundwater seeping from the stream bed and adjacent river bank. Due to the distance of the trench to the stream, the operation of the leachate trench does not influence groundwater levels immediately adjacent to the stream.

Assessment of effects on the environment 4_

Introduction 4.1

The proposed filling and closure of the Green Island Landfill has the potential to affect water quantity and quality. An assessment of effects addressed in this technical assessment includes:

- Effects to groundwater levels and flow;
- Effects to surface water flows;
- Effect of climate change and sea level rise: and
- Effects to water quality.

The water quality assessment is focussed on groundwater quality with consideration of surface water interaction. Surface water quality is primarily addressed in Surface Water Technical Assessment (GHD, 2023).

Effect to groundwater and surface water levels and 4.2 flows

The leachate collection system operates by drawing down water levels in the trench, this intercepts any leachate flowing from the landfill but also draws groundwater from the area outside of the trench. It is likely that this groundwater is hydraulically connected with surface water in the Kaikorai Stream, with the potential for groundwater abstraction to have a stream depleting effect.

The modelling assessment presented in Section 3 and Appendix F, indicates that approximately 30% of the water pumped from the leachate trench is derived from groundwater / connected surface water on the outside of the trench, in areas where the trench is close to the stream. This volume is estimated to be <0.5 L/s for the entire trench length. For areas further from the active stream channel, eg PS3, the modelling indicates no effect of the leachate trench operation on stream flows.

Overall, the assessment indicates that effect of the leachate trench abstraction on surface water flows is negligible when compared to the stream flows, even during low flow conditions, and volume of water in the estuary.

The underlying KEF and Abbotsford mudstone are not used for groundwater supply, therefore the abstraction of groundwater and localised reduction in groundwater levels around the landfill perimeter does not affect any groundwater users.

4.3 Effect of climate change and sea level rise

Modelling was undertaken to assess the effect of higher river levels on the leachate collection system. The modelling incorporated a higher road level (as recommend in the Design Report) and a higher river level (increase by 0.5 m). The model results indicated slightly higher inflows to the leachate system, this increase is well within the operating range of the leachate system, which can accommodate much higher stormwater flows.

Effects to water quality 4.4

This assessment has shown the perimeter leachate collection system is effective at creating a hydraulic barrier and intercepting leachate flowing from the landfill. It also draws groundwater from outside of the gravel leachate trench, this is of particular importance in areas where waste is present outside of the trench (from historic activities). This abstraction prevents the movement of potentially contaminated groundwater from outside the leachate trench, into surface water. As discussed in the Surface Water Assessment, surface water monitoring shows that there are no discernible adverse effects on surface water quality from the landfill activities.

The modelling assessment was used to assess the effect on groundwater flow rates and flow paths should pumping from the leachate trench cease for a prolonged period resulting in discharge to surface water. This modelling scenario is considered very unlikely in reality, as there would be a time lag for leachate levels to rise in the trench before the modelled flow rate was achieved. Furthermore, it is considered unlikely that all pump stations would be out of action for an extended period given historical performance and additional mitigation measures that are proposed in the Design Report. However for the purposes of this assessment, if leachate migration to surface water was occurring (at a rate of 0.5 – 0.8 L/s) we would expect to see a measurable change in water quality between the surface water monitoring points GI2 and GI3 (during low flow conditions) due to the very high contaminant load in the landfill leachate (as described in section 3.2). Instead, the water quality at the downstream GI3 is generally better than GI2. This further supports the conclusions of modelling assessment that the leachate trench is effective at intercepting landfill leachate.

The proposal includes the installation of a leachate collection trench in the southern valley, this will intercept leachate flowing from the southern side of the landfill. However, there will still be a gap in the leachate trench between MH8 and PS9. This area of the landfill sits directly on a ridge of Abbotsford Mudstone. The mudstone forms an effectively impermeable barrier to flow, therefore leachate migration off site is unlikely.

5. Recommendations

5.1 **Monitoring**

It is recommended that both groundwater monitoring and surface water monitoring is continued in line with the current consent conditions, with the following exceptions:

- It is recommended that isotopic analysis is removed as a consent condition. The current consent requires quarterly sampling for analysis of environmental isotopes. It is considered that isotopic analysis does not significantly improve the understanding of the groundwater/surface water system, and given the long delay (months) between collection of samples and reporting of results, chemical analysis is likely to provide a more timely indication of leachate mobility.
- It is recommended that the requirement to install a deep groundwater well within the landfill is removed as a consent condition. The installation of a deep well has the potential to create preferential flow paths from fill to the underlying geology.

The following total peak and average daily flow rates are proposed as a consent condition for the leachate collection system. These maximum rates allow for both groundwater and diverted stormwater during potentially extended periods of rainfall:

- Peak rate 20 L/s (1,728 m³/day)
- Average rate 5 L/s (432 m³/day or 157,680 m³/year).

It is recommended that the following water level and quality monitoring is undertaken (Table 5.1).

A comparison to the existing monitoring schedule is outlined in Table 5.2. The proposed changes simplify the current monitoring program, with all water types having the same quarterly and annual analytical suites. Some analytes have been removed from the schedule, as these analytes (i.e. faecal coliforms, volatile fatty acids) were considered not to be useful for detecting leachate migration into the environment.

Table 5.1 Recommended monitoring

Frequency	Measurement/Analyte	Locations
Monthly	Groundwater levels	A / B / C / D wells, pump stations and manholes.
Quarterly ¹	pH Electrical Conductivity Dissolved oxygen Boron Nitrate Nitrogen Ammoniacal Nitrogen Chloride PFAS/PFOA ³	- C and D wells - Representative sample from the leachate trench (PS3) - Surface water (GI1, GI2, GI3, GI 5 and estuary) within three hours of low tide - Western sediment pond - South western pond - Eastern sediment pond - South eastern constructed wetland - Eastern constructed wetland
Annual	Major Ions (Sodium, Potassium, Magnesium, Calcium, Bicarbonate, Sulphate and Chloride) Dissolved Reactive Phosphorous Metals² (Aluminium Arsenic, Cadmium, Chromium, Iron, Lead, Manganese, Mercury, Nickel, Zinc) Volatile Organic Compounds (VOC) Semi Volatile Organic Compounds (SVOC) Cyanide	- C and D wells - Representative sample from the leachate trench (PS3) - Representative sample of leachate (from gas well in landfill) - Surface water (GI1, GI2, GI3, GI 5 and estuary) within three hours of low tide - Western sediment pond

Frequency	Measurement/Analyte	Locations
	Chemical oxygen demand (COD)	- South western pond
	Biological oxygen demand (BOD)	- Eastern sediment pond
	PFAS/PFOA (after three years of monitoring on quarterly basis)	- South eastern constructed wetland
		- Eastern constructed wetland

¹Reduced frequency to 6 monthly, two years post closure

Water quality results are to be documented and reviewed and assessed on receipt, with full analysis and reporting in the annual monitoring report due 1 October every year. The annual monitoring report is to include a discussion on any water quality trends and the effectiveness of the leachate trench in intercepting landfill leachate.

Post closure it is recommended that the monitoring programme is reviewed and updated to reflect the changes to the landfill.

²Metal analysis – dissolved metals in groundwater and leachate trench (PS3), total metals in surface water

³Reduced frequency to annual after 3 years

Table 5.2 Comparison between current and proposed monitoring schedule

Schedule	Surface Water	Surface Water			Groundwater			PS3 (Leachate)		
	Current	Proposed Additional	Proposed Removed	Current	Proposed Additional	Proposed Removed	Current	Proposed Additional	Proposed Removed	
Monthly				Groundwater levels						
Quarterly	pH Electrical Conductivity Chloride Dissolved Oxygen Ammoniacal Nitrogen Nitrate Nitrogen	PFAS/PFOA	Total Organic Carbon Isotope Analysis Dissolved Metals, (annual only) Total Cyanide (annual only)	pH Electrical Conductivity	Dissolved oxygen Boron Nitrate Nitrogen Ammoniacal Nitrogen Chloride PFAS/PFOA	Isotope Analysis	Isotope Analysis	pH Electrical Conductivity Dissolved Oxygen Boron (previously measured annually) Nitrate Nitrogen Ammoniacal Nitrogen Chloride PFAS/PFOA	Isotope Analysis	
Annual		Major Ions (Sodium, Potassium, Magnesium, Calcium, Bicarbonate, Sulphate and Chloride) Metals (Aluminium Arsenic, Cadmium, Chromium, Iron, Lead, Manganese, Nickel, Zinc, Mercury) Volatile Organic Compounds (VOC) Semi Volatile Organic Compounds (SVOC) PFAS/PFOA Cyanide Chemical Oxygen Demand (COD) Biological Oxygen Demand (BOD)		BOD Major Ions (Sodium, Potassium, Magnesium, Calcium, Bicarbonate, Carbonate, Sulphate and Chloride) Cation Anion Ratio pH Conductivity Ammoniacal Nitrogen Nitrate nitrogen Dissolved Iron Dissolved Lead Dissolved Zinc Dissolved Oxygen	PFAS/PFOA Cyanide Chemical oxygen demand (COD) Mercury Nickel Cadmium Chromium Aluminium Volatile Organic Compounds (VOC) Semi Volatile Organic Compounds (SVOC)	Total Organic Carbon	Major lons (Sodium, Potassium, Magnesium, Calcium, Bicarbonate, Carbonate, Sulphate and Chloride) Cation anion ratio pH Electrical Conductivity COD and BOD Ammoniacal nitrogen Nitrate nitrogen Dissolved Oxygen Dissolved Reactive Phosphorus (DRP) Total Organic Carbon Acid soluble metals, including: Aluminium, Arsenic, Barium Cadmium, Chromium, Copper, Iron, Lead, Nickel, Manganese, Zinc Total Mercury Total Cyanide Volatile Organic Compounds (VOC) Semi Volatile Organic Compounds (SVOC)	PFAS/PFOA	Barium Sulphide Total phenols Faecal Coliforms OCP Polychlorinated Biphenyls Volatile fatty acids	

Conclusions 6.

This technical groundwater assessment has involved a review of historical data, including the historic distribution of waste, site monitoring data, and recent site investigation data. This information was used to inform a groundwater modelling assessment to estimate the seepage into the leachate collection trench and to simulate the leachate head within the landfill. In summary:

- The leachate collection trench is effective in intercepting landfill leachate and drawing groundwater from areas of waste outside the trench;
- The chemical signature of waters pumped from the leachate trench show that it is a mix of groundwater and leachate;
- The future modelling scenarios indicate a small reduction in leachate head in recently capped area (northern part of the landfill) and an increase of leachate head in areas receiving additional fill;
- Modelling indicates that the leachate trench in the southern valley will intercept leachate coming from the landfill and aid to lower leachate levels within the landfill and enhance leachate collection;.
- Potential for a small increase in dry weather flows in the leachate collection system from the additional fill placement and installation of the southern valley trench;
- With the continuing operation of the leachate collection system, and maintenance of the groundwater hydraulic barrier, no discernible effect on surface water quality is expected; and
- The effect on surface water flows from groundwater abstraction is negligible.

7. References

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GHD has relied on information from a number of sources, including but not limited to the following:

Environmental Impact Assessment (Beca, 1992)

Green Island Landfill Leachate Collection Trench Geological Report (Barry J Douglas Geological Consultants, 2002 – herein referred to as BDGC, 2002)

Green Island Landfill Gas Management Project. Stage 1: Investigation Works. URS, 2007.

Bund stability report (T&T, 2020)

Leachate collection summary (DCC) and associated plans

Claymine Assessment (DCC) and associated plans

GILF & RRRP master plan (Stantec 2020)

Green Island Annual monitoring reports prepared by GHD, 2022

Appendices

Appendix A

Hydraulic conductivity testing

A-1 Hydraulic conductivity testing

Rising and falling head tests of existing monitoring wells and new bore holes were undertaken in November and December 2022 to estimate the hydraulic conductivity of the underlying shallow aquifer. Up to two sets of rising and falling head tests were completed in each well, however due to the slow groundwater response in the low permeability sediments, less tests were completed in some wells. The raw data was initially processed and reviewed graphically and then, if test results were considered suitable (e.g. not too many fluctuations, and suitable measured response and recovery time), data were interpreted with Aqtesolv software using Bouwer and Rice (1976) and Hvorslev (1951) solutions.

The hydraulic conductivity estimate for monitoring wells MW2D is based on historical data from 1997, when the well was purged and groundwater levels recovered over a few weeks.

The results of the hydraulic conductivity testing are summarised in Table A.1, full results are provided in Table A.2 with analysis plots following.

Table A.1 Summary of hydraulic conductivity testing

Geology	Monitoring wells*	Hydraulic Conductivity Range (m/s)	Adopted Hydraulic Conductivity (m/s)
Upper Kaikorai Estuary Member (UKEM)	MW1C, MW5C, MW6C	8.4 x 10 ⁻⁷ to 2.8 x 10 ⁻⁶	1 x 10 ⁻⁶
Lower Kaikorai Estuary Member (LKEM)	BH100, BH101, BH103, BH108, MW2D, MW8C	6.2 x 10 ⁻¹⁰ to 3.3 x 10 ⁻⁶	1 x 10 ⁻⁷
Abbotsford Mudstone (AM)	BH104	< 1 x 10 ⁻⁹	1 x 10 ⁻⁹

^{*}Logs not available for older MW monitoring wells, inferred geology based on recent drilling investigation.

Wells BH102, MW0C, MW2C, MW3C, MW4C, MW4D likely screened over two layers

Test in BH104 does not recover fully (<85% recovery) – results approximately only.

Table A.2 Hydraulic conductivity testing

Monitoring wells	Screen lithology	Hydraulic test methodology*	Solution	Hydraulic Conductivity (m/s)	Comment
		FHT	Bouwer-Rice	1.9 x 10 ⁻⁷	
		RHT	Bouwer-Rice	3.3 x 10 ⁻⁷	
		FHT	Bouwer-Rice	1.8 x 10 ⁻⁷	
NAVA/O.O.	Likely screened over two	RHT	Bouwer-Rice	3.4 x 10 ⁻⁷	
MW0C	layers	FHT	Hvorslev	2.3 x 10 ⁻⁷	
		RHT	Hvorslev	4.6 x 10 ⁻⁷	
		FHT	Hvorslev	1.9 x 10 ⁻⁷	
		RHT	Hvorslev	4.0 x 10 ⁻⁷	
		FHT	Bouwer-Rice	1.9 x 10 ⁻⁶	
		RHT	Bouwer-Rice	8.4 x 10 ⁻⁷	
		FHT	Bouwer-Rice	9.6 x 10 ⁻⁷	
NNA40	LUCENA	RHT	Bouwer-Rice	1.1 x 10 ⁻⁶	
MW1C	UKEM	FHT	Hvorslev	1.5 x 10 ⁻⁶	
		RHT	Hvorslev	8.8 x 10 ⁻⁷	
		FHT	Hvorslev	1.0 x 10 ⁻⁶	
		RHT	Hvorslev	1.3 x 10 ⁻⁶	
		FHT	Bouwer-Rice	3.9 x 10 ⁻⁷	
	Likely screened over two	FHT	Bouwer-Rice	4.0 x 10 ⁻⁷	
MW2C	layers	FHT	Hvorslev	5.3 x 10 ⁻⁷	
		FHT	Hvorslev	5.0 x 10 ⁻⁷	
		RHT following purging	Bouwer-Rice	6.2 x 10 ⁻¹⁰	Taken from purged data
MW2D	LKEM	RHT following purging	Hvorslev	7.2 x 10 ⁻¹⁰	(weekly dip readings) in 1997 used for an estimate
MW3C		FHT	Bouwer-Rice	1.1 x 10 ⁻⁷	
	Likely screened over two layers	RHT	Bouwer-Rice	1.3 x 10 ⁻⁷	
	layers	FHT	Bouwer-Rice	1.2 x 10 ⁻⁷	

Monitoring wells	Screen lithology	Hydraulic test methodology*	Solution	Hydraulic Conductivity (m/s)	Comment
		RHT	Bouwer-Rice	9.6 x 10 ⁻⁸	
		FHT	Hvorslev	1.6 x 10 ⁻⁷	
		RHT	Hvorslev	1.1 x 10 ⁻⁷	
		FHT	Hvorslev	1.5 x 10 ⁻⁷	
		RHT	Hvorslev	1.0 x 10 ⁻⁷	
		FHT	Bouwer-Rice	6.6 x 10 ⁻⁷	
MW4C	Likely screened over two	FHT	Bouwer-Rice	6.8 x 10 ⁻⁷	
WWW4C	layers	FHT	Hvorslev	7.7 x 10 ⁻⁷	
		FHT	Hvorslev	8.0 x 10 ⁻⁷	
		FHT	Bouwer-Rice	1.2 x 10 ⁻⁷	Appears to be influenced
MW4D	Likely screened over two	RHT	Bouwer-Rice	1.1 x 10 ⁻⁷	by pumping or other unaccounted for factor,
WIVV4D	layers	FHT	Hvorslev	6.7 x 10 ⁻⁷	potentially tidal
		RHT	Hvorslev	2.0 x 10 ⁻⁷	
		FHT	Bouwer-Rice	1.5 x 10 ⁻⁶	
		RHT	Bouwer-Rice	1.0 x 10 ⁻⁶	
		FHT	Bouwer-Rice	1.1 x 10 ⁻⁶	
MW5C	UKEM	RHT	Bouwer-Rice	1.3 x 10 ⁻⁶	
MIVVOC	UKEWI	FHT	Hvorslev	2.0 x 10 ⁻⁶	
		RHT	Hvorslev	1.2 10 ⁻⁶	
		FHT	Hvorslev	1.4 x 10 ⁻⁶	
		RHT	Hvorslev	1.6 x 10 ⁻⁶	
MW6C		FHT	Bouwer-Rice	1.0 x 10 ⁻⁶	
		RHT	Bouwer-Rice	2.3 x 10 ⁻⁶	
	UKEM	FHT	Bouwer-Rice	9.3 x 10 ⁻⁷	
	UNEIVI	RHT	Bouwer-Rice	2.5 x 10 ⁻⁶	
		FHT	Hvorslev	1.3 x 10 ⁻⁶	
		RHT	Hvorslev	2.3 x 10 ⁻⁶	

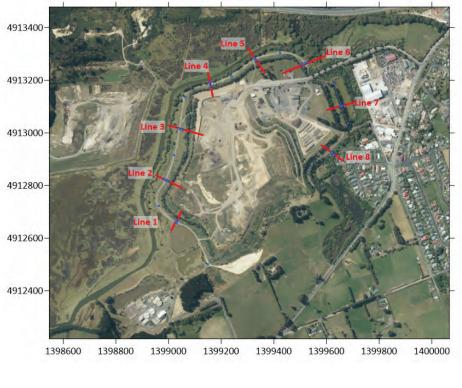
Monitoring wells	Screen lithology	Hydraulic test methodology*	Solution	Hydraulic Conductivity (m/s)	Comment
		FHT	Hvorslev	1.1 x 10 ⁻⁶	
		RHT	Hvorslev	2.8 x 10 ⁻⁶	
		RHT	Bouwer-Rice	2.4 x 10 ⁻⁶	
		FHT	Bouwer-Rice	3.3 x 10 ⁻⁶	
BH100	LKEM	RHT	Bouwer-Rice	2.2 x 10 ⁻⁶	
БПІОО	LNEW	RHT	Hvorslev	2.5 x 10 ⁻⁶	
		FHT	Hvorslev	2.3 x 10 ⁻⁶	
		RHT	Hvorslev	2.9 x 10 ⁻⁶	
		FHT	Bouwer-Rice	3.3 x 10 ⁻⁷	
		FHT	Bouwer-Rice	2.7 x 10 ⁻⁷	
BH101	LKEM	RHT	Bouwer-Rice	6.6 x 10 ⁻⁷	
БПІЛІ		FHT	Hvorslev	3.9 x 10 ⁻⁷	
		FHT	Hvorslev	3.6 x 10 ⁻⁷	
		RHT	Hvorslev	6.6 x 10 ⁻⁷	
		FHT	Bouwer-Rice	5.4 x 10 ⁻⁷	
		RHT	Bouwer-Rice	2.8 x 10 ⁻⁷	
		FHT	Bouwer-Rice	4.0 x 10 ⁻⁷	
BH102		RHT	Bouwer-Rice	2.8 x 10 ⁻⁷	
БП102	Screened over two layers	FHT	Hvorslev	6.6 x 10 ⁻⁷	
		RHT	Hvorslev	2.6 x 10 ⁻⁷	
		FHT	Hvorslev	4.5 x 10 ⁻⁷	
		RHT	Hvorslev	2.7 x 10 ⁻⁷	
BH103	IKEM	RHT	Bouwer-Rice	2.1 x 10 ⁻⁸	
סחוט	LKEM	RHT	Hvorslev	3.0 x 10 ⁻⁸	
	Mudstone	RHT	Bouwer-Rice	7.0 x 10 ⁻¹⁰	Does not recover to 85% of
		RHT	Hvorslev	8.0 x 10 ⁻¹⁰	displacement. Approximate only.
BH108	LKEM	FHT	Bouwer-Rice	3.1 x 10 ⁻⁸	

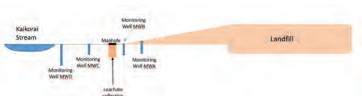
Monitoring wells	Screen lithology	Hydraulic test methodology*	Solution	Hydraulic Conductivity (m/s)	Comment
		RHT	Bouwer-Rice	2.2 x 10 ⁻⁸	
		FHT	Hvorslev	3.4 x 10 ⁻⁸	
		RHT	Hvorslev	1.5 x 10 ⁻⁸	

^{*}FHT/RHT = Falling/rising head test undertaken using displacement slug

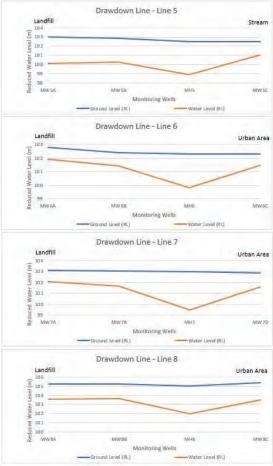
Appendix B

Groundwater level plots









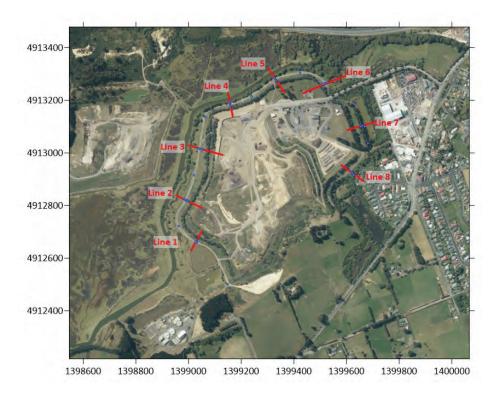


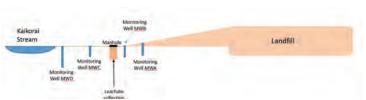
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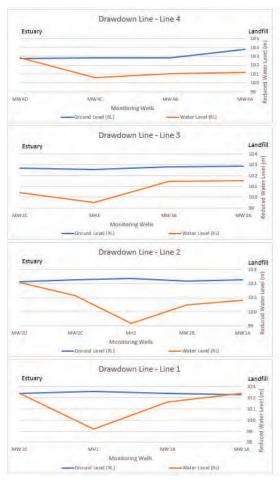
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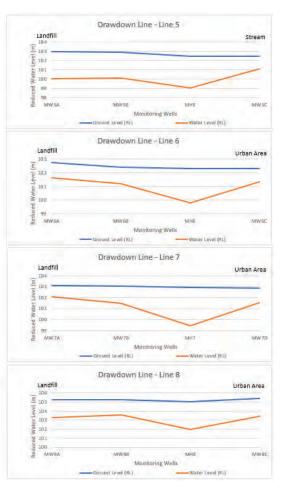
Figure B-1

Sources: Aerial (LINZ: Aerial Dunedin 0.4m Rural, 2013, NZGD2000); Computed using Green_Island_water_levels.xlsb.xlsx











Monthly Transects - Green Island October 2021

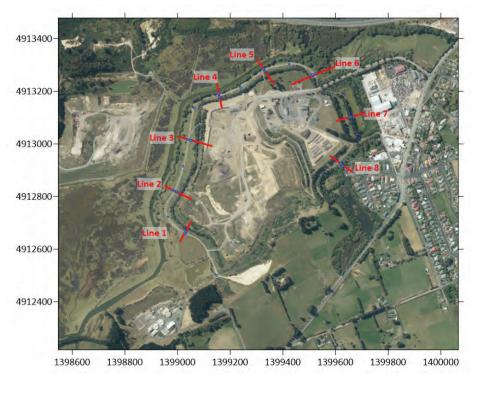
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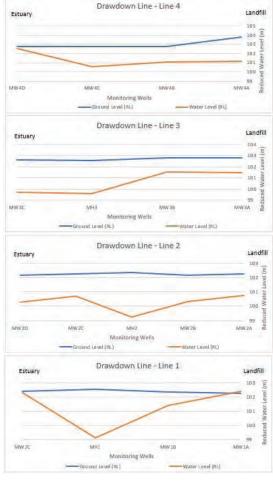
Figure B-2

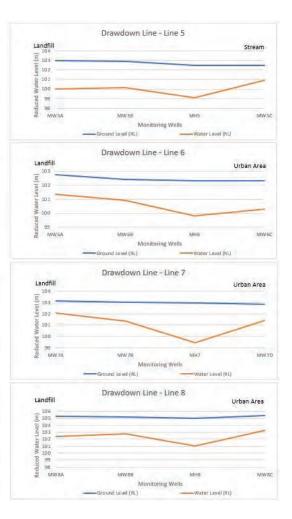
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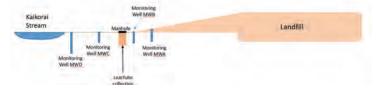
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Monthly Transects - Green Island February 2022

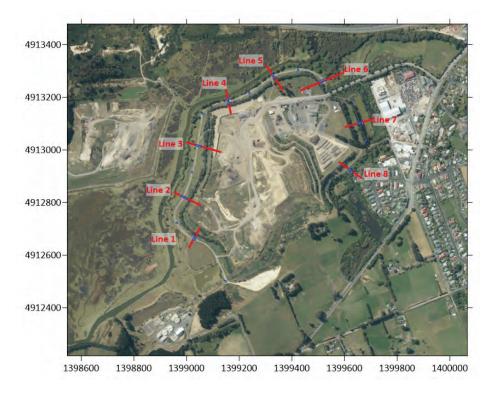
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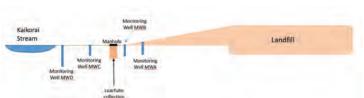
Figure B-3

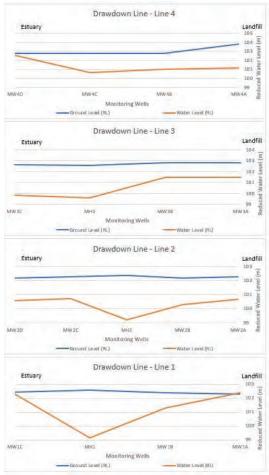
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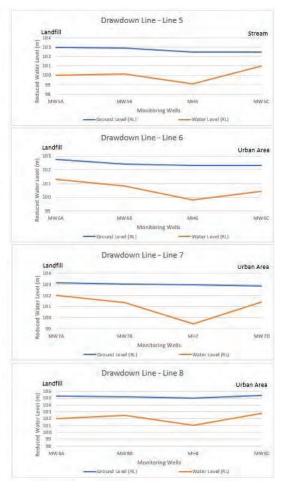
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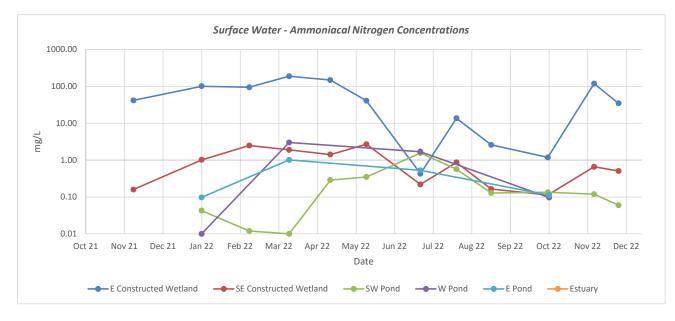
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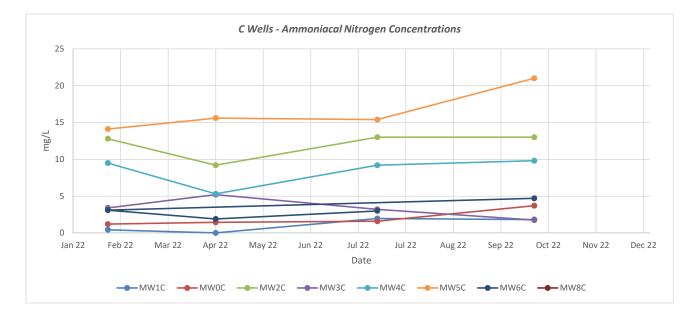
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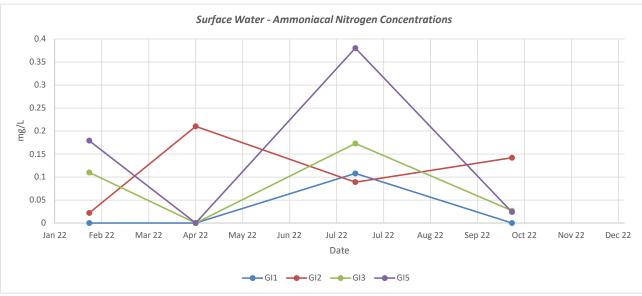
Figure B-4

Appendix C

Groundwater quality







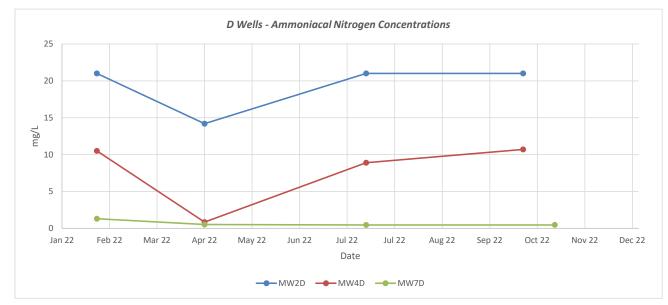
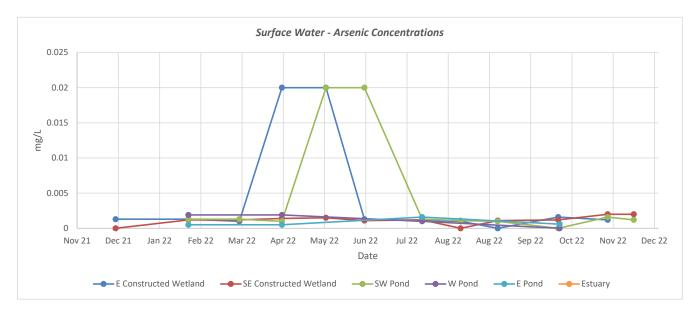
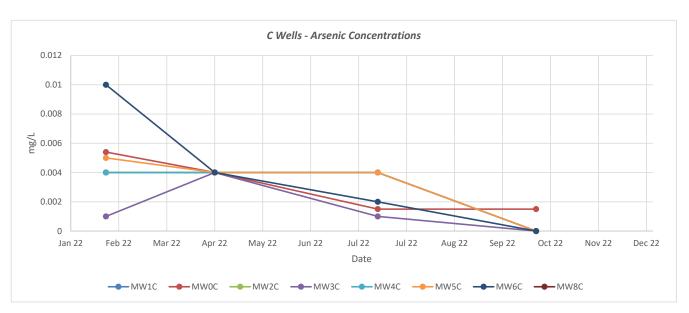
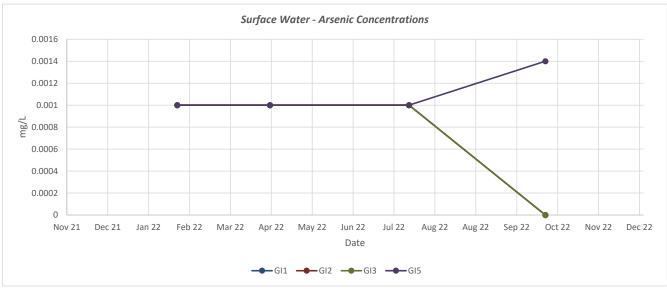
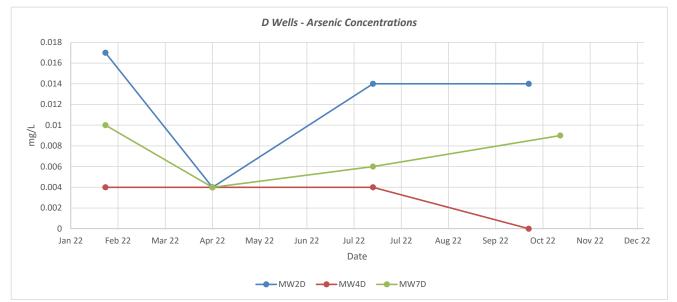


Figure C1-1 Ammonical Nitrogen



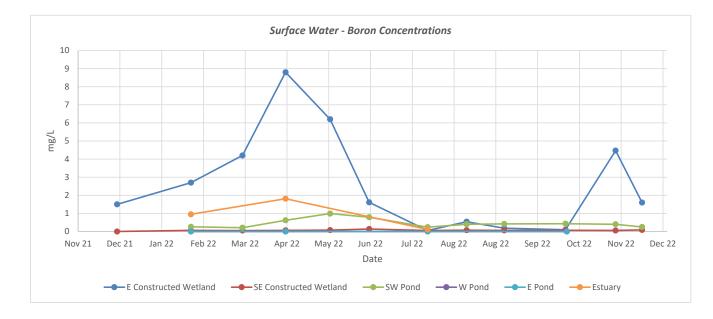






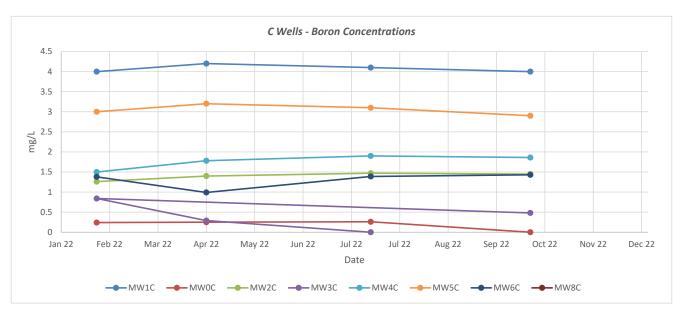
Note: All arsenic concentrations at locations GI1, GI2, GI3 & GI5 were less than the detection limit

Figure C1-2 Arsenic

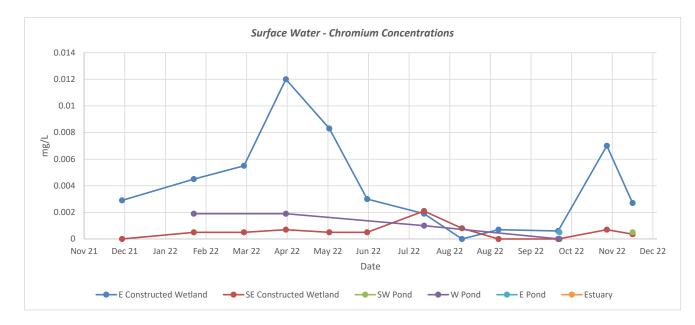


Note: Boron analysis not undertaken in samples GI1, GI2, GI3, GI5

Figure C1-3 Boron







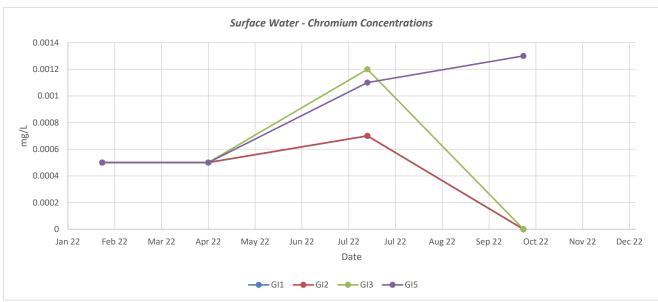
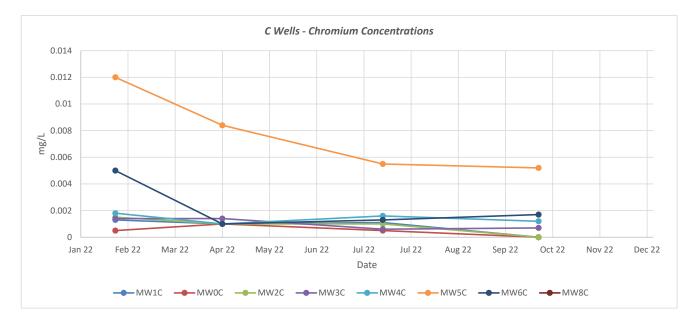
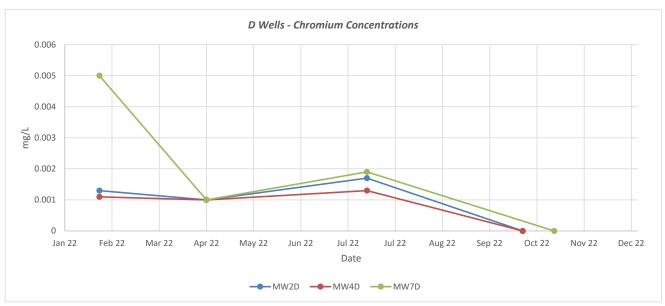
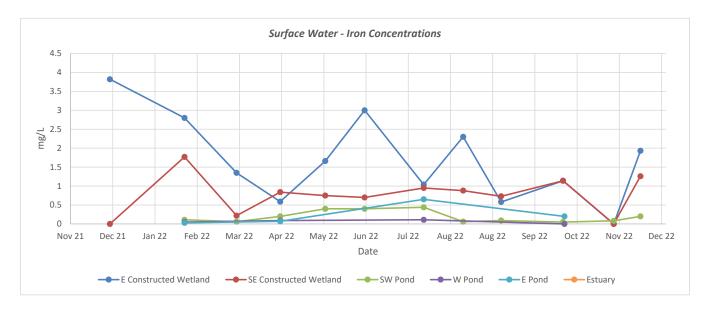
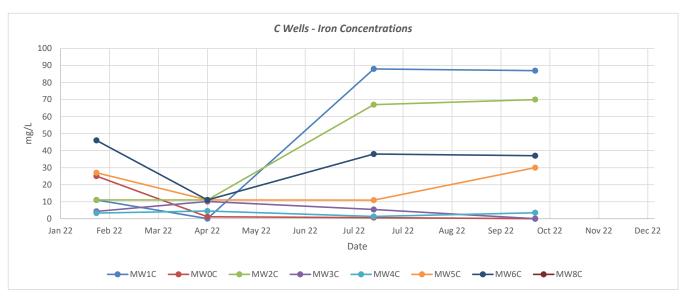


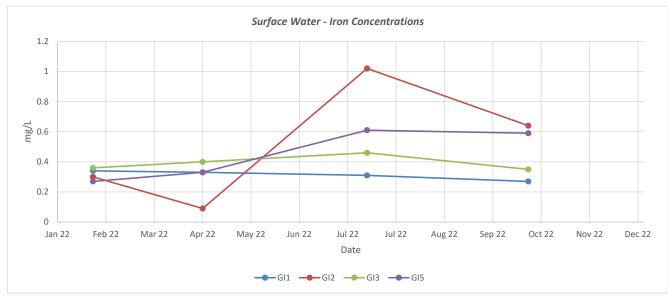
Figure C1-4 Chromium











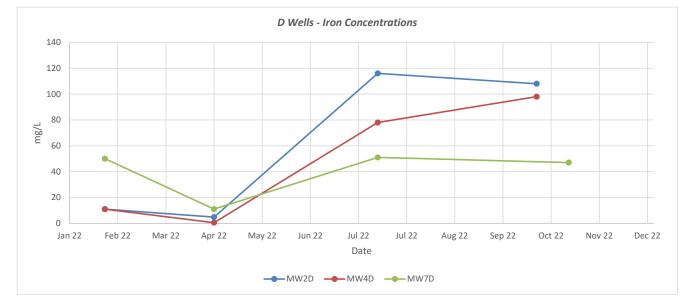


Figure C1-5 Iron

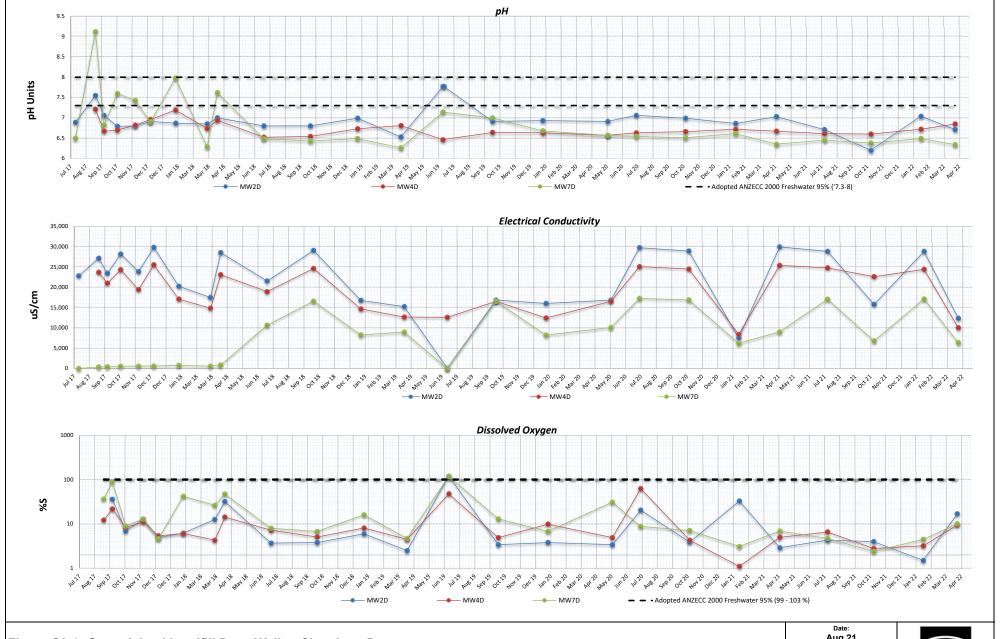


Figure C2-1: Green Island Landfill Deep Wells - Chemistry Data

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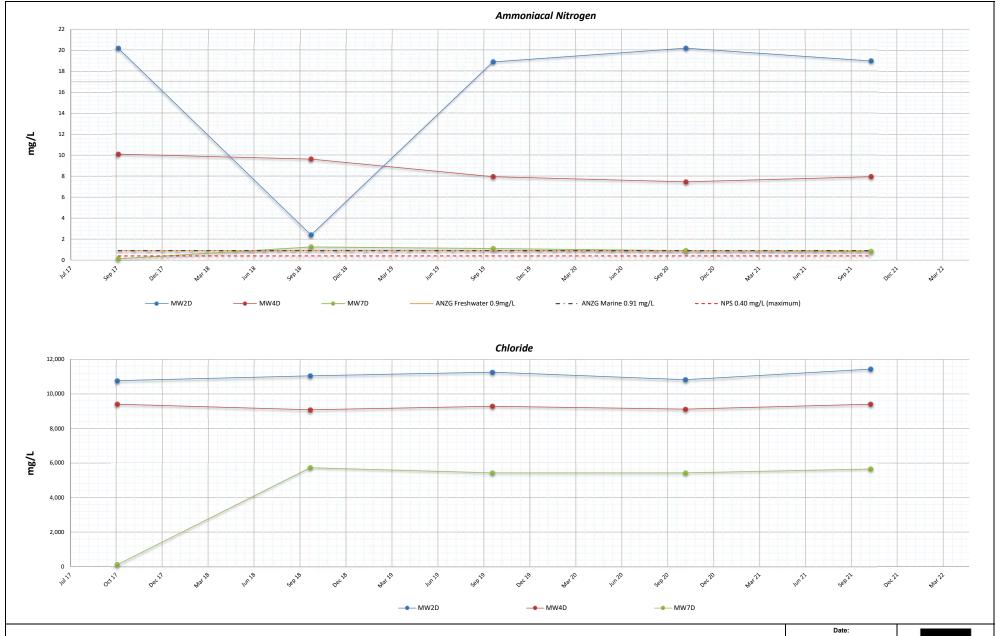


Figure C2-2: Green Island Landfill Deep Wells - Chemistry Data



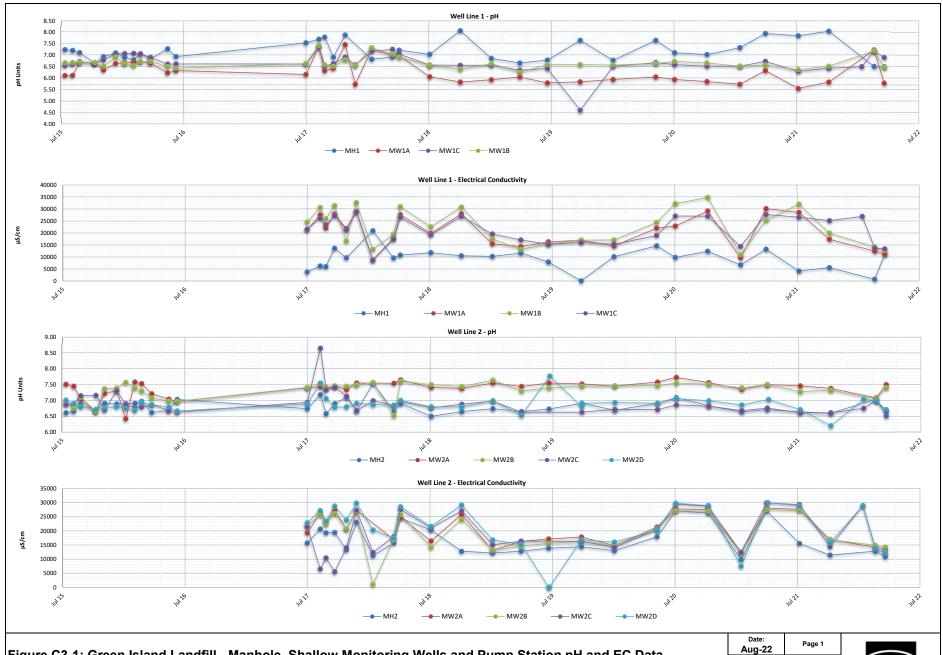
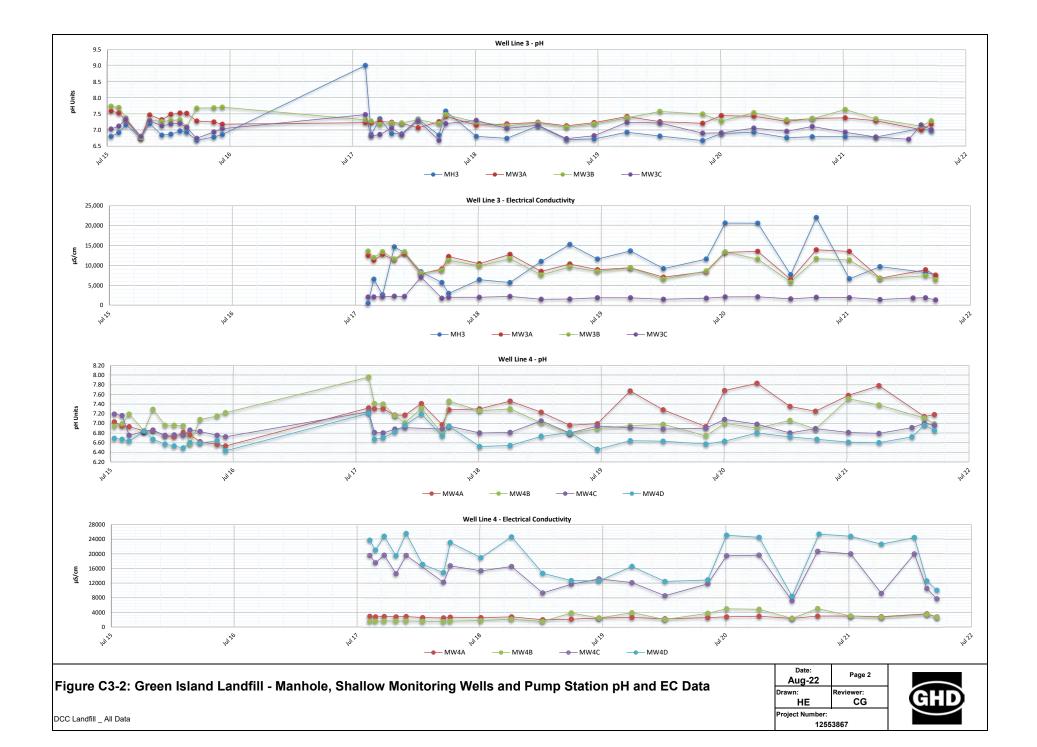
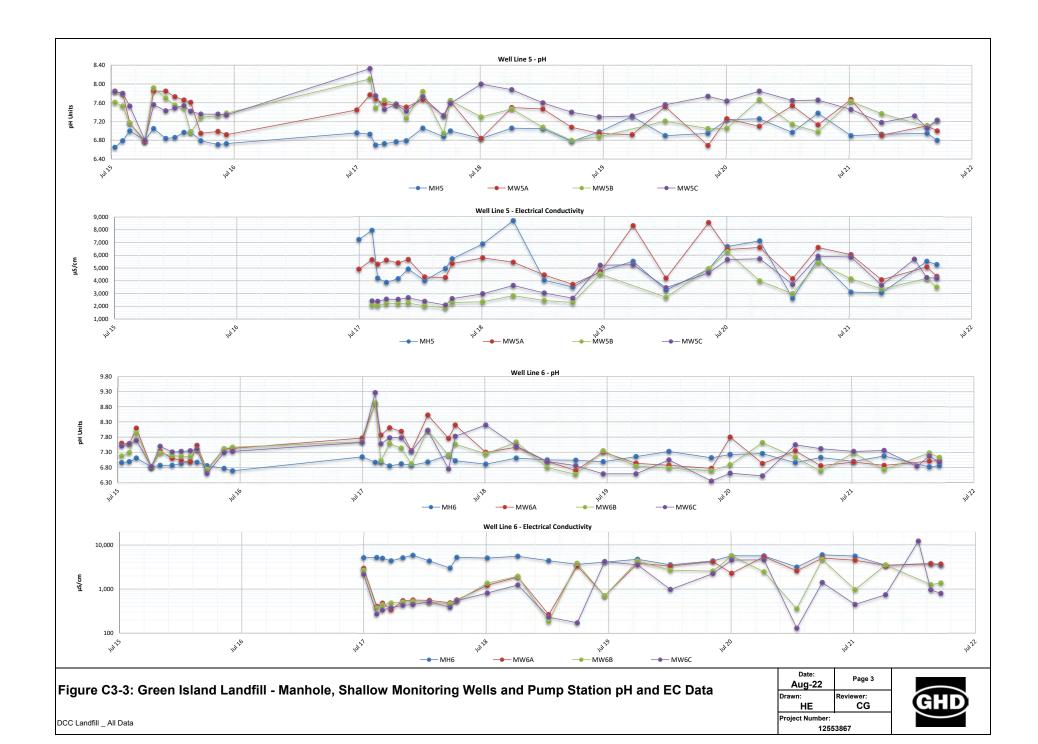


Figure C3-1: Green Island Landfill - Manhole, Shallow Monitoring Wells and Pump Station pH and EC Data







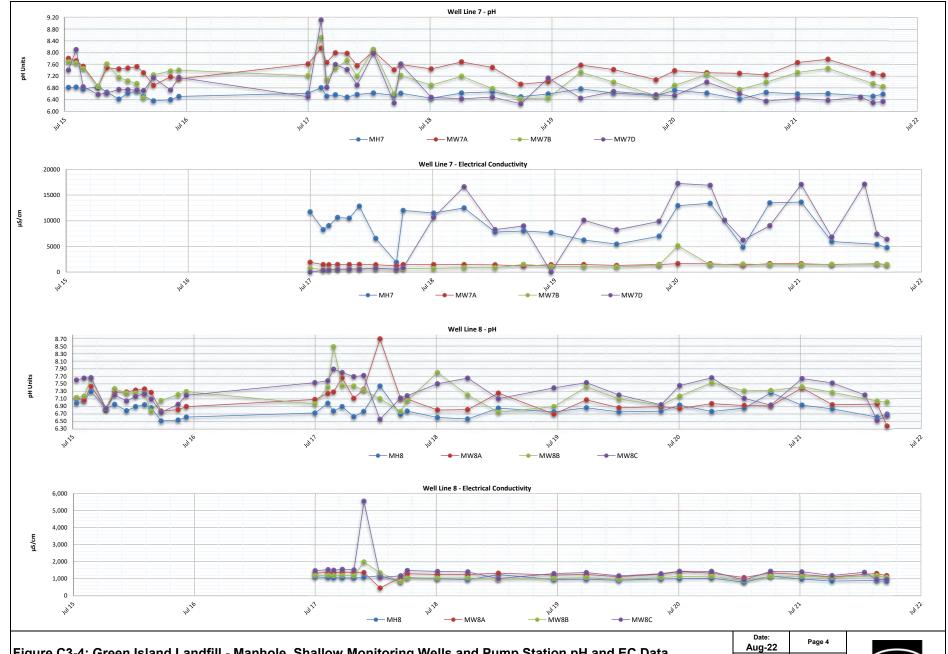


Figure C3-4: Green Island Landfill - Manhole, Shallow Monitoring Wells and Pump Station pH and EC Data



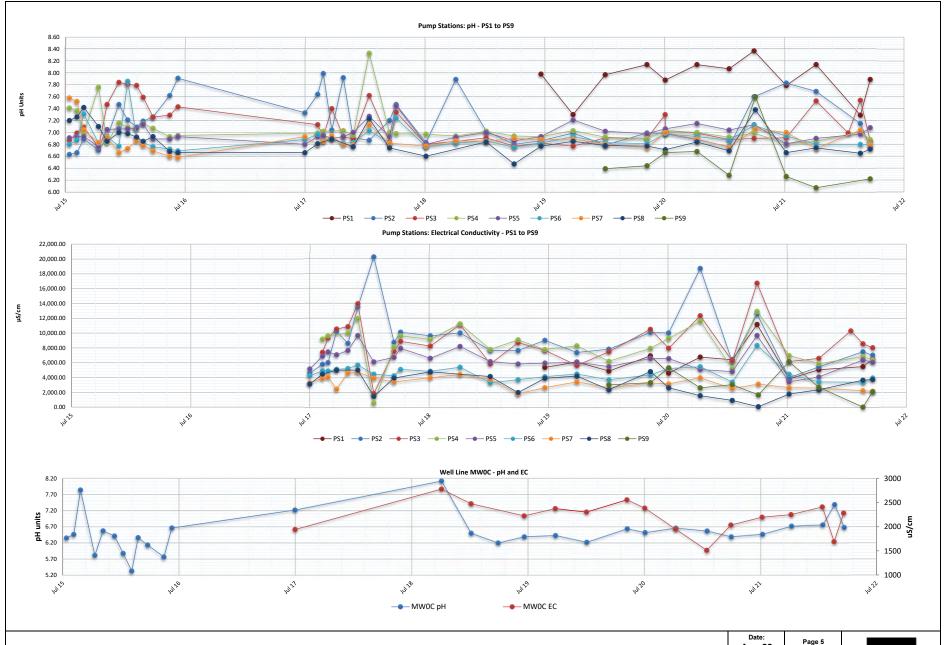


Figure C3-5: Green Island Landfill - Manhole, Shallow Monitoring Wells and Pump Station pH and EC Data



Appendix D

Waste history and distribution



Green Island Landfill

Landfilling History and Targeted Contaminated Land Assessment Report

Dunedin City Council
7 March 2023



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Printed date	14/03/2023 4:24:00 PM
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Author	Cecilia Gately
Project manager	Rose Sullivan
Client name	Dunedin City Council
Project name	GILF Closure Consents
Document title	Landfilling History and Targeted Contaminated Land Assessment Report
Revision version	Rev 01
Project number	12547621

Document status

Status Code	Revision	Author	Reviewer		Approved for issue		
			Name	Signature	Name	Signature	Date
S4	Rev 2	Cecilia Gately	Stephen Douglass	Just .	Stephen Douglass	Jus .	7 March 2023

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1. Introduction

1.1 Background

Existing Otago Regional Council (ORC) resource consents for the operation and management of the Green Island Landfill (the Site), expire in October 2023. It is understood that Dunedin City Council (DCC) is considering options for disposal of waste beyond this date, and eventual closure scenarios for the Landfill.

The Green Island Landfill (the Site) is a municipal landfill facility situated on Taylor Street, to the west of Brighton Road, approximately 10 km southwest of the suburb of Green Island and central Dunedin. A Site Location Plan is provided as Figure 1, overleaf.

GHD Limited (GHD) has been engaged, as part of a wider project to prepare planning documents for application to consent future landfilling and closure, to undertake a review of the history of landfilling at the Site and an assessment of the contamination status of soil recovered during drilling works being undertaken as part of a concurrent GHD geotechnical investigation. The geotechnical investigation was focused on the perimeter of the landfill, with boreholes located mainly around the leachate collection drain / access track area. Further details of borehole locations are provided in Section 4 of this report.

The findings of this report will provide contaminated land status context for both the geotechnical and hydrogeological assessment repots and help inform the design report and consenting process.

1.2 Purpose and Scope of the Environmental Assessment

The purpose of this piece of work is to improve the understanding of the history and extent of landfilling at the Site and assess soil contaminant status in the geotechnical investigation boreholes. This work helps inform the technical reports being prepared as part of the Site's continued operation and closure works and consent application strategy.

The following scope of works was undertaken:

- A review of historical aerial photographs for the Site sourced from Retrolens and DCC.
- A review of historical site investigation reports, Site plans, and compliance reports for the Site.
- Collection of samples from soil cores retrieved during the GHD geotechnical investigation.
- Submission of selected samples to the laboratory for analysis of identified contaminants of concern.
- Comparison of analytical results to adopted guidelines and standards.
- Preparation of a report outlining the history and extent of landfilling at the Site and discussing the findings of the site investigation.

1.3 Assumptions

GHD has made the following assumptions during the preparation of this report:

- Information obtained from third parties and DCC is complete and accurate.
- That the Site will remain in commercial / industrial land use until closure and thereafter will be used for recreational purposes.

Figure 1: Site Location Plan – Green Island Landfill



1.4 Limitations

This report: has been prepared by GHD for Dunedin City Council and may only be used and relied on by Dunedin City Council for the purpose agreed between GHD and Dunedin City Council as set out in Section 1 of this report. GHD otherwise disclaims responsibility to any person other than Dunedin City Council and Council officers, consultants, the hearings panel and submitters associated with the resource consent and notice of requirement process for the Green Island Landfill Closure Project arising in connection with this report.

GHD otherwise disclaims responsibility to any person other than Dunedin City Council arising in connection with this report. GHD also excludes implied warranties and conditions, to the extent legally permissible.

The services undertaken by GHD in connection with preparing this report were limited to those specifically detailed in the report and are subject to the scope limitations set out in the report.

The opinions, conclusions and any recommendations in this report are based on conditions encountered and information reviewed at the date of preparation of the report. GHD has no responsibility or obligation to update this report to account for events or changes occurring subsequent to the date that the report was prepared.

The opinions, conclusions and any recommendations in this report are based on assumptions made by GHD described in this report (refer section(s) 1 through 7 of this report). GHD disclaims liability arising from any of the assumptions being incorrect.

The opinions, conclusions and any recommendations in this report are based on information obtained from, and testing undertaken at or in connection with, specific sample points. Site conditions at other parts of the site may be different from the site conditions found at the specific sample points.

Investigations undertaken in respect of this report are constrained by the particular site conditions, such as the location of buildings, services and vegetation. As a result, not all relevant site features and conditions may have been identified in this report.

Site conditions (including the presence of hazardous substances and/or site contamination) may change after the date of this Report. GHD does not accept responsibility arising from, or in connection with, any change to the site conditions. GHD is also not responsible for updating this report if the site conditions change.

GHD has prepared this report on the basis of information provided by Dunedin City Council and others who provided information to GHD (including Government authorities), which GHD has not independently verified or checked beyond the agreed scope of work. GHD does not accept liability in connection with such unverified information, including errors and omissions in the report which were caused by errors or omissions in that information.

2. Site Description

2.1 Site setting

The Green Island Landfill (the Site) is a municipal landfill facility situated on Taylor Street, to the west of Brighton Road, approximately 10 km southwest of the suburb of Green Island and central Dunedin. The facility is currently managed and operated by Waste Management Ltd., on behalf of DCC who own the landfill.

The Site is located in a reclaimed wetland area within the Kaikorai Estuary, which is part of the larger Kaikorai Catchment; a 55 km² area bounded by the Kaikorai Stream and Abbots Creek and the topography of Chain Hills, to the northwest and north, Kaikorai and Round Hills to the north and northeast and Saddle Hill, to the west.

The Kaikorai Stream historically ran through the Site but was later diverted along the western boundary of the Site to run in a southwest and southerly direction, towards the Kaikorai Lagoon and ultimately the sea. The stream forms the northern and western limits of the landfill before flowing into the Pacific Ocean near Waldronville.

The landfill is approximately 38 hectares in size and is delineated by the legal descriptions in Table 1 below.

Table 1: Legal Descriptions for Green Island Landfill

Legal Descriptions

Pt Secs 44 and 45 Green Island Bush SD

Secs 54-55, 63, 65 Block VII

Section 119 Block VII Dunedin and East Taieri SD

Several activities are currently being undertaken within the boundaries of the landfill including municipal waste disposal, compost production, liquid waste and sludge disposal alongside the operation of a waste transfer station and a recycling centre. A Site layout plan is presented as Figure 2, overleaf.

As can be seen on Figure 2, the working face is currently located towards the centre-south of the landfill, the composting area is located near the eastern boundary and the waste transfer station and recycling station in the north-eastern portion of the landfill. Landfilling is complete over the northern and eastern portion and continues over the remainder of the Landfill.

A network of landfill gas collection wells is installed at the landfill. Additional wells have been installed over time with the development of the landfill. The configuration of the wells changes over time to allow for improved gas collection and collection of gas from areas of new waste. The gas is piped to the Green Island Waste Water Treatment Plant (GIWWTP) for use in the generation of electricity.

A leachate collection trench runs around the perimeter of the northern, western and southern boundaries of the Site. An access track is generally aligned over the top of this trench along much of its length.

Two stormwater sedimentation ponds are located on the landfill, one on the southwestern site boundary (West Pond) and one on the northeast site boundary (East Pond), see Figure 2 overleaf.

2.2 Contaminated land (HAIL) status

The requirements of the Resource Management (National Environmental Standard for Assessing and Managing Contaminants in soil to protect Human Health) Regulations (NESCS, 2011) applies when a selected activity on a piece of land where an activity or industry on the Ministry for the Environment (MfE) hazardous activities and industries list (HAIL) is, has, or is more likely than not to have occurred. Activities such as soil disturbance, subdivision and change in land use are regulated by the NESCS

The Site is known to be part of a landfill and as such is categorised as having HAIL activity G3 (landfill sites) occurring on it.

Figure 2: Site layout Plan



3. Landfill History

3.1 Site history

A landfill has been present on the eastern side of the Kaikorai Estuary since 1954. It is understood that waste filling began on the eastern side of the upper Kaikorai Estuary and by the late 1970s, the Site had become the main landfill for Dunedin.

Unregulated and uncontrolled landfilling occurred at the Landfill until the 1980s, when DCC began to manage waste disposal activities through a national planning approach for the area. The landfill was granted resource consents under the RMA in 1994 and a combined leachate interception trench and collection system was constructed at the Site. This comprised nine (9) pump stations interconnected via a gravel-filled trench with an inbuilt perforated collector drain located around the landfill toe and was retrofitted around the majority of the perimeter of the landfill.

This pump network is set up to maintain a hydraulic gradient towards the trench, minimising the amount of leachate migrating beyond the interceptor trench. The trench is not currently present along the southern boundary of the landfill, between MW0 and PS9, as shown on figure 006116-19-01 Green Island Landfill – Leachate collection and environmental monitoring system, 2004, included in Attachment A. A stormwater interception ditch is located along this boundary which is channelled towards pump station PS1. This gap in the system ws to allow planned placement of waste over this area. In the last few years DCC have decided not to place waste in this area. The interception trench allows for the leachate to be collected and discharged to the Green Island WWTP, via a pipeline, located to the southwest of the landfill.

A network of groundwater / leachate monitoring wells was installed in a series of lines crossing perpendicular to the interception trench, to monitor groundwater / leachate levels across the trench to confirm hydraulic containment of the shallow groundwater. This network consists of both shallow and deep monitoring wells and each line is located approximately halfway between each pump station.

A schematic cross section plan of the landfill and the location of the leachate collection drain and monitoring well arrangement is presented in Figure 3 below and in figure 006116-19-01 included in Attachment A.

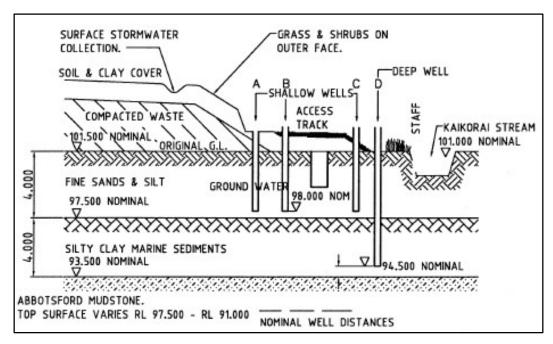


Figure 3: Schematic cross section through an example monitoring well line and the leachate collection trench (Source DCC Landfill Annual Survey Plans – July 2004 Sheet No. G11).

An HDPE liner was historically placed between the leachate interception trench and the Kaikorai Stream to minimise flow from the stream and groundwater migrating eastwards towards the landfill. During this time (HDPE liner placement), a clay bund was also installed around the site boundary to contain both the landfill (deposited waste) and leachate.

3.2 DCC Reports and Plans

A selection of reports provided to GHD by DCC have been reviewed to understand the history of the landfill and its evolution. These are discussed in the following sections.

3.2.1 Beca Steven EIA

In 1992 Beca Steven prepared an Environmental Impact Assessment¹ (EIA) report for the landfill. This report states the following:

Section 4.1.2 The Site

In 1954, an industrial landfill was started on the eastern side of the upper Kaikorai Estuary. At the time, the Kaikorai Stream was still seriously contaminated from industrial dumping, storm water runoff and household waste. There was little thought given to the impact that such an activity might have on the wetland environment in the vicinity, although much of the wetland area was farmed. The same reasons that made the area ideal for industrial development (i.e. close to the city but out of the Otago Harbour catchment), also made it ideal as a refuse site. In the early years of the landfill, as was common practice around the country, little attention was paid to what was dumped or buried on the site.

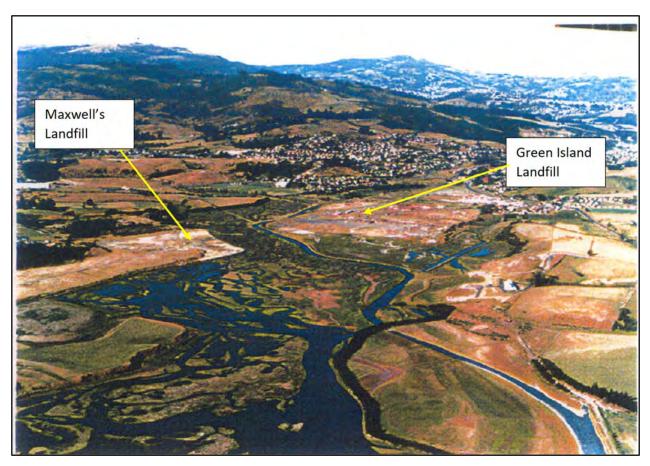
By the late 1970s Green Island had become the main landfill for all of Dunedin, taking in an average of 58,000 tonnes of refuse annually.

Currently the landfill at Green Island and the adjacent private Maxwell's landfill just across the Kaikorai Stream (started in 1968) cover over 30% of the total estuarine landscape of the valley to a maximum depth of 10 metres. Adjacent to the Green Island Landfill the Kaikorai Stream has been channelled behind bunding for the purposes of the drainage and protection of farmland. One consequence of this has been that the bunded areas form in effect oxidation ponds helping to control and treat leachates.

Section 4.2.2 Existing Features of the Site

The Green Island landfill is situated at the head of the Kaikorai Estuary and is separated by a bund constructed of in situ materials from the Kaikorai Stream. The photograph in Figure 4.2 (see below) shows the position of the site within the locality (Maxwell's landfill is to the left and Green Island DCC landfill to the right).

¹ Beca Steven (1992) Environmental Impact Assessment of the Extended Green Island Sanitary Landfill. October 1992.



The site itself is essentially a flat area on top of a raised platform. The flat area presents an untidy appearance due to the presence of windblown rubbish over much of the site and the limited "finishing" of the filled areas which has taken place.

The primary uses of the site are for refuse processing and disposal by land filling.

A prominent use of the site is the stockpiling of car bodies prior to periodic shipment to Auckland for scrap metal recovery.

Section 4.3.4 Refuse Acceptance Policy

Normal domestic and commercial wastes are accepted automatically and whether they contain hazardous or special wastes is left to the disposer to identify, although the landfill operator can inspect any load if there is reason to suspect the material may contain special or hazardous wastes.

Routine special wastes such as dead animals and small volumes of offal are accepted and disposed of by burying at the toe of the landfill.

Section 4.3.6 Surface Water Management

Streams and watercourses approaching the site are diverted around it - one to the south of the landfill which runs outside the southern bund, and one which is channelled down the eastern side of the landfill and discharged to Kaikorai Stream. The watercourse to the east is not well separated from the landfill and contamination of it by leachate occurs over the lower reaches.

Surface water from the area of the completed existing landfill is collected in channels and discharged away from the face. However many of the channels are not well formed or graded, and are affected by landfill settlement. While collecting and preventing some surface water from entering the landfill they are not totally effective. Nevertheless the upper levels of the landfill are relatively dry.

None of the upstream catchments to the south and east are sufficient in size to generate large flows and the opportunity for them to flood the site is virtually non-existent.

To upgrade stormwater control for the existing landfill operation at Green Island, the following mitigating works were recommended in the Existing Green Island Landfill EIA:

- Constructing and operating two sedimentation ponds within the landfill site
- providing a system of surface drains to safely convey stormwater within the site.
- providing for the safe conveyance of stormwater around the site without causing erosion of streambanks or the landfill. This is incorporated in the proposed construction of the outer perimeter face.
- implementing sound sediment control practices.
- conducting a monitoring programme on the discharge from the sedimentation ponds.

Section 4.3.7 Leachate and Groundwater controls

As described in the Existing Green Island Landfill EIA the recommended method of collecting leachate contaminated groundwater is in a gravel filled trench around the landfill perimeter with a perforated collector drain. Pump stations at about 200m spacing will draw down the groundwater table so that an inward hydraulic gradient is formed. Leachate contaminated groundwater from under the landfills will be intercepted by the trenches which will also draw water from the stream side and the discharge will be piped to the Green Island Wastewater Treatment Plant.

However, because the contamination of groundwater has most likely been confined to the immediate vicinity of the landfill little has travelled beyond the bunded area as evidenced by the water testing carried out to date. No groundwater wells are used in the area. The improved groundwater quality after collection of the leachate, will not be a noticeable benefit in itself, but indirectly will improve the Kaikorai Stream by preventing leachate discharge to it.

Summary

The following is understood:

- Development of the landfill was ad hoc for two to three decades before a proper management plan was put in place.
- Municipal, commercial / industrial, and special and hazardous waste were accepted at the landfill.
- Prior to the installation of the leachate collection trench in 1994, there was very little historical control of leachate collection or discharge and as such leachate was discharging into the Kaikorai Stream.
- Prior to the construction of the sedimentation ponds, stormwater management was historically very limited.
- The stream is contaminated, as is much of the catchment, due to historical commercial and industrial activities upstream of the Site.

3.2.2 Trench installation report

According to the Green Island Landfill Leachate Trench Geological Report², the landfill was developed in the upper reaches of Kaikorai Estuary without modern liner and drainage systems beneath the waste and fill to control leachate migration to the underlying sediments and groundwater.

In 1994, a 1,757 m long, gravel filled trench was constructed around most of the perimeter of the landfill to intercept groundwater and leachate flowing from the site. Pumps were installed in the trench to collect leachate / groundwater from the trench and to maintain a hydraulic barrier to manage discharge of leachate to the surrounding environment.

This Report included photographs and stratigraphic logs describing the soil conditions encountered during the trench excavation. An assessment of stratigraphy in this report observed that landfill refuse and/or fill material was the upper most material at 41 of the 76 locations logged and that the maximum thickness of landfill material observed was 2.6 m. The figure with locations of the soil logs was not included in the document, however the locations of the different trench profile logs can be interpreted from their position in relation to individual pump station, manhole and air vent locations.

² Green Island Landfill Leachate Trench Geological Report, Barry J. Douglas Geological Consultants, August 2002

A summary of the locations of where landfill material was encountered, its thickness and composition is provided in Table 2. Please refer to Figure 006116-19-01 in Attachment A for trench profile locations (logs).

Table 2: Landfill material location, thicknesses and composition

Trench profile number	Location	Thickness of landfill (m)	Comments
17	6m north of PS3	0.5	Earthfill intermixed with bricks and shells, black leachate stained refuse
18	33m north of PS3	0.1	Earthfill intermixed with refuse
20	мн3	1.5	Earthfill and mixed refuse
21	29m north of MH3	1.6	Mixed refuse
22	36-42m north of MH3	2	Mixed refuse
23	17-18m south of PS4	0.8	Mixed refuse
25	30.5m north east of PS4	1.4	Mixed refuse
26	38m north east of PS4	1.25	Mixed refuse
27	42m west of AV4	2.1	Mixed refuse
29	0.5 west of AV4	2.6	Mixed refuse
30	AV4	2.6	Mixed refuse
31	4m east of AV4	0.65 (north wall) 2.6 (south wall)	Landfill
32	13m east of AV4	0.5	Brown earthfill intermixed with refuse
34	32m east of AV4		Landfill removed during and after trench construction
42	36m east of PS5	0.4	Intermixed refuse and earthfill
43	49m east of PS5	0.3	Brown earthfill intermixed with refuse
51	МН6	2.0	Landfill
52 (A)	Beneath access road (46m north west of PS7)	~6.0	Mostly earthfill with minor intermixed solid rubble (this material was placed in the 1970's-1990's period)
52 (B)	40m north west of PS7	3.75	Mostly earthfill with minor intermixed solid rubble

53	19m north west of PS7	1.1	Earthfill
55	19m south east of PS7	1.6	Mixed refuse
56	30m south east of PS7	1.6	Mixed refuse
57	49m south east of PS7	0.95	Earthfill and refuse (coke, sawdust, vegetation)
58	50m north of MH7	0.95	earthfill
59	27m north of MH7	1.5	Earthfill and black refuse
60	4-5m north of MH7	1.6	Black refuse
61	МН7	1.2	Black refuse
62	34.7m south of MH7	1.6	Intermixed earthfill and refuse
63	18m north of PS8	1.75	Earthfill and refuse
64	7-8m north of PS8	1.5	Mostly earthfill mixture of sandy and muddy sediments
65	28m south of PS8	2.45	Mostly brown earthfill
66	48m south of PS8	2.4	Mixed refuse
67	60.5m south of PS8	1.9	Landfill
68	МН8а	2.35	Landfill
69	40m west of MH8a	1.95	Intermixed earthfill and refuse
70	64m west of MH8a	2.1	Intermixed earthfill and refuse
71	мн8	1.7	Intermixed earthfill and refuse
72	16m north of PS9	0.95	Earthfill (Landfill)
74	27m north of PS9	1.3	Earthfill (Landfill)
75	37m north of PS9	1.4	Earthfill (Landfill)
76	55m north of PS9 (end of trench)	1.45	Earthfill (Landfill)

Notes:

PS – Pump station

MH – Manhole

AV – Air vent

As can been seen for the above table, refuse material is present on the outside of the leachate collection trench from north of pump station PS3 on the western side of the landfill, to Manhole MH8 on the eastern side of the landfill. The greatest thickness of material was present between Pump station PS4 through to Air vent AV4.

3.2.3 Landfilling contour plan

GHD reviewed a figure depicting the estimated locations of the tip faces at the landfill from 1964 to 2001. This plan has been provided to GHD by DCC and its provenance is unknown, though is likely hand drawn based on aerial photographs review and site history knowledge of previous site managers.

In general, this figure supports other documents and photographs reviewed, in that it shows landfilling operations starting in the southeast portion of the site and advancing to the north and west over time. The historical tip faces appear to be estimates, and the figure does not show landfilling activities outside of the leachate trench. A copy of this figure is included in Attachment A. The historical placement of waste pre installation of the leachate interception trench is discussed further in Section 3.4.

3.2.4 Masterplan 2021

A landfill gas (LFG) masterplan for the landfill was prepared by Tonkin and Taylor in 20213.

The report states that in July 2020, there was approximately 4.8M tonnes of waste in the landfill. The history of landfilling at the landfill is summarised in Table 3. The purpose of including this history was to understand the potential generation of landfill gas from historically placed waste. Although the landfill was operational since 1954, tonnage data was first recorded in 1964.

Table 3: Landfilling history (source – Tonkin and Taylor Landfill gas masterplan report (2021))

Filling period	Average fill depth (m)	Description	Capped	LFG extracted
1954 - 1976	6 m	87600 m ² area to the east of the current filling area.	This area has been capped with hardfill, soils and topsoil. This area is currently utilised as the transfer station and other logistical activities associated with the site.	No
1977 - 1992	9 m	119,300 m ² area in the centre of the site.	This area was capped with intermediate soil cover. Filling over the cover material recommenced in 2002.	Yes, existing wells
1993 - 2001	9 m	72,500 m ² area in the western part of the site.	This area is capped with intermediate soil cover.	No Extraction is proposed as part of future filling in this area.
2002 - 2020	12 m	Filling recommenced in the central part of the site in 2002 and was continuing in late 2020.	No final capping installed yet, intermediate cover in place for completed areas. Final capping currently being designed by others, including 600 mm low permeability soil layer and topsoil layer on top of an intermediate cap layer.	Yes

The report states that the landfill has received a mix of municipal waste, commercial and industrial waste, and construction and demolition waste.

This report does not provide a plan showing the extent of the landfilling.

³ Tonkin and Taylor (2021)Landfill Gas Masterplan, Green Island Landfill. Job number 1008787.5010.v2. Dated May 2021.

3.3 Historical Aerial Photographs

As part of this report a review of historical aerial photographs from the online repository Retrolens was conducted to assess the landfilling sequencing and historical extent of the landfill.

In general, this review shows that landfilling began at the site before 1958. Landfilling then progressed from the southeast portion of the current property to the north until the eastern portion of the property was covered with fill by 1967. From there, filling advanced northwards and to the west over time. The approximate current landfill footprint was filled by 2000. The 2013 aerial photograph generally shows the Site in its current configuration.

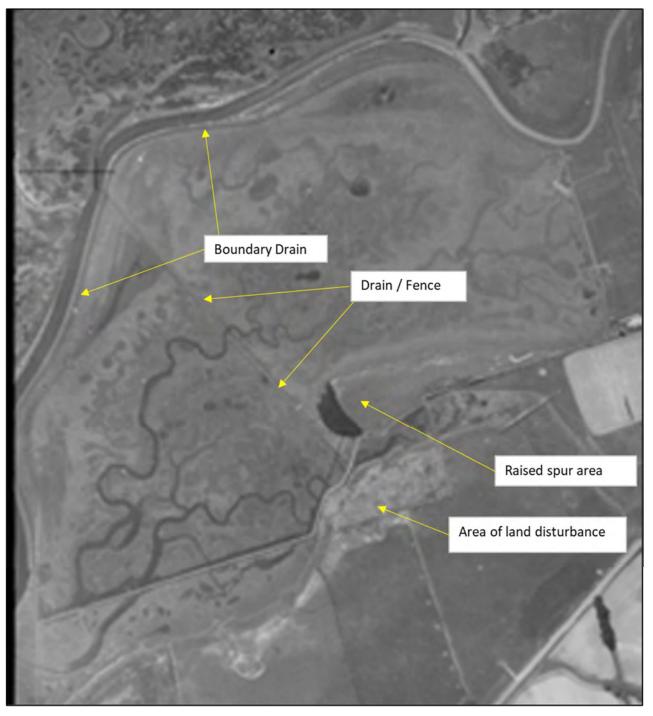
In addition to the aerial photographs obtained from Retrolens, DCC provided GHD with aerial photographs of the Site from 1994 and 1999. These two aerial photographs are of a higher resolution than those obtained from Retrolens. A summary description of the aerial photographs reviewed is provided in the following sections.

3.3.1 1942 Aerial Photograph

No evidence of landfilling is apparent in the 1942 aerial photograph. There is land disturbance evident on the southern boundary, possibly winning material from the hillside for creation of vehicle access to the stream-edge to construct boundary drains.

There appears to be a drain / trench that runs on the inside of the Kaikorai Stream and continues around the boundary of the existing landfill property. There also appears to be another drain or fence which crosses the Site diagonally from northwest to southeast across the wetland.

A raised spur can be noted on the southern site boundary.



(Source - Retrolens)

3.3.2 1947 Aerial Photograph

No evidence of landfilling is apparent in the 1947 aerial photograph.

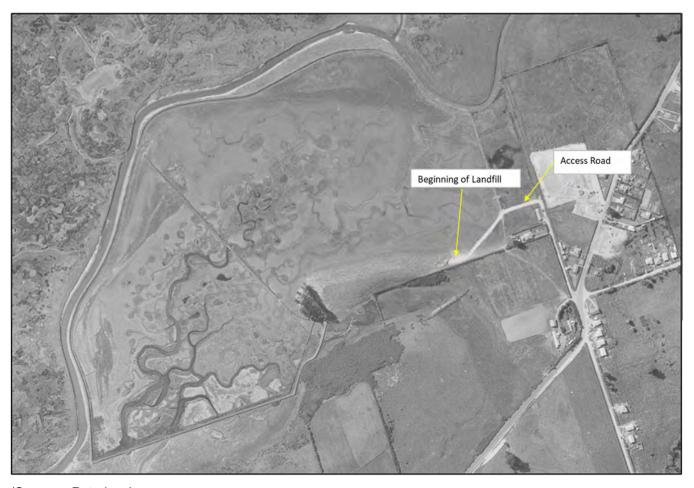
It can be noted that there appears to be a spur of raised ground in the southern portion of the Site with trees at the top. It appears that this portion of the Site increases in height towards the west.



(Source - Retrolens)

3.3.3 1958 Aerial Photograph

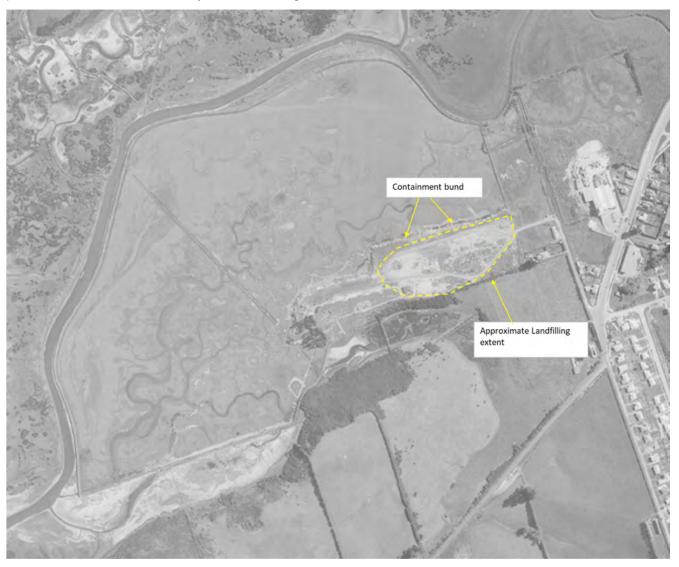
There appears to be an access road into the future landfill at the southeast corner of the current facility. At the end of the access road, it appears that landfilling or initial site development has commenced at the site. It is likely that the area on the southern boundary is being raised / built up to create a road to allow access into the landfill.



(Source - Retrolens)

3.3.4 1962 Aerial Photograph

Landfilling has extended westwards along the southern boundary. A structure, maybe a containment bund, is present on the northern boundary of the landfilling area.



(Source – Retrolens)

3.3.5 1970 Aerial Photograph

Landfilling has continued on the southern boundary of the site and has extended to the north to the edge of the boundary drain. The area of the higher spur of land from the 1940's and 1950's is clearly obvious as a different colour (lighter) along the southern boundary, indicating that this area continues to be used as a source of soil material used during landfilling works.

The southeast corner of the landfill (where filling began) has vegetation growing on it, indicating that filling in that area was complete by 1970.

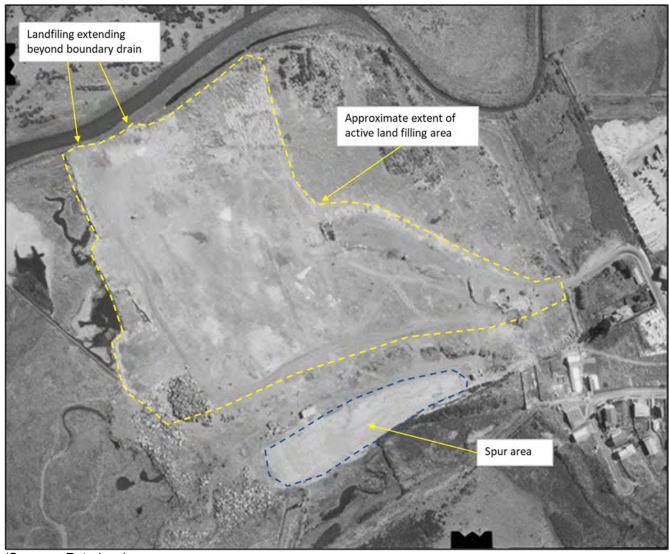


(Source - Retrolens)

3.3.6 1978 Aerial Photograph

Landfilling continues to the north, west and to the northwest. Along a portion of the northern boundary of the landfill, fill has been deposited as far as the inner drain and appearing to infill it. Again, the area of the higher spur of land, visible in the 1940's photographs, appears to have soil disturbance activities being undertaken on it and likely continues to be used as a source area for soil to be used during landfilling activities (cover material etc.).

A portion (southern) of the drain that ran northwest to south east across the Site has been infilled.

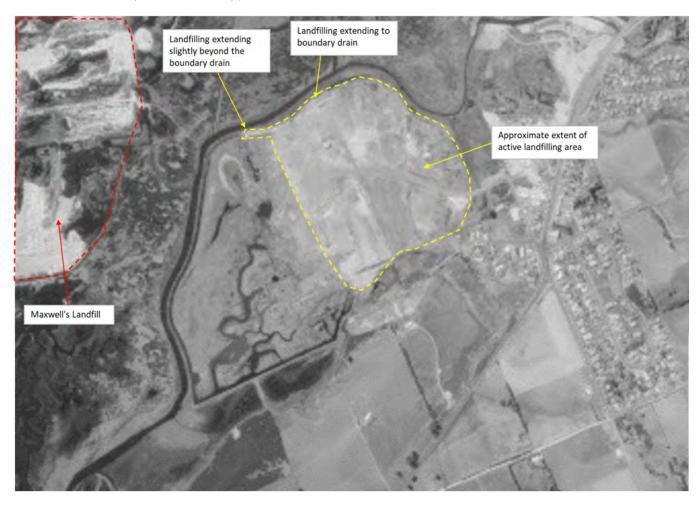


(Source – Retrolens)

3.3.7 1985 Aerial Photograph

Filling appears to have progressed north-eastwards across the footprint of the landfill area, with minimal expansion to the west. Waste appears to be have infilled the boundary drain along the central portion of the northern boundary and likely extend beyond it towards the Kaikorai Stream in isolated areas.

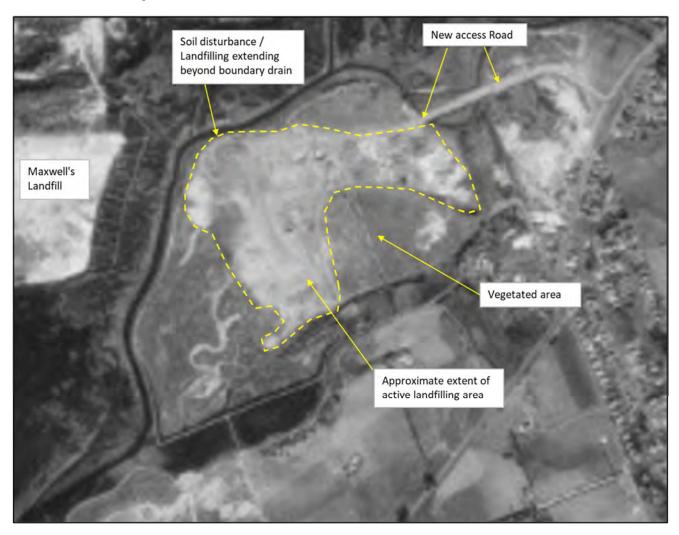
Maxwell's Landfill is present on the opposite side of the Kaikorai Stream.



(Source - Retrolens)

3.3.8 1990 Aerial Photograph

The access road was moved to the northern portion of the Site since 1985. Areas of revegetation are visible on the southeast portion of the site. Fill has expanded to the west especially in the northern half of the landfill. It appears that the boundary drain has been infilled further westwards along the northern boundary with evidence of soil disturbance / landfilling on the northern side of it in one area.



(Source - Retrolens)

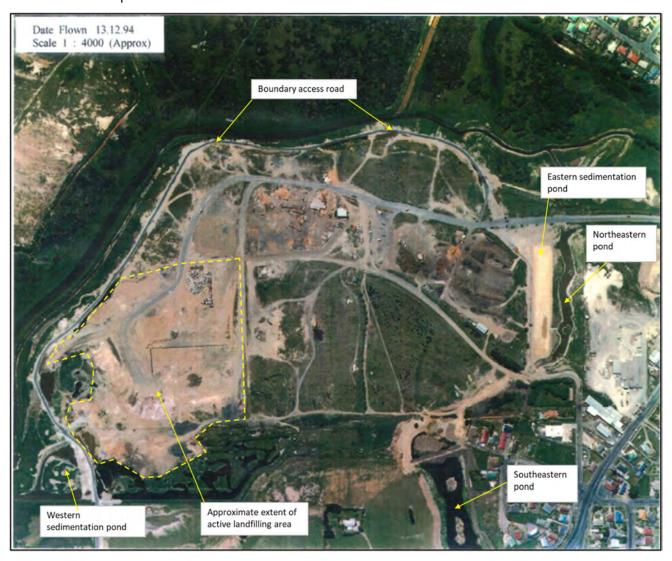
3.3.9 1994 Aerial Photograph

In the 1994 aerial photograph, an access road has been constructed around the northern and western boundary of the Site (along Kaikorai Stream). This access road generally follows the current leachate interceptor trench. Access to the Site is from the east into the northeast corner of the Site, which is the current configuration.

Active filling at the landfill appears to have been concentrated on the western portion of the site and particularly onto the wetlands at the southwest corner. Several support buildings appear to have been built since 1990. Revegetation is ongoing in the southeast quadrant of the site.

It appears that that access track / leachate collection trench has been constructed along the western and northern boundaries. The trench / drain which had run along the southern boundary has been infilled.

It appears that the Eastern and Western sedimentation ponds are under construction and that the north-eastern and south-eastern ponds have been construction.



(Source - DCC)

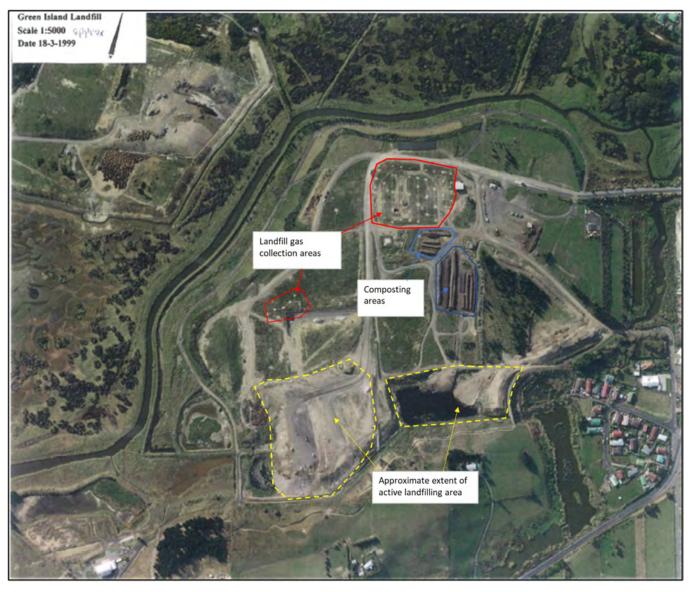
3.3.10 1999 Aerial Photograph

In the 1999 aerial photograph, it appears that the landfill footprint has expanded to include most, if not all, of the current footprint.

The weighbridge has been built in its current location along the main access road to the site.

Active filling is concentrated at the southeast corner and along the central southern boundary of the site. Revegetation is ongoing across much of the site (including the central, eastern and western portions) indicating that active filling is not being undertaken in these areas for some time. Wind-break planting has been established on the northern and western boundary of the landfill.

There appears to be a landfill gas collection network under construction or use in the northern and part of the western portion of the site.



(Source - DCC)

3.3.11 2013 Aerial Photograph

By 2013, the Site is generally in its current configuration. The resource recovery facility / transfer station, wind break planting, the access track (and interceptor trench), and the compost facility are all visible in this aerial photograph. Landfilling appears to be active in the northern portion of the Site. The area of capping completed in 2009 / 2010 is noticeable as vegetated (green) in the southeast corner of the landfill, as is an area of capping underway at that time immediately adjacent to the west.



(Source - Retrolens)

3.4 Landfilling extent

A review of the available historical aerial photographs and leachate collection trench construction report indicates that waste is present on the outside of the leachate collection trench in places, with waste up to 6 m thick on the north eastern boundary in a short section (approximate 20-40 m length). Generally, the thickness of the waste is less than 2.6 m in other areas, and averaging 1.63 m depth over the other 40 locations mapped from the log.

The aerial photographs (1978 and 1982) indicate that landfilling activities likely extended very close to the edge of the Kaikorai Stream banks along the northern boundary with some waste placement also likely to having occurred in these areas. The likely extent of the landfilling is shown on Figure 4, overleaf.



4. Site works

Field work was carried out in compliance with the project specific Job Safety and Environmental Analysis (JSEA) plan. All site investigation staff were inducted to the Site by Waste Management Ltd, the operators of the landfill. Environmental sampling was undertaken between 1 November and 9 November 2022, comprising the following:

- A GHD environmental scientist was on site to assess and sample soil from eight boreholes installed as part of the geotechnical investigation (BH-100 through BH-104, BH-107, BH-108 and BH-111).
- A GHD environmental scientist was on site for the installation of monitoring wells in boreholes BH100, BH-101, and BH-104.
- Soil samples from boreholes BH-102, BH-103, BH-107, BH-108 and BH-111 were collected from core boxes by a GHD environmental scientist. The remainder of the samples were collected during the drilling of BH-100, BH-101 and BH-104.
- Collection of 26 original and 2 duplicate soil samples from the above-mentioned boreholes. Samples were
 collected from multiple depths in each borehole. The sample depths were chosen based on an assessment of
 the soil by a GHD environmental scientist.

A site investigation points location plan is presented as Figure 5, overleaf.

As the Site is a known piece of contaminated land, a land use consent was required to drill and install the boreholes. Resource consent RM21.467 (to drill 17 investigation sites on contaminated land for the purpose of geotechnical investigation) was granted by ORC to DCC to undertake these works on 10th October 2021.

Prior to breaking ground, subsurface utilities clearance was undertaken by a specialist contractor (Fulton Hogan), using both a Cable Avoidance Tool (CAT) and Ground Penetrating Radar (GRP) at all borehole locations. A Ground Penetration Permit to Work was issued by an authorised GHD permitter.

Logging of the material encountered during drilling was undertaken by a GHD geotechnical engineer in accordance with NZ Geotechnical Society (2005) field description of soil and rock. Borehole logs are included in Attachment B.



4.1 Soil sampling methodology

The geotechnical investigation was mainly focussed on the peripheries of the landfill to better understand liquefaction potential of the material present in these areas. Consequently, the environmental investigation was limited to the assessment of the material collected during the drilling works in these areas.

Environmental soil samples were collected from a total of eight boreholes at multiple depths at each borehole to target different depths and lithologies to gain an understanding of the material present. GHD collected two duplicate samples for quality control and assurance purposes.

Soil samples were collected in accordance with standard GHD procedures. Samples were placed directly into laboratory supplied containers and then placed in an iced chilly bin and couriered to ALS - Analytica Laboratories under standard GHD Chain of Custody (CoC) procedures. Due to capacity issues at Analytica Laboratories, a portion of the polycyclic aromatic hydrocarbon and semivolatile organic compounds analysis was subcontracted to Eurofins Environmental Testing NZ Limited (Eurofins).

A total of 26 primary soil samples were selected for analysis. Based on the historical land use and observations made during sampling, selected soil samples were analysed for the identified contaminants of concern as follows:

- Heavy metal suite (arsenic, beryllium, boron, cadmium, chromium, copper, lead, mercury, nickel and zinc) x26 samples
- Polycyclic Aromatic Hydrocarbons (PAHs) x23 samples
- Semi Volatile Organic Compounds (SVOCs) x13 samples
- Total Petroleum Hydrocarbons (TPH) x1 sample
- Asbestos in soil (presence/absence) x18 samples
- Asbestos in soil (semi-quantitative) x1 sample
- Ammonia x8 samples
- Two duplicate samples, one of which was analysed for heavy metals and PAHs and the other was analysed for heavy metals only.

4.2 Field Observations

Solid waste and/or fill material was encountered in all eight of the boreholes that were sampled as part of this environmental assessment.

Boreholes BH-100 to BH-104, which were installed on the perimeter of the known landfill area, contained fill and/or solid waste to depths up to 3.95 metres below ground level (m bgl). Boreholes BH-107, BH-108 and BH-111 were installed on the landfill and contained fill to at least 7.5 m bgl, 11.2 m bgl and 4.6 m bgl respectively.

A summary of the borehole locations and the thickness of fill is provided in Table 4 below. Further details can be found on the borehole logs in Attachment B.

Table 4: Borehole locations and fill thicknesses summary

Borehole Number		Thickness of Fill / Waste (m)	Location
BH-100	12.95	2.75	Inside / within the trench
BH-101	12.95	1.3	Outside the trench
BH-102	14.95	1.6	Inside the trench (Bund slope)
BH-103	13.15	1.2 – 1.5 (Core loss)	Outside the trench
BH-104	9.95	3.95 – 4.5 (Core loss)	Inside the trench

BH-107	20.0	7.5	Northern portion of landfill (Inside trench)
BH-108	20.85	11.2 – 13.5 (Core loss)	Southern portion of the Landfill
BH-111	19.95	4.6 – 5.5 (Core loss)	Inside the trench – former tyre storage area

A Photo-ionisation detector (PID) was used to screen soil samples for volatile contaminants at all locations, with the exception of borehole BH-100. PID measurements are included in Table 6. It should be noted that some of the PID readings were collected during drilling works while the others were measured during the collection of soil samples from core boxes. For the samples that were collected from the core boxes, the measured PID values are likely lower than what would have been measured if the samples were collected during the drilling works, as the loss of volatiles would likely have occurred over the storage interval.

Table 5: PID measurements

Borehole Name	Date	Depth (m bgl)	PID reading (ppm VOC in headspace)	Measured in field / Core box	
	1 Nov. 2022	0.5	0.0		
DU 404	1 Nov. 2022	1.5	0.4	lo field	
BH-101	1 Nov. 2022	2.5	0.1	In field	
	1 Nov. 2022	3.5	0.2		
	1 Nov. 2022	0.6	0.0		
BH102	1 Nov. 2022	1.5	1.9	Cara hav	
	1 Nov. 2022	2.5	0.8	Core box	
	1 Nov. 2022	3.5	0.0		
BH-103	1 Nov. 2022	0.2	0.0		
	1 Nov. 2022	1.2	0.1	Core box	
	1 Nov. 2022	2.5	0.0		
	9 Nov. 2022	0.5	0.9		
BH-104	9 Nov. 2022	1.0	17.4	In field	
	9 Nov. 2022	2.0	1.7		
	1 Nov. 2022	0.5	0.0		
BH-107	1 Nov. 2022	1.5	0.0	Core box	
	1 Nov. 2022	2.5	0.0		
	9 Nov. 2022	0.5	0.9		
BH-108	9 Nov. 2022	1.9	2.0	Core box	
	9 Nov. 2022	3.3	27.9	(core collected earlier that day)	
DU 444	7 Nov. 2022	0.5	1.4		
	7 Nov. 2022	1.5	7.2		
BH-111	7 Nov. 2022	2.0	5.0	Core box	
	7 Nov. 2022	3.0	3.9		
	7 Nov. 2022	3.2	2.1		

Note: ppm VOC – parts per million Volatile Organic Compound

5. Regulatory Context

5.1 Applicable Soil Contaminant Standards

5.1.1 The National Environmental Standard for Assessing and Managing Contaminants in Soil to Protect Human Health

The User's Guide: National Environmental Standard for Assessing and Managing Contaminants in Soil to Protect Human Health (MfE, 2012) ⁴ details Soil Contaminant Standards (SCSs) for seven inorganic substances and five organic compounds (or groups of compounds). SCSs are available for these substances and compounds when present in land used for five land use scenarios. The contaminants analysed at this site for which SCSs are available are arsenic, cadmium, chromium, copper, lead, mercury, DDT (Dichlorodiphenyltrichloroethane) and benzo(a)pyrene equivalent (BaP). The NESCS applies to a "piece of land" on which a HAIL activity has occurred or is currently occurring.

The land use category selected for the purposes of this investigation was Commercial / Industrial and is described in the NESCS User Guide as "Commercial / Industrial site with varying degrees of exposed soil. Exposure of outdoor workers to near-surface soil during routine maintenance and gardening activities with occasional excavation as part of maintaining subsurface utilities. Also, conservatively applicable to outdoor workers on a largely unpaved site".

Because the landfill area may, at some point in the future, be used for recreational purposes (eg.as a reserve or park), a Recreational land use scenario was also considered. The NESCS User Guide describes Recreational land use as "Public and private green areas and reserves used for active sports and recreation. This scenario is intended to cover playing fields and suburban reserves where children play frequently. It can also reasonably cover secondary school playing fields but not primary school playing fields."

These land-use exposure scenarios have been adopted for screening purposes to include potential receptors including site workers during any future construction works, current and future users of the Site and ongoing maintenance / excavation workers at the site.

The intention of the NESCS is the protection of human health from contaminated land and the appropriate assessment of the risk to human health prior to the undertaking of the regulated activities (e.g. Soil disturbance or land use change).

If the investigation demonstrates that the contaminants tested are at, or below, screening criteria concentrations, the regulations of the NESCS will not apply should any of the regulated activities be undertaken over the Site.

NESCS SCS criteria adopted for the Site are presented in Table 1 through to Table 3 in Attachment C.

5.1.2 Health and Safety at Work (Asbestos) Regulations 2016

The management and/or removal of asbestos in soils is regulated under the Health and Safety at Work (Asbestos) Regulations 2016⁵ (Asbestos Regulations). However, the Asbestos Regulations do not provide guidance regarding the definitions of what constitutes an asbestos contaminated site, in particular, with regard to soil. Rather, the Regulations simply states that the Asbestos Regulations apply where a competent person advises that the disturbance and/or removal of soil is likely to lead to airborne contamination at a level that exceeds trace concentrations.

⁴ Ministry for the Environment, 2012. Users Guide: National Environmental Standard for Assessing and Managing Contaminants in Soil to Protect Human Health. Wellington: Ministry for the Environment

⁵ Health and Safety at Work (Asbestos) Regulations 2016 (15 February 2016) made under sections 24(1) (m), 211, and 218 of the Health and Safety at Work Act 2015

The New Zealand Guidelines for Assessing and Managing Asbestos in Soil (BRANZ Guidelines 2017)

The BRANZ Guidelines 2017⁶ provide a methodology to ensure that management of asbestos in soil meets regulatory requirements and an acceptable level of managed risk. This methodology is consistent with the MfE CLMG's for New Zealand and the NESCS for the assessment of asbestos in soil.

If asbestos is detected in soils using a laboratory presence/absence test, the BRANZ Guidelines 2017 then provides an additional guidance for further soil sampling and criteria for the definition of whether the removal of such material is considered licenced asbestos removal (Class A or B), asbestos-related-works or unlicensed works under the Asbestos Regulations.

The adopted asbestos criteria are presented in Table 6 in Attachment C.

5.2 Other applicable Human Health Standards

For contaminants of concern that are not listed as priority contaminants, the NESCS references the Contaminated Land Management Guidelines No. 2: Hierarchy and Application in New Zealand of Environmental Guideline Values (CLMG No.2) to provide guidance.

In the absence of New Zealand risk based human health criteria for certain contaminants of concern, such as nickel and zinc, the Australian National Environment Protection Measure 2013⁷ (NEPM) Recreational (C) and Commercial/Industrial (D) guideline values have been adopted for this investigation.

For TPH, naphthalene and pyrene, Tier 1 screening criteria from the MfE Guidelines for Assessing and Managing Petroleum Hydrocarbon Contaminated Sites in New Zealand⁸ have been selected. Tier 1 soil acceptance criteria for Commercial / Industrial land use and Residential land use, All Pathways, Sand 1 - 4 m depth have been adopted. As there are no recreational land use guideline values available, the residential values have been adopted as a conservative screen for this pathway.

The adopted criteria are presented in Tables 1 through to 4 in Attachment C.

5.3 Background Soil Concentrations

Background soil concentrations of heavy metals for the Site were obtained from the Landcare Research (2015) report⁹, Tables 13 through Table 15. The area of the landfill is classified as fill and as such does not have background heavy metal concentrations derived. As material has been received at the Site from all over Dunedin, the highest background value from each of the dominant adjacent geological units for each metal was adopted for comparison purposes.

There are currently no published background soil concentrations for PAHs in Dunedin. As such, PAH values established for Christchurch urban soils¹⁰ were adopted for comparative purposes only.

Background concentrations for OCP pesticides were taken from the following two documents:

- Landcare Research (2015). Background soil concentrations of selected trace elements and organic contaminants in New Zealand. Table 21.
- Ministry for the Environment (1998). Ambient concentrations of selected organochlorines in soils. Table F5.1 mean values.

The adopted criteria are presented in Table 1, Table 2 and Table 3 in Attahment C.

⁶ New Zealand Guidelines for Assessing and Managing Asbestos in Soil - Building Research Association of New Zealand (BRANZ), November 2017

⁷ National Environmental Protection Council (NEPC) (2013). National Environmental Protection Measure (Assessment of Site Contamination) as amended in 2013 Schedule B1, Health Investigation Levels (HIL) for soil contaminants.

⁸ Ministry for the Environment (1999, revised 2011). Guidelines for Assessing and Managing Petroleum Hydrocarbon Contaminated Sites in New Zealand. Module 4- Tier 1 Soil screening Criteria.

⁹ Landcare Research (2015). Background soil concentrations of selected trace elements and organic contaminants in New Zealand

¹⁰ Environment Canterbury, 2007, Background concentrations of polycyclic aromatic hydrocarbons in Christchurch urban soils. Report R07/19

5.4 Otago Regional Council Regional Plans

The operative Regional Plan for Otago (Waste) was enacted in April 1997. Section 5 of this Plan details the policies, methods and rules regarding Contaminated Sites. A contaminated site is defined as a site at which hazardous substances occur at concentrations above background levels and where assessment indicates it poses, or is likely to pose, an immediate or long term hazard to human health and/or the environment.

Hazardous substances are defined as follows:

Hazardous substances are substances which impair human, plant or animal health, or which may adversely affect the health or safety of any person or the environment, whether or not they are contained in or form part of any other substance or thing. These include pesticides, petrol, oil, cleaners and paint.

Rule 5.6.1 of the Plan states the following:

Rule 5.6.1 Hazardous wastes at contaminated sites (discretionary activity)

- The disturbance of land; or
- 2. The discharge of hazardous waste into water; or
- 3. The discharge of hazardous waste onto or into land in circumstances that may result in that hazardous waste (or any other hazardous waste emanating as a result of natural processes from that hazardous waste) entering water; or
- 4. The deposit of any hazardous waste, in, on or under land; or
- 5. The discharge of hazardous waste into air at or from a contaminated site;

is a discretionary activity.

The operative Regional Plan for Otago (Water for Otago) was enacted in February 1998 and updated in June 2021. Section 7 of the Regional Plan, details policies for the discharge of stormwater and the Rules associated with these policies are contained in Section 12 (water takes, use and management). However, many of the Rules related to the discharge of stormwater within this Section have been repealed.

Rule 12.B.3 (Discharge of hazardous substances, hazardous wastes, specified contaminants, and stormwater, and discharges from industrial or trade premises and consented dams) states that the discharge of stormwater to water, or onto or into land in circumstances where it may enter water, is a restricted discretionary activity. The restrictions include the potential for soil contamination.

Should contaminants be found present in the soil at the Site at concentrations that there is a hazard posed to human health and/or the environment, then a consent may be required under the Regional Council's rules for certain activities to be undertaken. Further discussion is provided in Section 7.3 on consenting requirements.

6. Results

6.1 Soil analytical results

Analytical results compared against the adopted standards and guidelines are presented in Tables 1 through Table 6 in Attachment C. The laboratory reports are included in Attachment D.

In summary:

- Heavy metals
 - Metals including arsenic, cadmium, chromium, lead, nickel and zinc were detected at concentrations above the adopted Dunedin background soil concentrations in 13 of the 28 samples analysed.
 - No heavy metals were reported at concentrations above the adopted background values in samples collected from BH-100 and BH-107.
 - Above background concentrations were limited to the upper 2.5 m of material in boreholes BH-101, BH-102 and BH-103. However, they extended to at least 3.3 bgl in boreholes BH-108 and BH-111.
 - No metals concentrations were detected above the relevant human health screening criteria.
 - Refer to Table 1 for tabulated analytical metals results.

PAHs

- Various PAHs were detected above adopted background soil concentrations in eight samples (not including the field duplicate sample).
- No PAHs were reported at concentrations above the adopted background values in samples collected from BH-100 and BH-101, BH-102 and BH-103.
- Above background concentrations of PAHs were distributed through the whole soil profile (surface to 3.3 m bgl).
- PAH, as BaP toxicity equivalence quotient (TEQ), was greater than the NESCS SCS for Commercial/Industrial and Recreational use in one sample (12547621-BH-104(0.5)) collected at borehole BH-104 in fill material.
- Refer to **Table 2** for the tabulated PAH results.

SVOCs

- Various SVOC were detected above background soil concentrations in five samples, two collected from BH-103 (0.2 and 1.2 m bgl), two collected from BH-107 (1.5 and 2.5 m bgl) and the other from BH-111 (1.5 m bgl).
- The pesticide 4,4'-DDD was reported present at a low concentration in boreholes BH-103, BH-107 and BH111 in the samples described above. The pesticides 4,4'-DDE and 4,4'-DDT were also reported present in the samples collected and analysed from BH-103.
- No SVOCs were detected at concentrations above any considered human health or environmental screening criteria.
- Refer to Table 3 for a summary of SVOC results. Only those SVOCs which were detected at
 concentrations above the LOR are presented in the table. The full set of results can be found in the
 laboratory reports.
- TPH was not detected above any of the considered human health or environmental screening standards.
 Refer to **Table 4** for tabulated TPH results.
- Ammonia was detected above the laboratory detection limit in four of eight samples. The highest
 concentrations were found in the samples collected from soils likely to be rich in organic material (on the basis
 of their lithological description). Refer to **Table 5** for a summary of ammonia results. There are no available
 guideline criteria available to compare these results against.

Asbestos

- Asbestos fines were reported present in one sample collected from BH-107 at a depth of 0.5 m bgl (12547621-BH-107(0.5)), and the sample was further scheduled for semi-quantitative analysis. Asbestos was not detected in the sample examined for semi-quantitative analysis.
- Asbestos was not detected in any other soil samples collected as part of this investigation.
- Refer to **Table 6** for a summary of asbestos results.

6.2 Quality Assurance and Quality Control

GHD quality assurance/quality control (QA/QC) procedures to assess data quality were maintained throughout the project.

The QA/QC programme undertaken as part of the assessment by GHD included the following:

- All fieldwork was undertaken by suitably qualified and trained staff. The work managed by a suitably qualified and experienced practitioner (SQEP) and the report was reviewed by a SQEP, as required by the NESCS.
- Collection of two duplicate soil samples for analysis of heavy metals and PAHs.
- Soil samples collected throughout works were dispatched to Analytica Laboratories in Hamilton on the day of collection under standard chain of custody procedures and the analysis was undertaken within the sample holding times.
- Analytica Laboratories is an internationally recognised laboratory endorsed by International Accreditation New Zealand (IANZ) which represents New Zealand in the International Laboratory Accreditation Cooperation (ILAC). Through the ILAC mutual recognition arrangement (ILAC-MRA) this accreditation is internationally recognised. The tests were performed in accordance with the terms of accreditation.

6.2.1 Analytical results

GHD collected two field duplicate samples for quality assurance/quality control (QA/QC) one of which was analysed for both heavy metals and PAHs and the other for heavy metals only.

A quantitative measure of the accuracy of the analytical results received from the laboratory was conducted using calculated relative percentage difference (RPD) values. The RPD values were calculated using the following equation:

RPD (%)
$$= \frac{\langle Co - Cs \rangle}{\langle \frac{Co + Cs}{2} \rangle} \times 100$$

Where Co = concentration obtained from the original sample.

Cs = concentration obtained from the duplicate sample.

The usual acceptance criteria for relative percentage difference (RPD) are between 30 and 50% in soils (CLMG, No.5 section 3.9.1 (2011)). However, larger RPDs are allowed for different analytes (up to 100%). A large percentage differential can occur particularly in soils due to the following:

- A small analytical differential between two samples based on the low levels of detection from the primary and duplicate soil sample.
- Soil samples collected from a non-homogenous (heterogeneous) soil profile.

RPD values for the duplicate samples (field) analysed for heavy metals for this assessment ranged from 0% to 61.6%. However, the majority of the RPDs, 17 out of 20, were less than 50%.

RPD values for the duplicate samples (field) analysed for PAHs for this assessment ranged from 127.5% to 187.9%.

The high RPD values for the PAH analysis are likely due to the heterogenous nature of the material at that borehole location (BH-108).

The duplicate quality assurance and quality control (QA/QC) results are provided in Attachment C, Table 7 and the laboratory analytical reports are provided in Attachment D.

The results of the whole QA/QC program are considered to provide an acceptable degree of confidence in the sampling and analytical program.

7. Conclusions

7.1 Landfilling Extent

A review of the available historical aerial photographs and leachate collection trench construction report indicates that waste is present on the outside of the leachate collection trench in places, with waste up to 6 m thick on the north eastern boundary in a short section (approximate 20-40 m length). Generally, the thickness of the waste is less than 2.6 m in other areas, and averaging 1.63 m depth over the other 40 locations mapped from the log.

The aerial photographs (1978 and 1982) indicate that landfilling activities likely extended very close to the edge of the Kaikorai Stream banks along the northern boundary with some waste placement also likely to having occurred in these areas. The trench construction report indicates that waste / landfill material is present on the outside of the trench at the following locations:

- 6m north of PS3 through to 32m east of AV4
- 36m east of PS5 to 49m east of PS5
- MH6 through to 18m north of PS8

Waste is likely to extend either side of these locations. In addition, earthfill is also present from south of PS8 through to the end of the trench (55 m north of PS9).

The likely extent of the landfilling is shown on Figure 4.

7.2 Soil contaminant status

An environmental site investigation was undertaken in conjunction with a geotechnical investigation in October and November 2022. Soil samples were collected from eight of the boreholes drilled during the investigation, six of which were located around the periphery of the landfill and two in the northern portion of the landfill (area of oldest landfilling). These samples were laboratory analysed for the identified contaminants of concern. The material in the boreholes was logged by a geotechnical engineer and observations of soil type and indications of contamination were also made by an environmental scientist.

Municipal solid waste (MSW) and fill material was found present in all the boreholes assessed.

The analytical results indicated that contaminant concentrations at above background values were present through the soil profile from surface to a depth of 3.3 m bgl. One sample, at BH-104(0.5 m bgl), had a reported a BaP equivalent concentration which exceeded the NESCS SCS value for both Commercial / Industrial and Recreational land use.

Asbestos fines were reported present in only one of the 19 samples analysed. Low concentrations of TPH were reported present in the one sample analysed for this contaminant. Elevated concentrations of ammonia were reported in samples collected from borehole horizons described as peat, silty sand, fill / organic silt and fill / wood in sand. These materials are likely to have elevated organic content.

Pesticides (4,4'-DDE, 4,4'-DDD and 4,4'-DDT), at low concentrations, were reported present in five of the thirteen samples analysed for these contaminants.

7.3 Regulatory requirements

7.3.1 NESCS

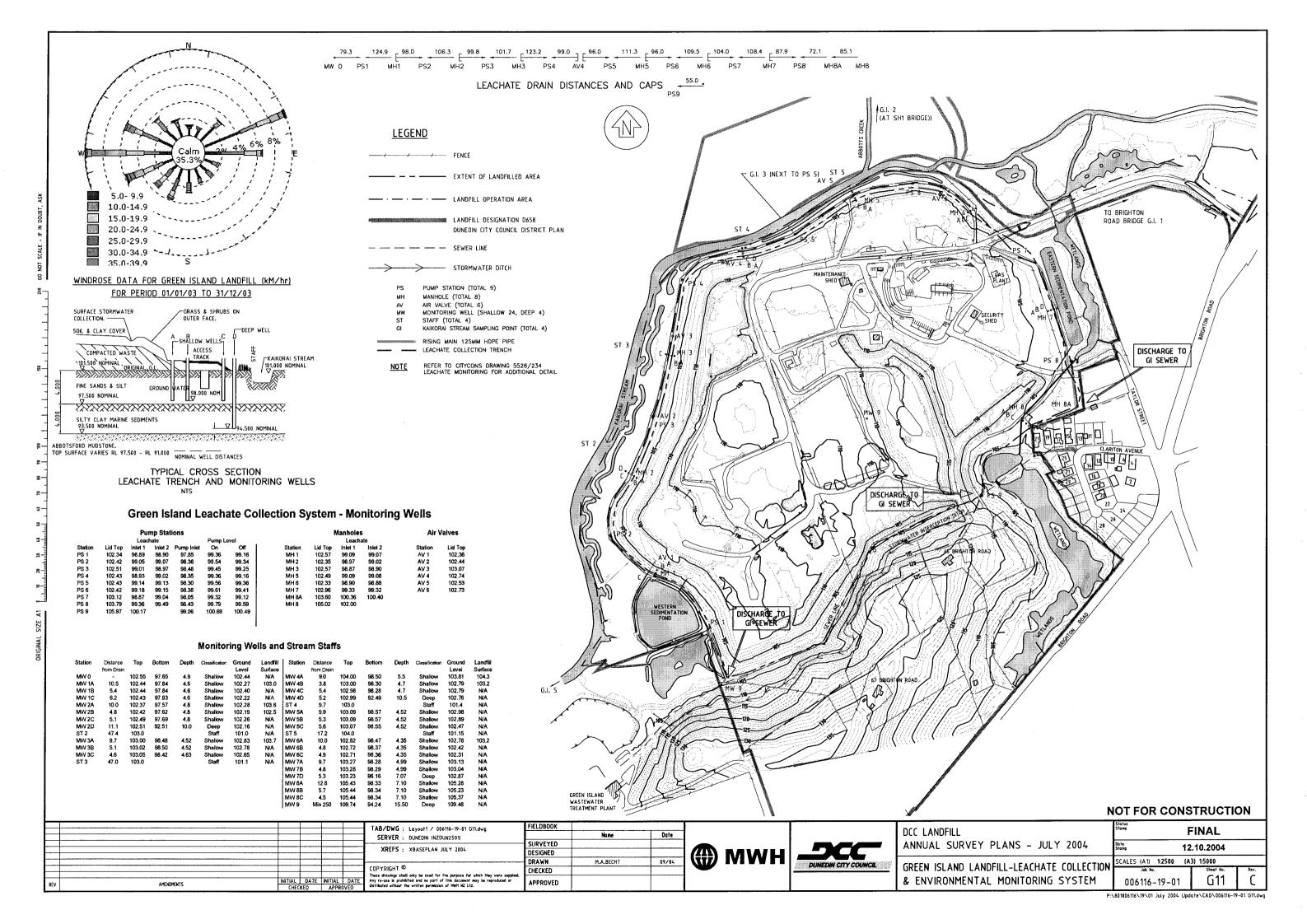
This investigation has confirmed that HAIL activity G3 (landfill sites) has occurred on the Site and as such the requirements of the NESCS apply to any future redevelopment works (e.g. soil disturbance and land use change).

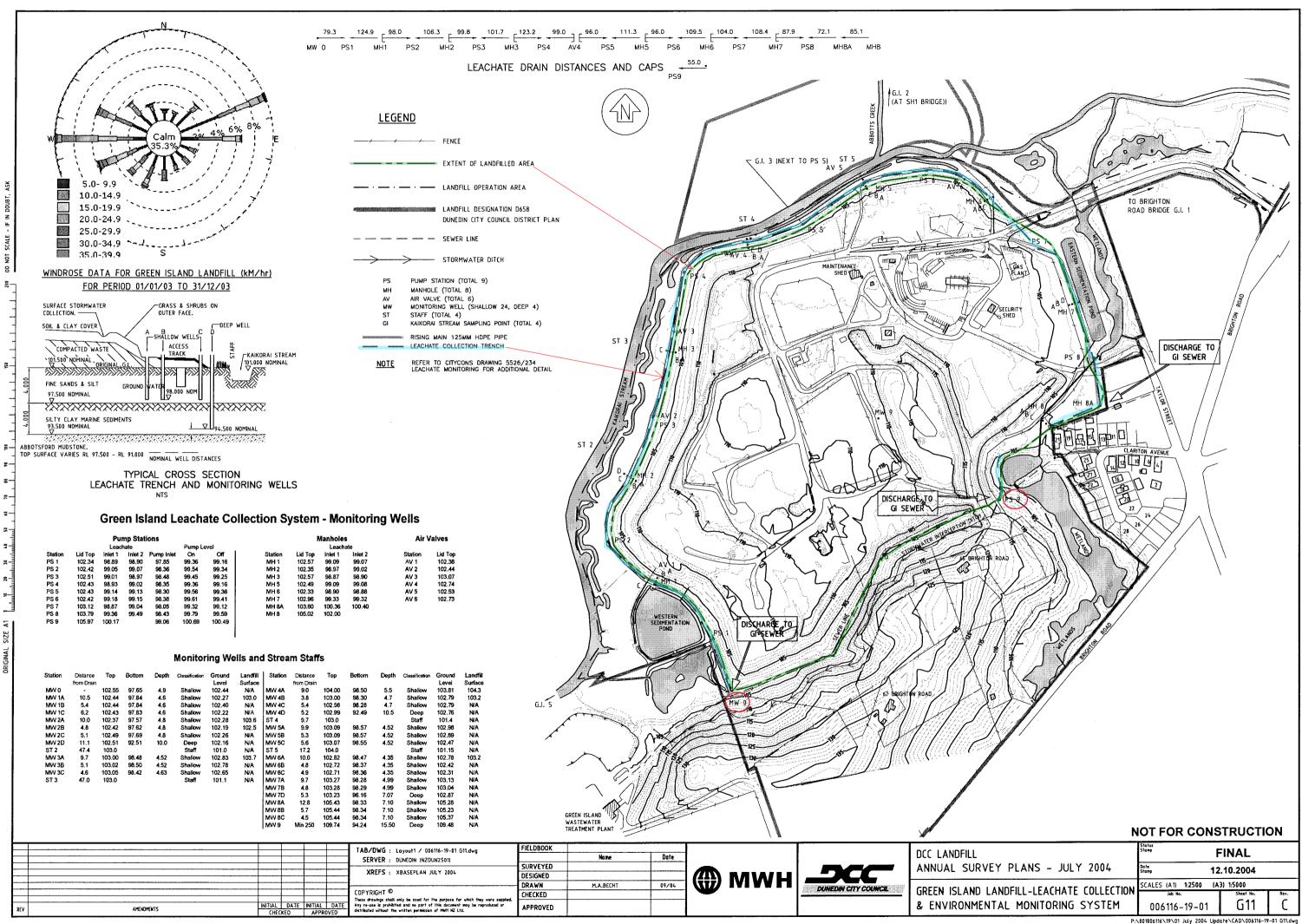
7.3.2 Regional and Local Council Requirements

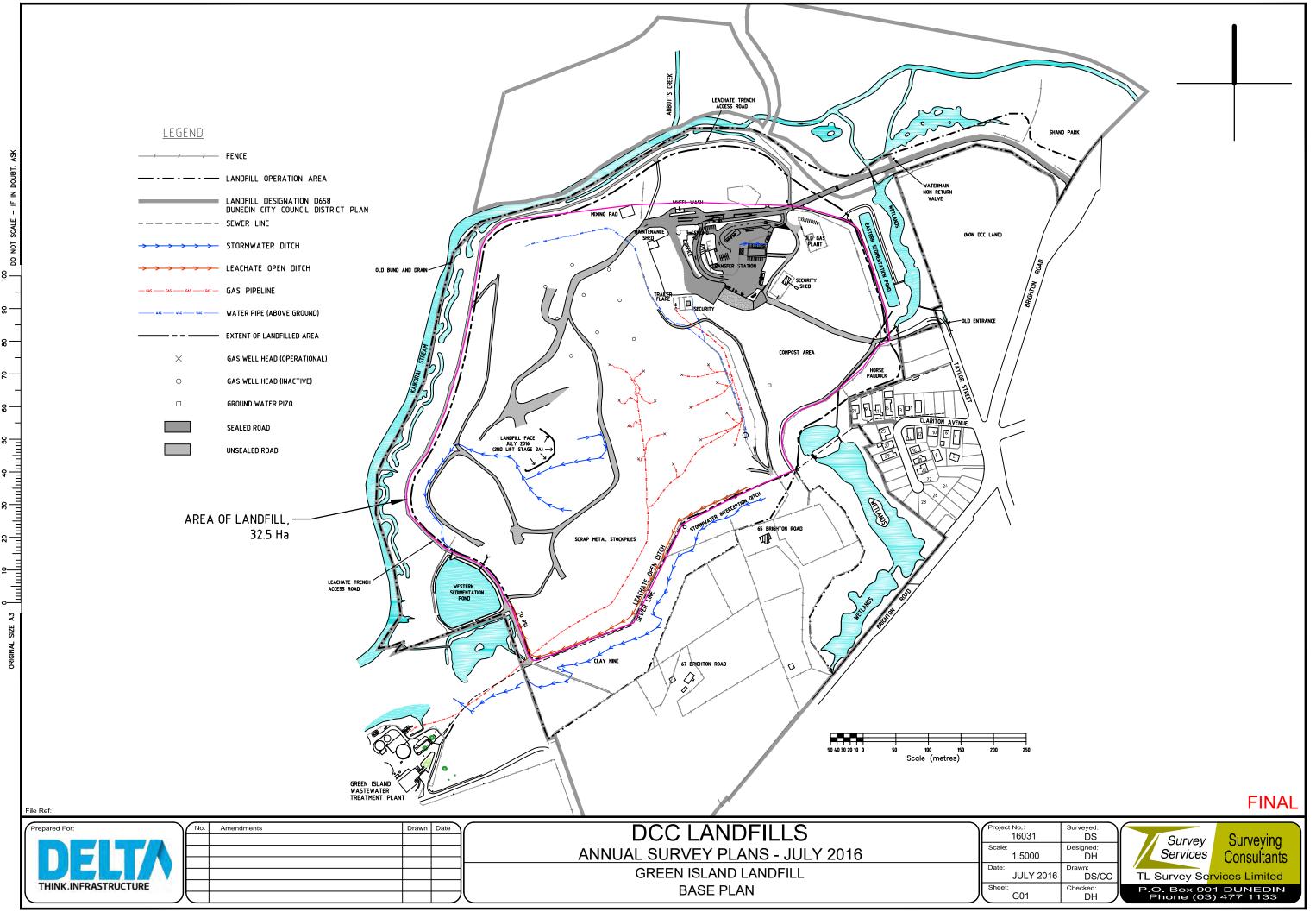
It is understood that a review of Regional and Local Council planning requirements is to be undertaken by Boffa Miskell. As such, further discussion of any requirements has not been undertaken in this report.

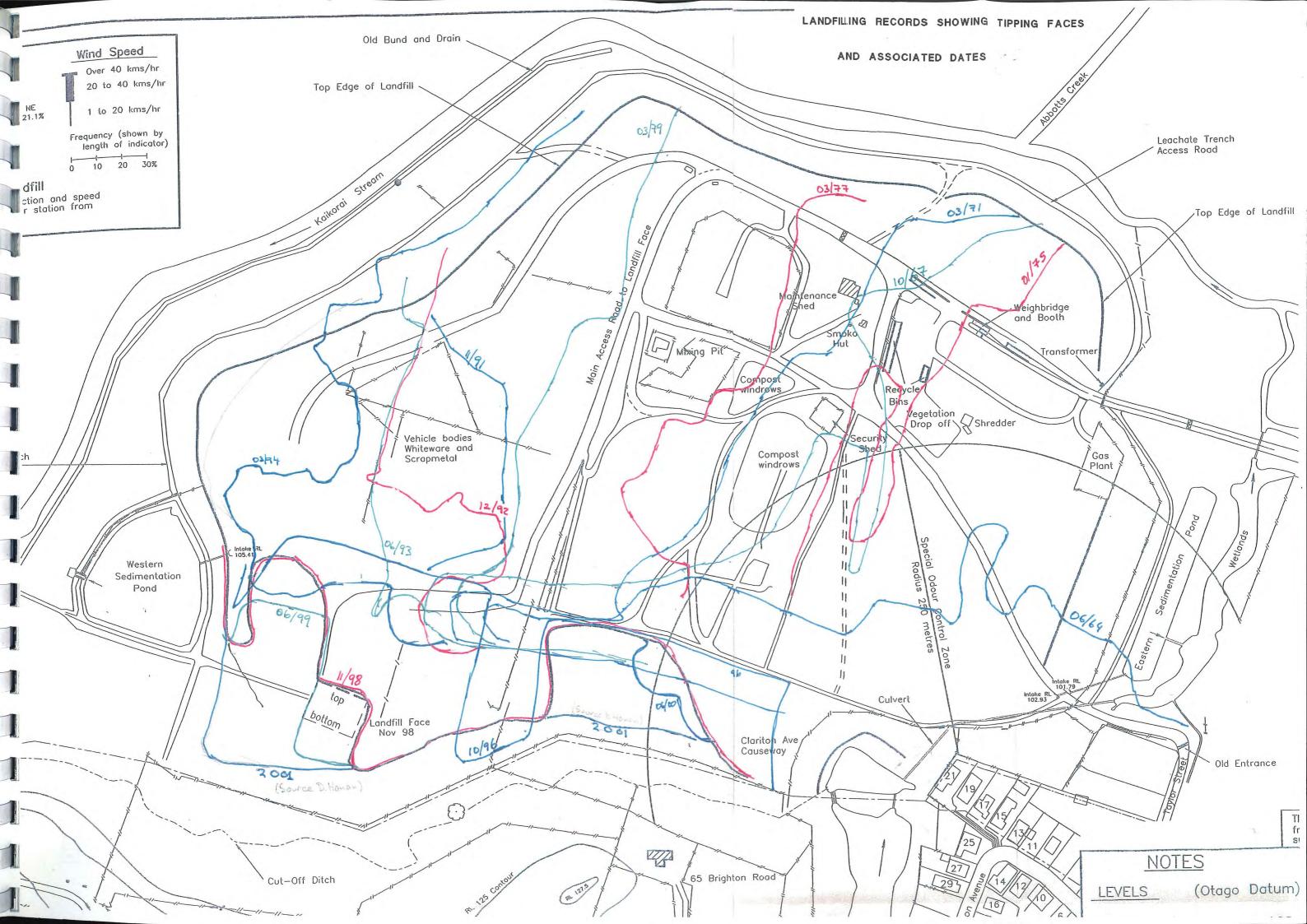
Attachments

Attachment A Historical Plans and Figures









Attachment B Borehole Logs

GLOSSARY OF SYMBOLS



Bentonite

This standard sheet should be read in conjunction with all test hole log sheets and any idealised geological sections prepared for the investigation report.

GENERAL ABBREVIATIONS

Activity type / drilling method

Dual tube OP Observation pit/trench Pressuremeter test hole Casing advancement PMCA **EXP PQTT** PQ triple tube coring Logged exposure **GCOP** GCO probe RC Rotary cored

RCG

HA Hand Auger Rotary drilling in common ground HV Hvdro Vacuum excavation RO Rotary open hole

HQTT HQ triple tube coring SCP Static cone penetrometer

ICBR In situ CBR test SH Shaft

IDFN In situ density test SNC Sonic core drilling **INST** Instrument SPT Standard penetration test

IVAN In situ vane test TP Trial pit/trench MHA Machine Hollow auger Triple tube coring TT MSA Machine Solid auger VC Vibrocore NQ triple tube coring W Wash boring **NQTT**

Open barrel OB

Sampling type

AMAL Amalgamated sample LB Large bulk disturbed sample (for earthworks testing)

Bulk disturbed sample LDS В Large Disturbed Sample **BLK** Block sample Μ Mazier type sample С Core sample Ρ Piston sample

TW **CBR** CBR mould sample Thin walled push in sample D Small disturbed sample U Undisturbed sample - open drive ES Soil sample for environmental testing U100 U110 Undisturbed Sample EW Water sample for environmental testing U76 U76 Undisturbed Sample

Thin wall open drive tube sampler G Gas sample UT

J Jar W Water sample

Other testing

Falling Head Permeability Test Total blows - SPT Value Ν

PK Packer Test

PP Pocket Penetrometer (suffixed by value in kPa)

PT Pressuremeter Test

R Rising Head Permeability Test

Shear Vane Test (suffixed by value in kPa, peak/residual values) SV

UTP Unable to penetrate (shear vane testing)

Target depth TD HCL Hydrochloric acid

WELL SYMBOLS

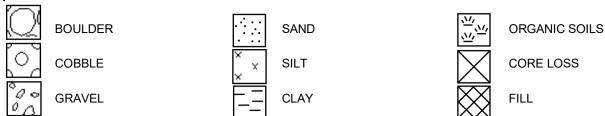


GROUNDWATER SYMBOLS



SOIL SYMBOLS

Main Components



Note: Composite soil types will be signified by combined symbols, e.g.



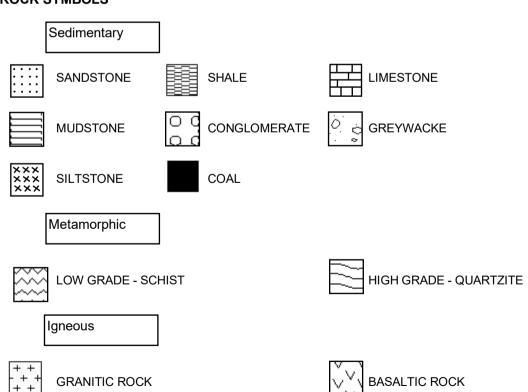
SOIL DESCRIPTION ABBREVIATIONS

Consiste	ency	S-F	Soft to firm
D	Dense	St	Stiff
D-VD	Dense to very dense	St-VSt	Stiff to very stiff
F	Firm	VD	Very dense
F-St	Firm to stiff	VL	Very loose
Н	Hard	VL-L	Very loose to loose
L	Loose	VS	Very soft
L-MD	Loose to medium dense	VS-S	Very soft to soft
MD	Medium dense	VSt	Very stiff
MD-D	Medium dense to dense	VSt-H	Very Stiff to hard
S	Soft		•

Moisture Condition

ט	Dry
D-M	Dry to moist
M	Moist
M-W	Moist to wet
S	Saturated
W	Wet

ROCK SYMBOLS



Note: Additional rock symbols may be allocated for a particular project. Interbedded rock will be represented using alternatively the above symbols

ROCK DESCRIPTION ABBREVIATIONS

Rock Strength

Extremely weak EW EW - VW Extremely to very weak

VW Very weak VW - W Very weak to weak

W Weak

W-MS Weak to moderately strong

MS Moderately strong

MS - S Moderately strong to strong

S Strona

S - VSStrong to very strong

Very strong VS

VS - ES Very strong to extremely strong

ES Extremely strong

Weathering

RS Residual soil

CW-RS Completely weathered to residual soil

CW Completely weathered

Highly weathered to completely weathered HW-CW

HW Highly weathered

Moderately weathered to highly weathered MW-HW

Moderated weathering MW

SW-MW Slightly weathered to moderately weathered

Slightly Weathered SW

UW-SW Unweathered to slightly weathered

UW Unweathered (fresh)

DEFECT DESCRIPTION ABBREVIATIONS

Fracture Type

Bedding Plane СВ Cross Bed

CI Cleavage Crushed Seam CS

CZ Crush zone FΙ Foliation

FΖ Fractured Zone (>250 mm)

JS Joint set JΤ Joint

SF **Sheared Surface**

SM Seam

Sheared Seam SS

SZ Sheared Zone (>250 mm)

VN Vein

Inclination

Sub-horizontal SB G Gently inclined Μ Moderately inclined S Steeply inclined

VS Very steeply inclined

SV Sub-vertical **Aperture**

Т Tight VN

Very Narrow Narrow Ν

Moderately Narrow MN MW Moderately Wide

Wide W VW Very Wide

Roughness

Slickensided sl Rough Smooth sm

Texture

Ы Planar St Stepped U Undulating

Infilling or Coating

CN Clean

Carbonaceous Χ

CLAY Clay Chlorite KT Calcite CA Iron Oxide Fe MΙ Micaceous Quartz Ω7 VΕ Veneer

Joint Set Counts

X 2 2 joints

X 3 3 joints

X 4 4 joints

X 5 5 joints X 6

6 joints

X 7 7 joints

8 joints 8 X

9 joints X 9

> 10 joints

Spacing

EC Extremely closely spaced VC Very closely spaced С Closely spaced

MWModerately widely spaced

W Widely spaced VW Very widely spaced

Core Recovery Parameters

100

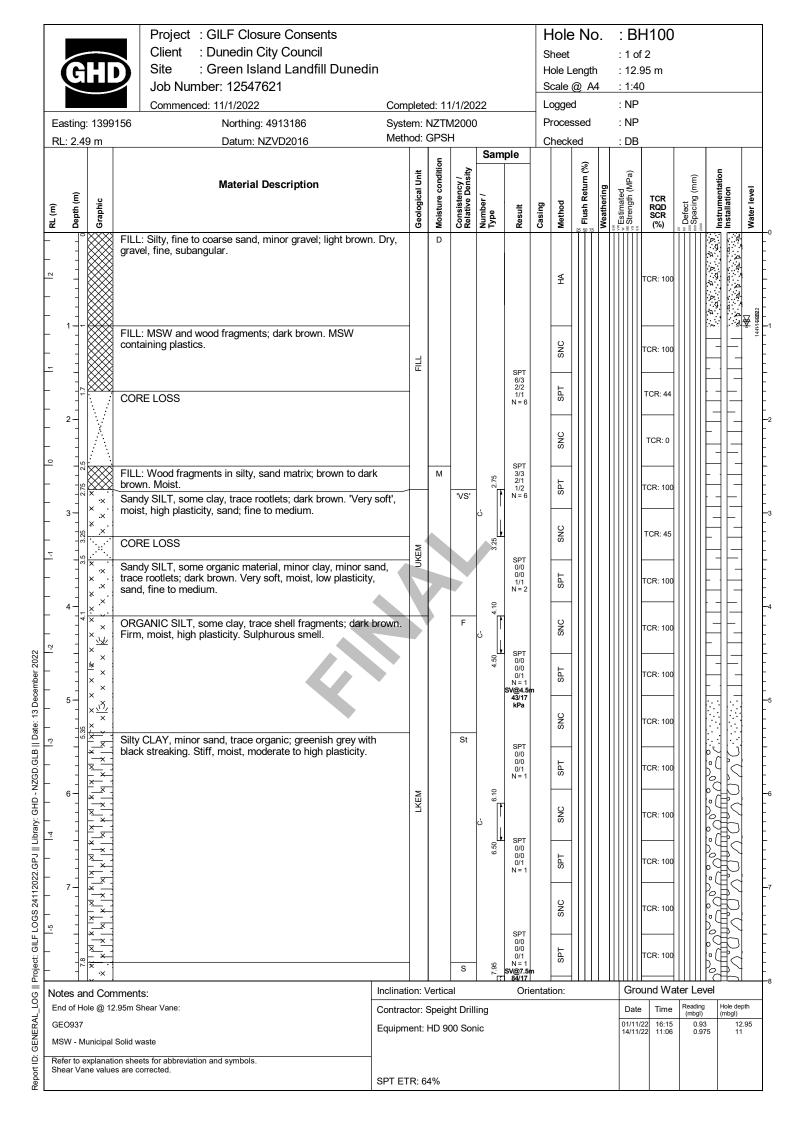
TCR - Total Core Recovery % SCR - Solid Core Recovery %

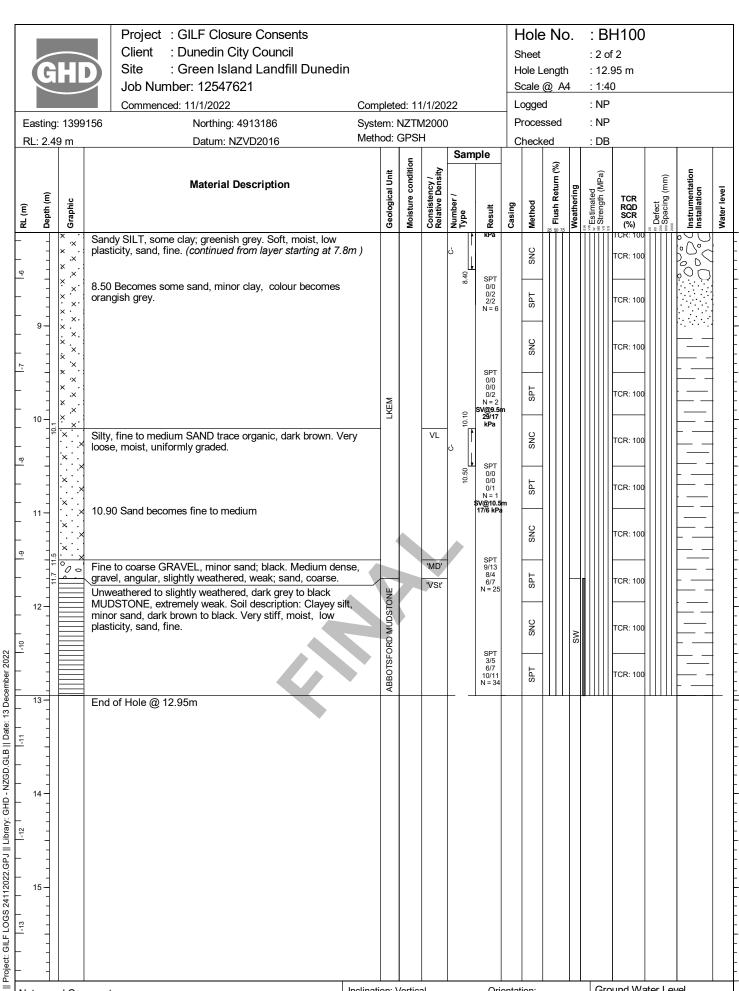
RQD - Rock Quality Designation %

Visual Defects



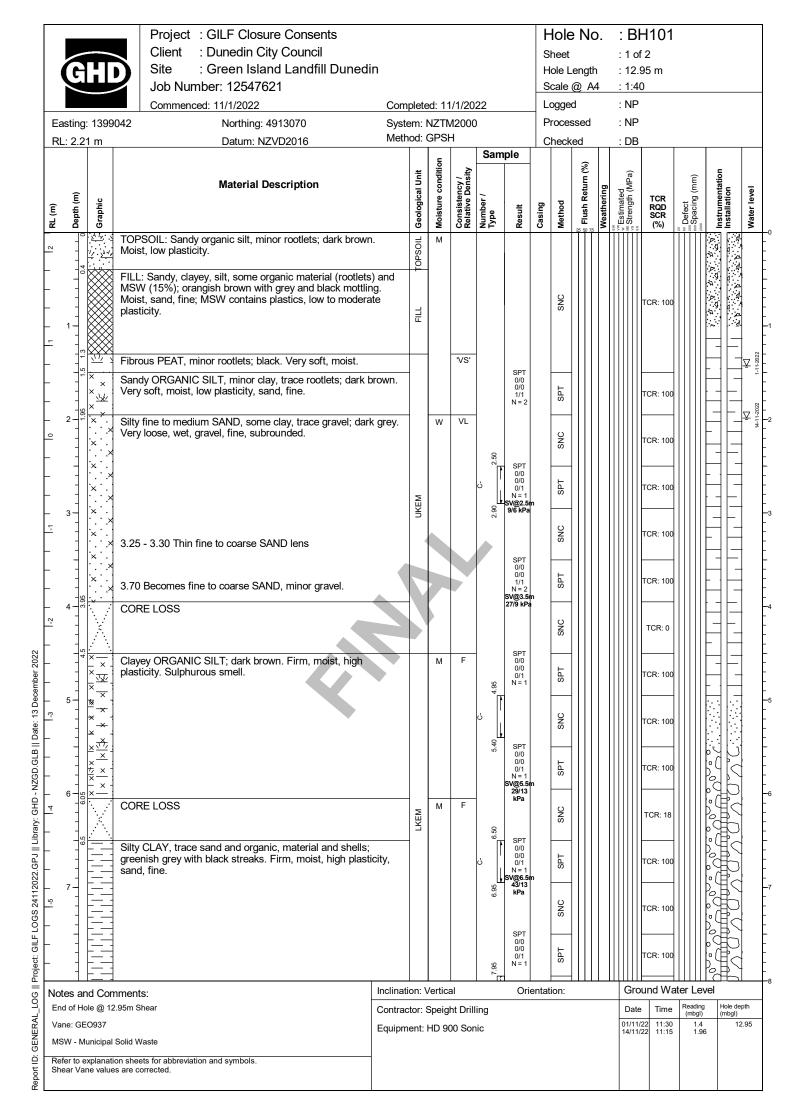
Visual representation of defect angle from horizontal (example shown is 45°)

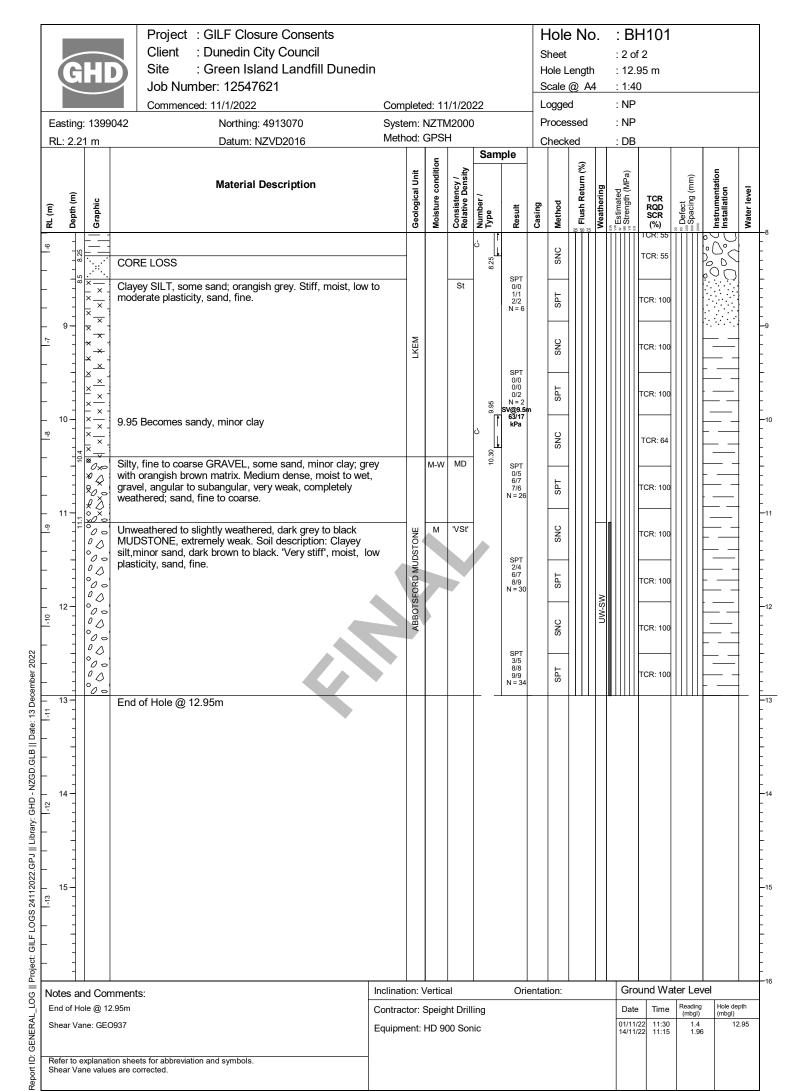


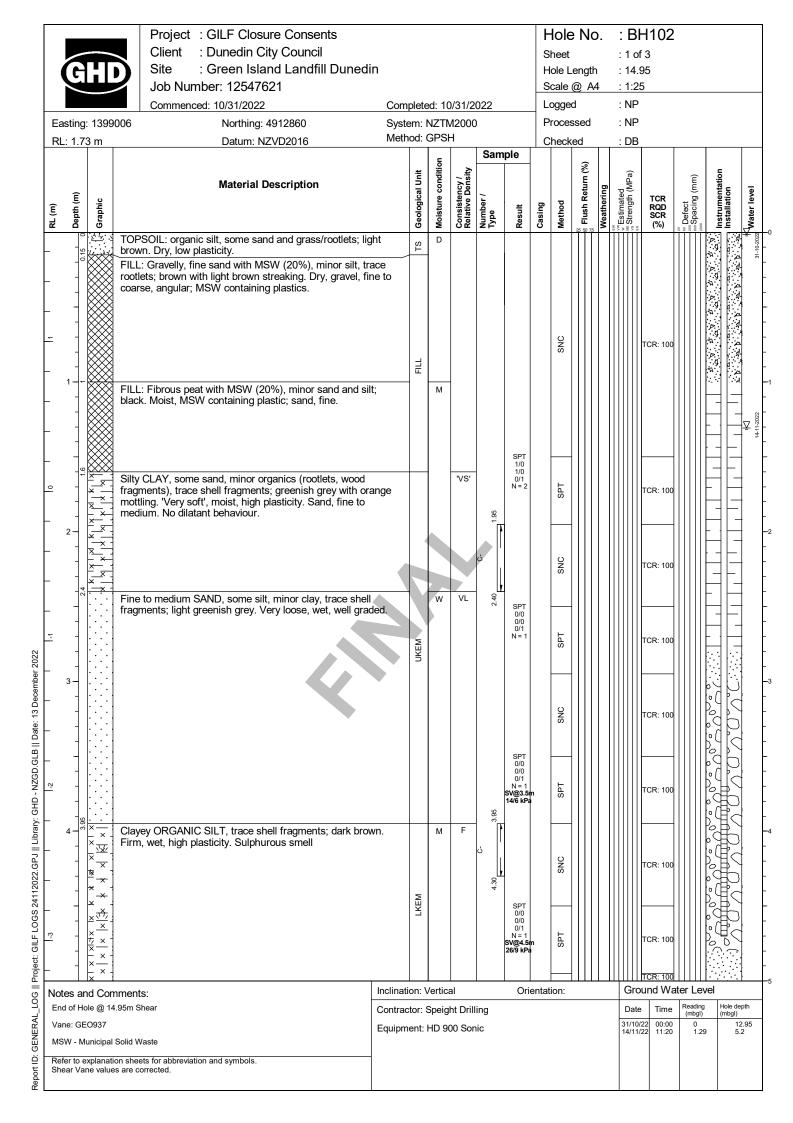


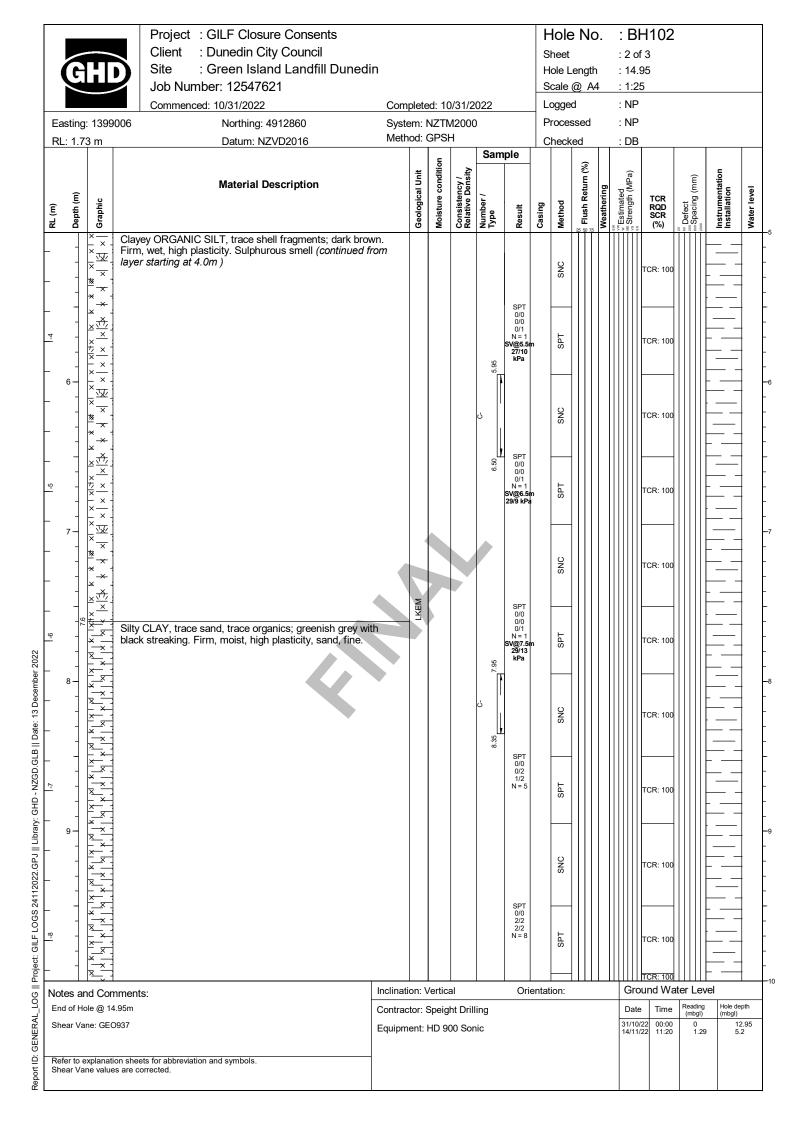
|| Project: GILF LOGS 24112022.GPJ || Library: GENERAL_LOG ≘ Report I

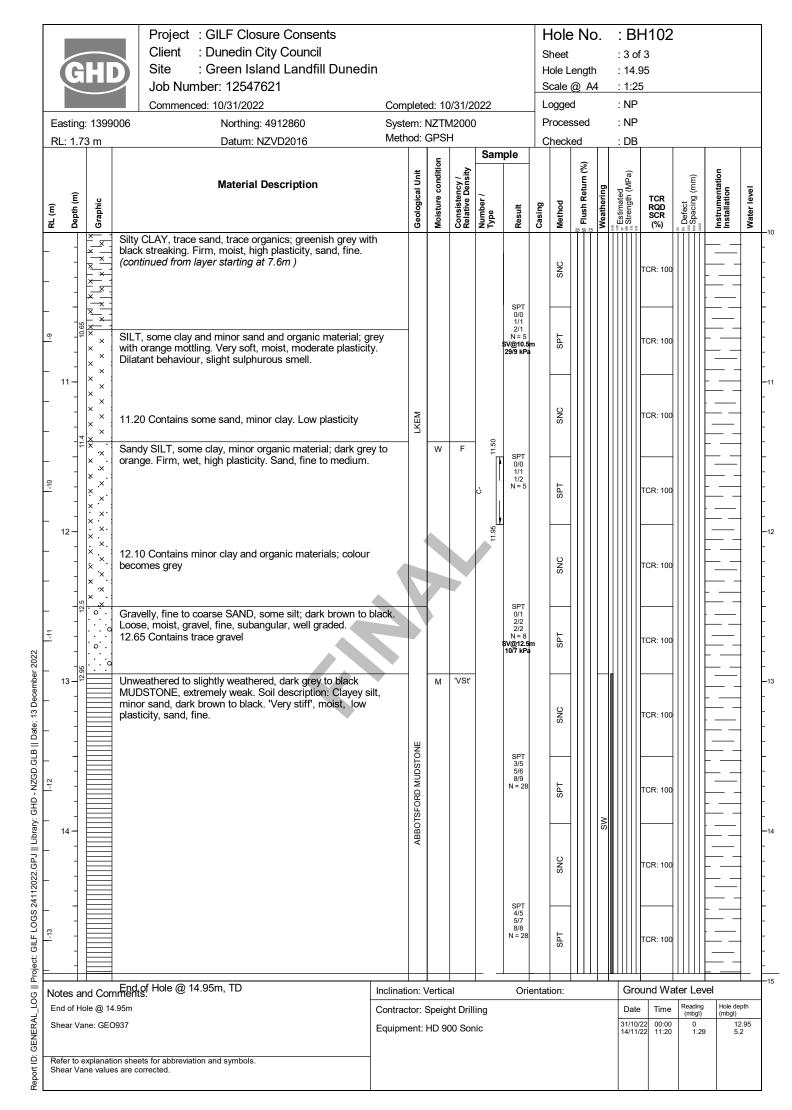
Ground Water Level Orientation: Inclination: Vertical Notes and Comments: End of Hole @ 12.95m Date Contractor: Speight Drilling Time Shear Vane: GEO937 01/11/22 14/11/22 16:15 11:06 0.93 0.975 12.95 11 Equipment: HD 900 Sonic Refer to explanation sheets for abbreviation and symbols. Shear Vane values are corrected. SPT ETR: 64%

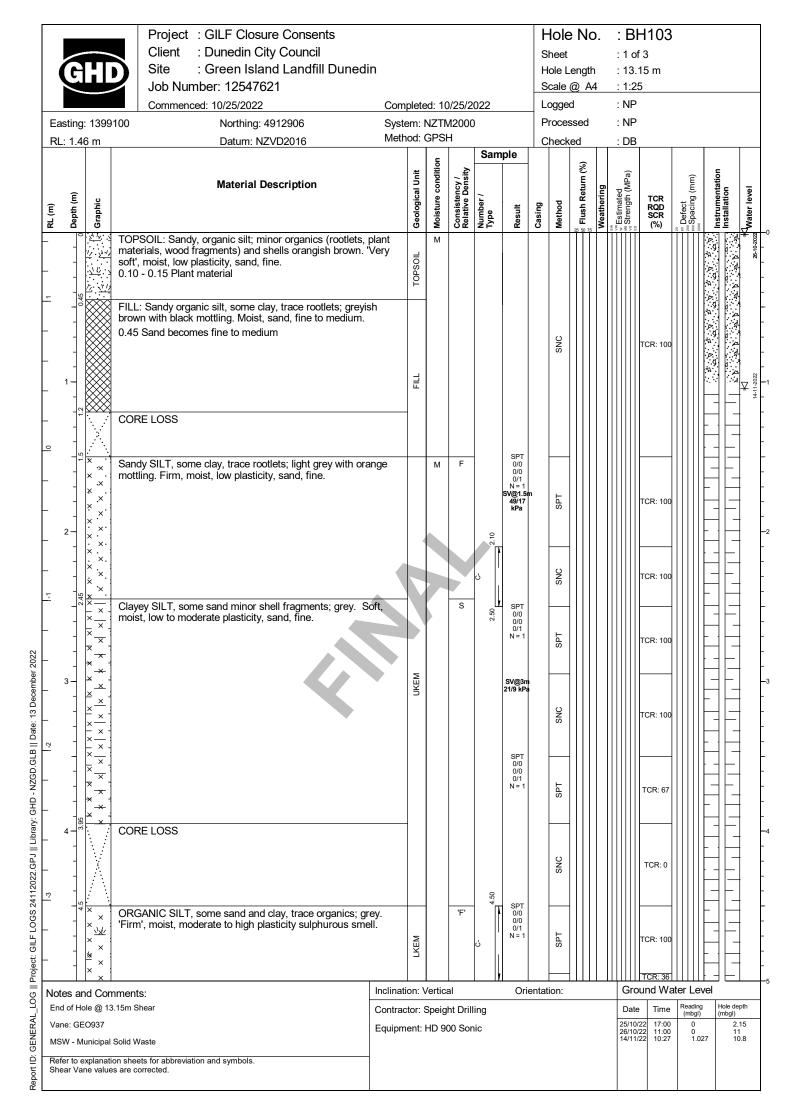


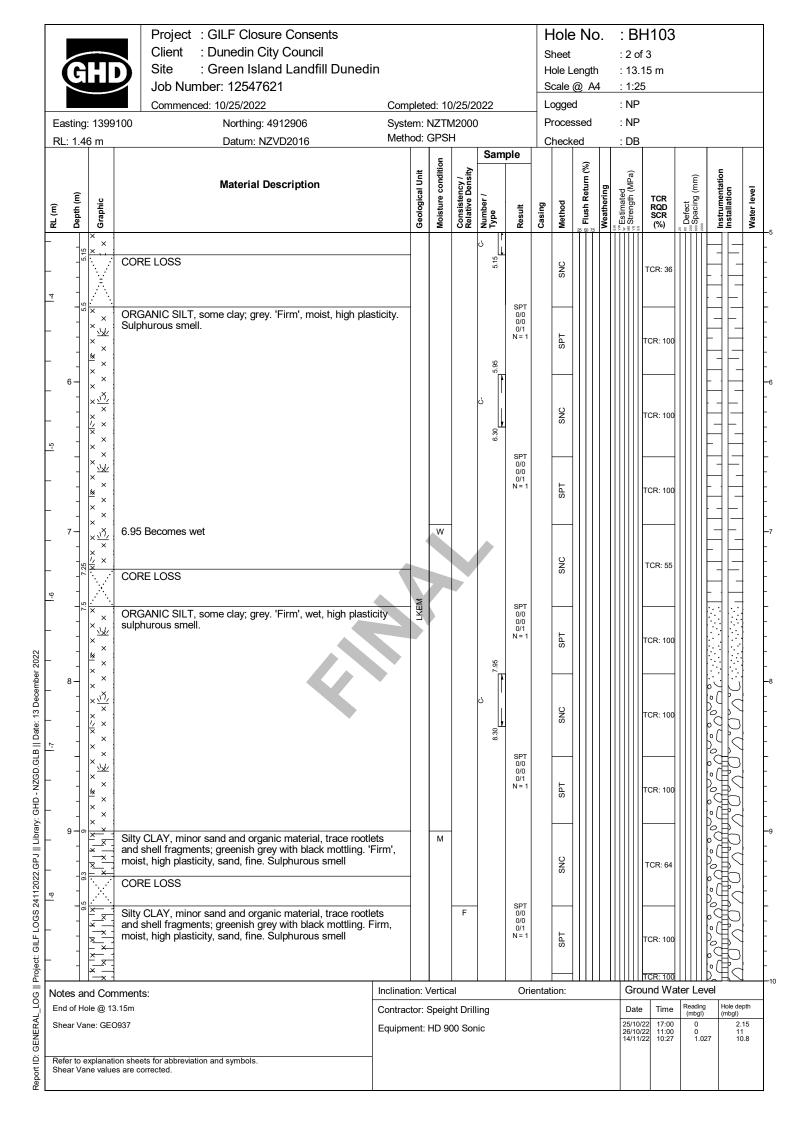


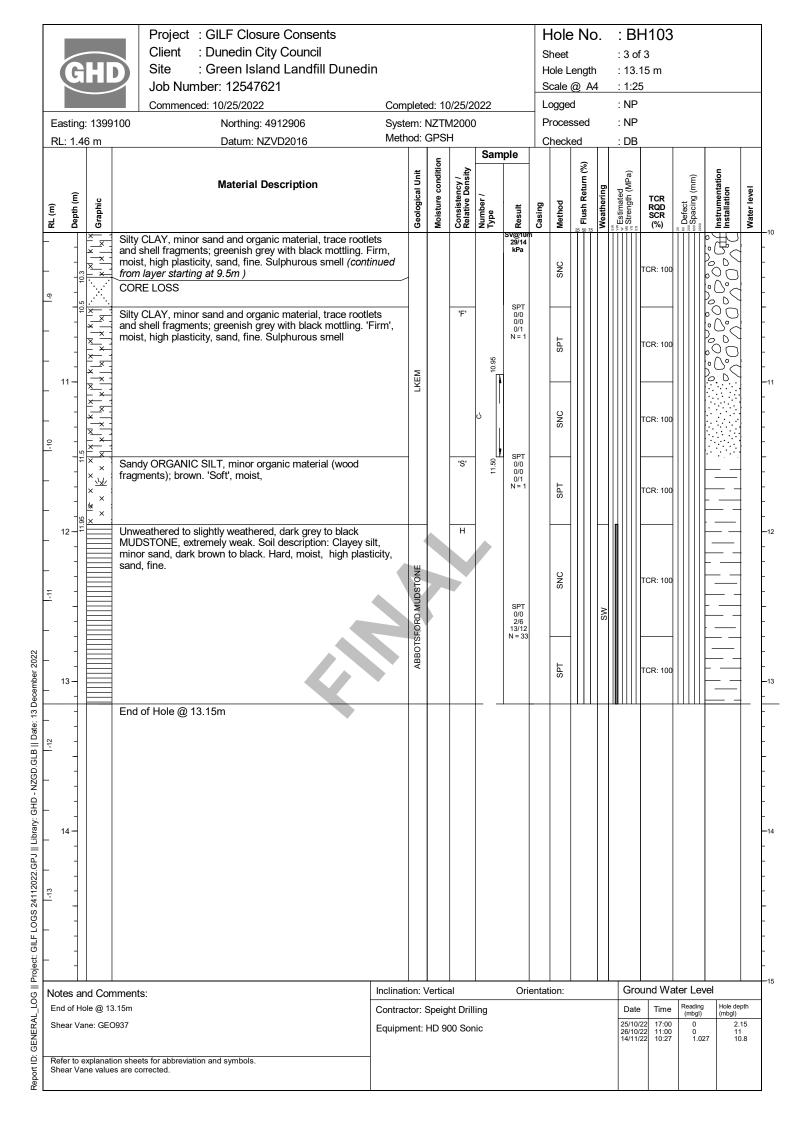


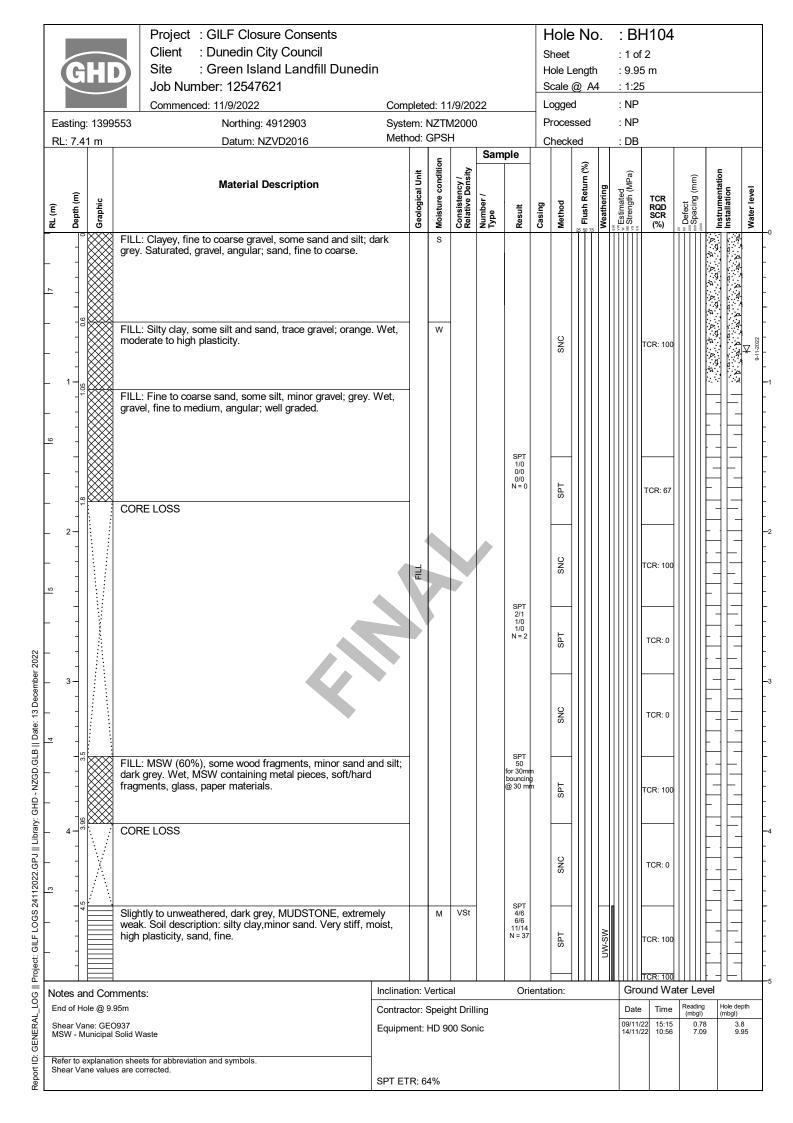


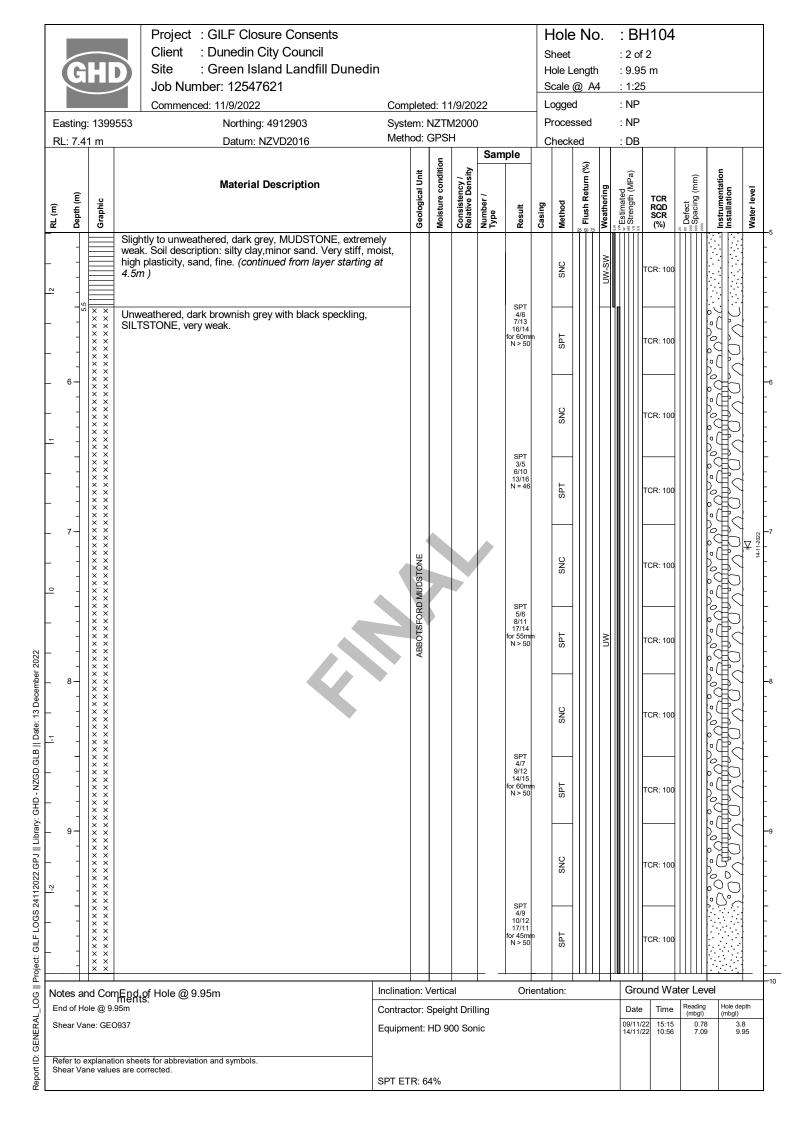


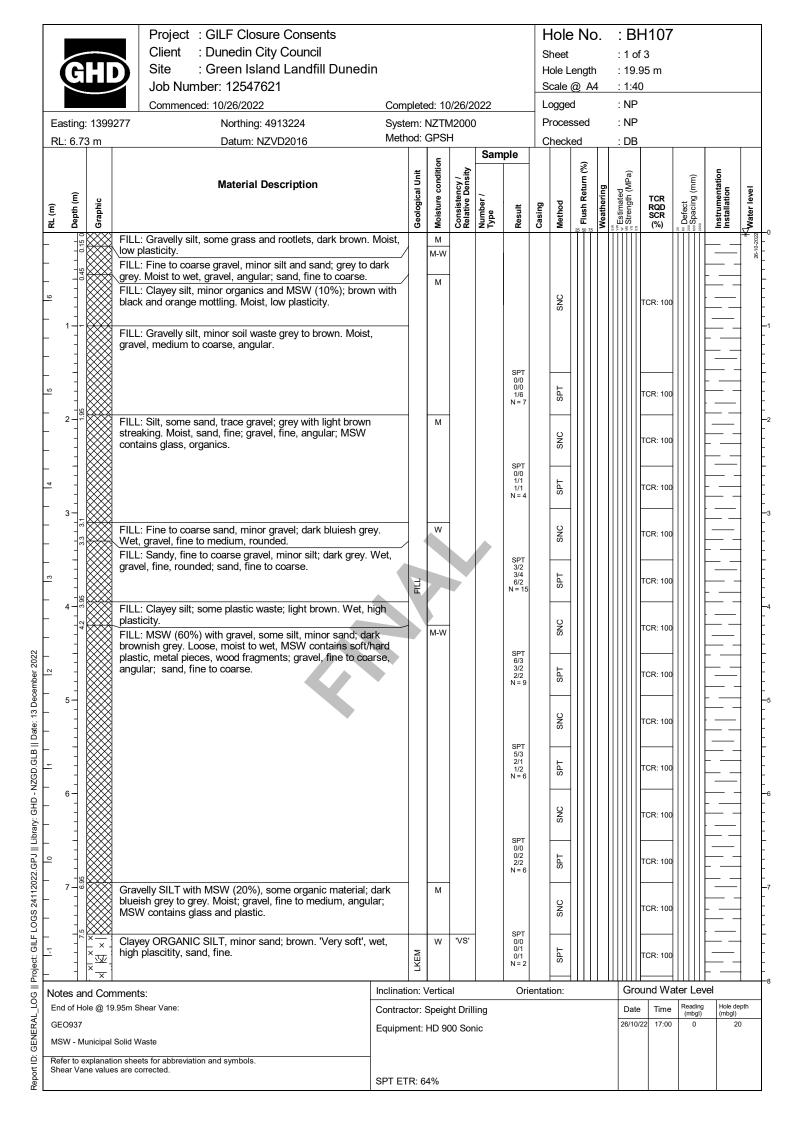












Project : GILF Closure Consents Hole No. : BH107 Client : Dunedin City Council Sheet : 2 of 3 Site : Green Island Landfill Dunedin : 19.95 m Hole Length Job Number: 12547621 Scale @ A4 : 1:40 Commenced: 10/26/2022 Logged : NP Completed: 10/26/2022 : NP Processed Easting: 1399277 Northing: 4913224 System: NZTM2000 Method: GPSH RL: 6.73 m Checked Datum: NZVD2016 : DB Sample Moisture condition 8 Consistency / Relative Density **Geological Unit** Strength (MPa) Spacing (mm) Flush Return Material Description Water level Estimated, Number / Type TCR RQD SCR Graphic Method Defect Casing RL (m) Depth (Result CORE LOSS SNC TCR: 27 Clayey ORGANIC SILT, minor sand; brown. 'Very soft', wet, 0/0 0/0 0/1 × high plascitity, sand, fine. SPT TCR: 100 砅 N= SNC TCR: 100 0/0 1/1 0/1 N = 3 SPT TCR: 78 × SNC W TCR: 100 Sandy SILT, trace organics and shell fragments; grey with orange mottling. 'Very soft', moist, low plasticity, sand, fine to 'VS' 0/0 0/0 0/1 N = ·× SPT TCR: 100 ٠× SNC LKEM TCR: 100 SPT 0/0 0/0 0/1 N = SPT TCR: 89 SNC TCR: 100 GHD - NZGD.GLB || Date: 13 December 2022 0/0 0/0 0/1 N = 4 φ SPT TCR: 73 SNC TCR: 100 0/0 0/0 2/2 N = 4 .× SPT TCR: 100 Clayey ORGANIC SILT, minor sand; brown. 'Very soft', wet, 'VS' W × high plascitity, sand, fine. SNC TCR: 100 L Silty, fine to medium SAND, grey to brownish grey. 'Loose', LOGS 24112022.GPJ || Library: SPT 15/50 for 0mm bouncing @ 50 mm 14.40 Sand becomes fine to coarse TCR: 100 VD 0 φ COBBLE, some gravel; black with white flecks. Very dense, ABBOTSFORD MUDSTONE moist, cobble, angular; gravel, medium to coarse, angular. Moderately weathered SILTSTONE, extremely weak. Soil SNC TCR: 100 description: silt, minor sand; greenish grey with orange mottling. Hard, wet, low plasticity, sand, fine. Ή' W || Project: GILF 5/6 7/12 16/15 N > 50 ဝှ SPT **Ground Water Level** Inclination: Vertical Orientation: Notes and Comments: GENERAL_LOG End of Hole @ 19.95m Contractor: Speight Drilling Time Shear Vane: GEO937 26/10/22 17:00 20 Equipment: HD 900 Sonic

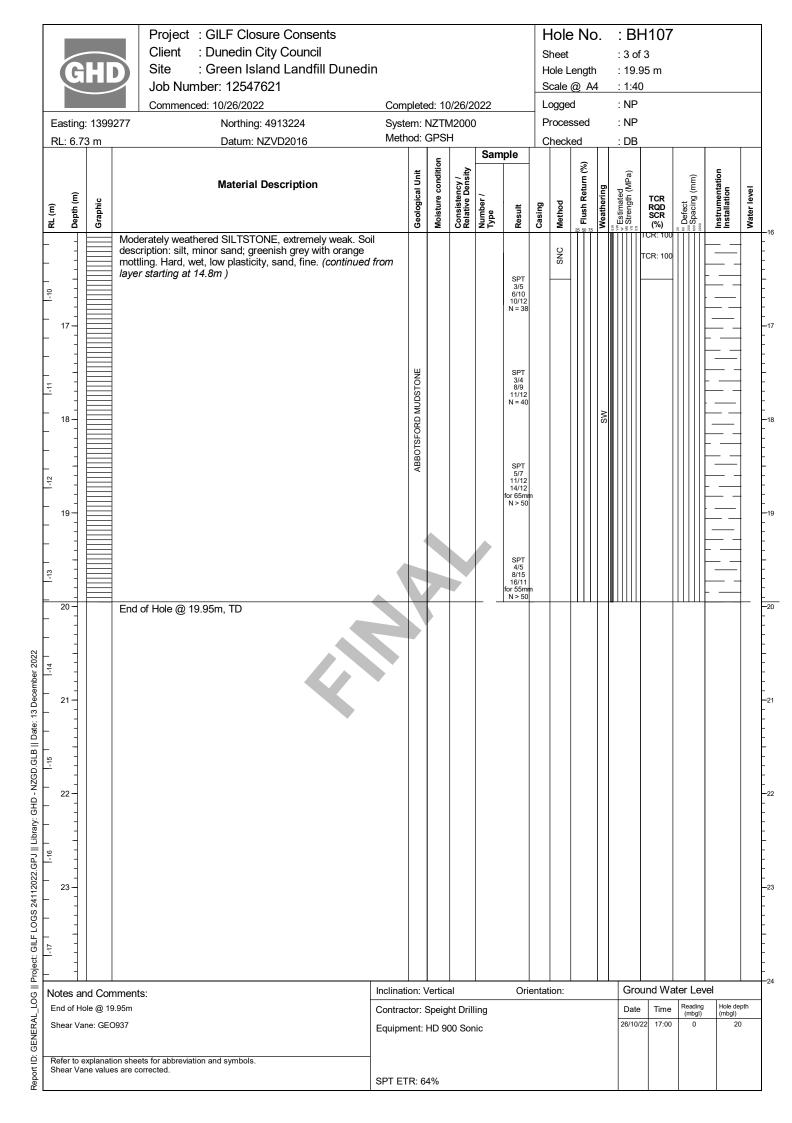
SPT ETR: 64%

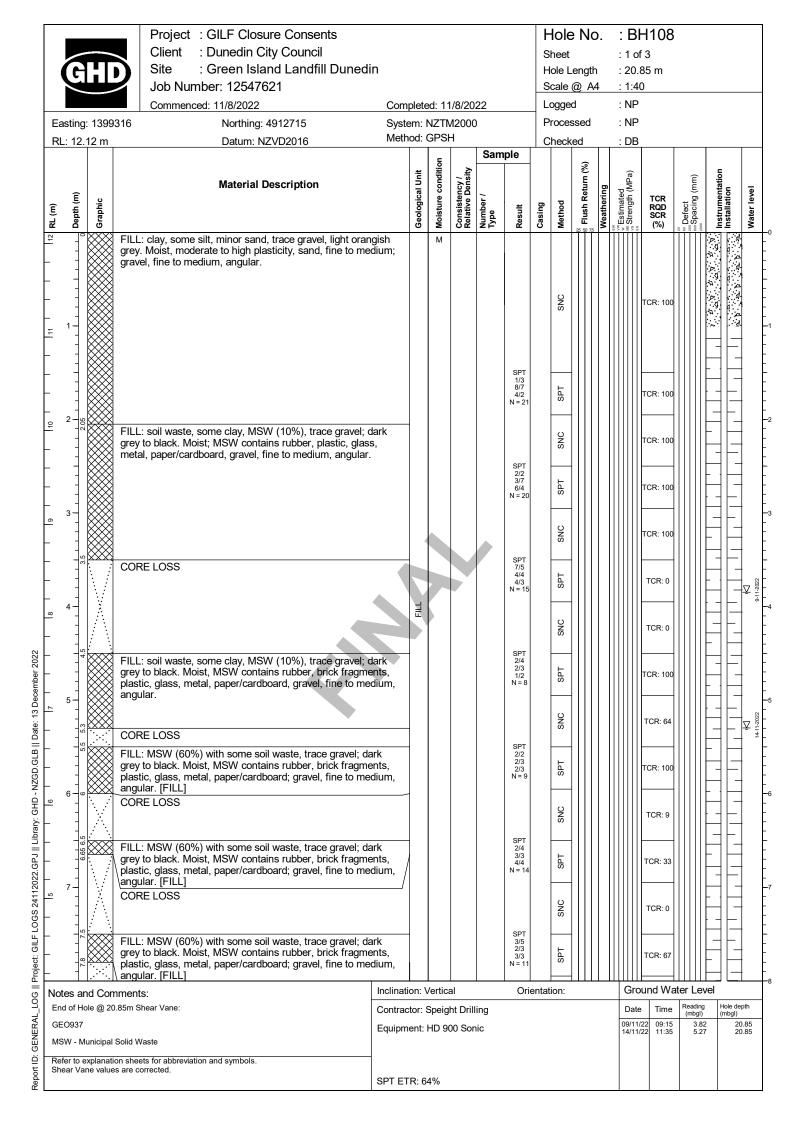
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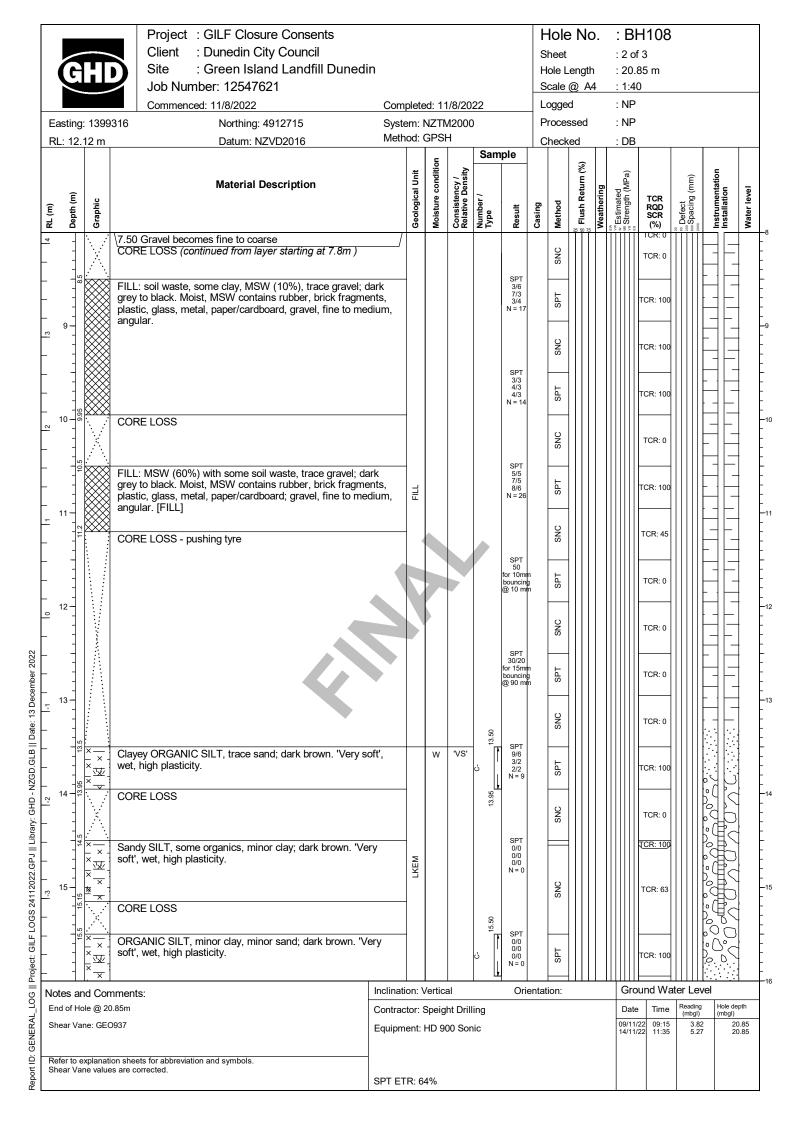
Report I

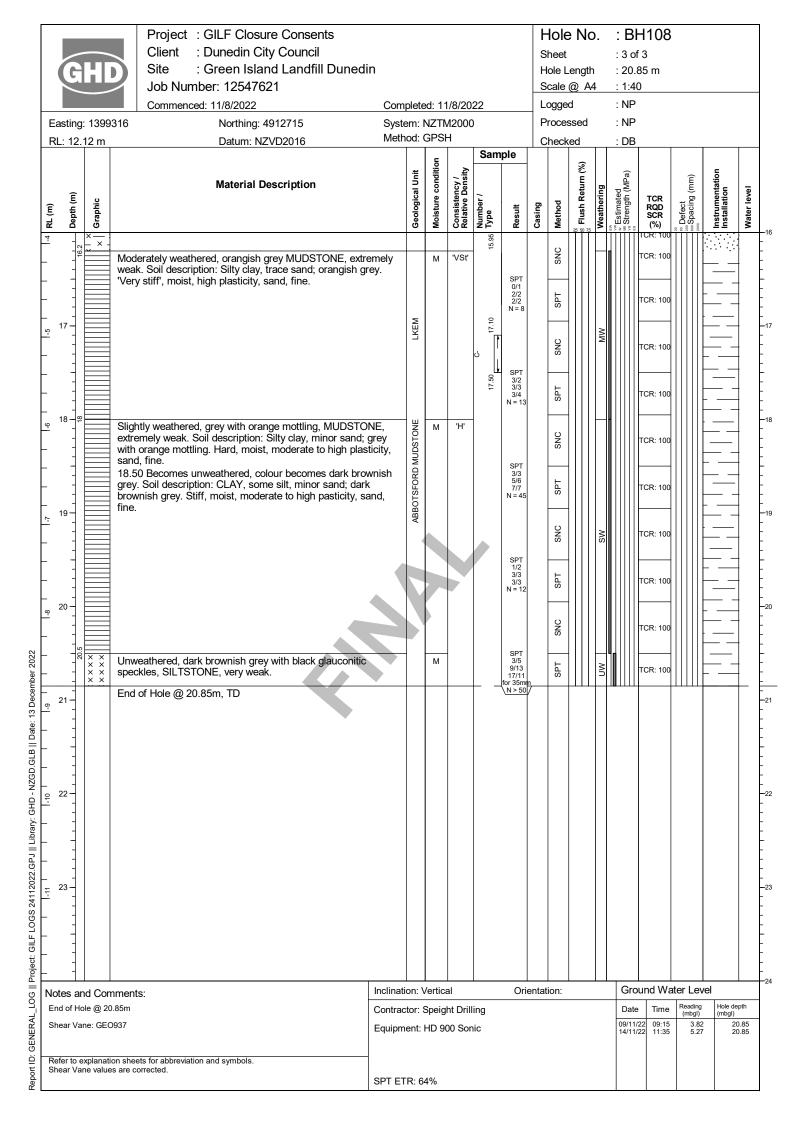
Refer to explanation sheets for abbreviation and symbols.

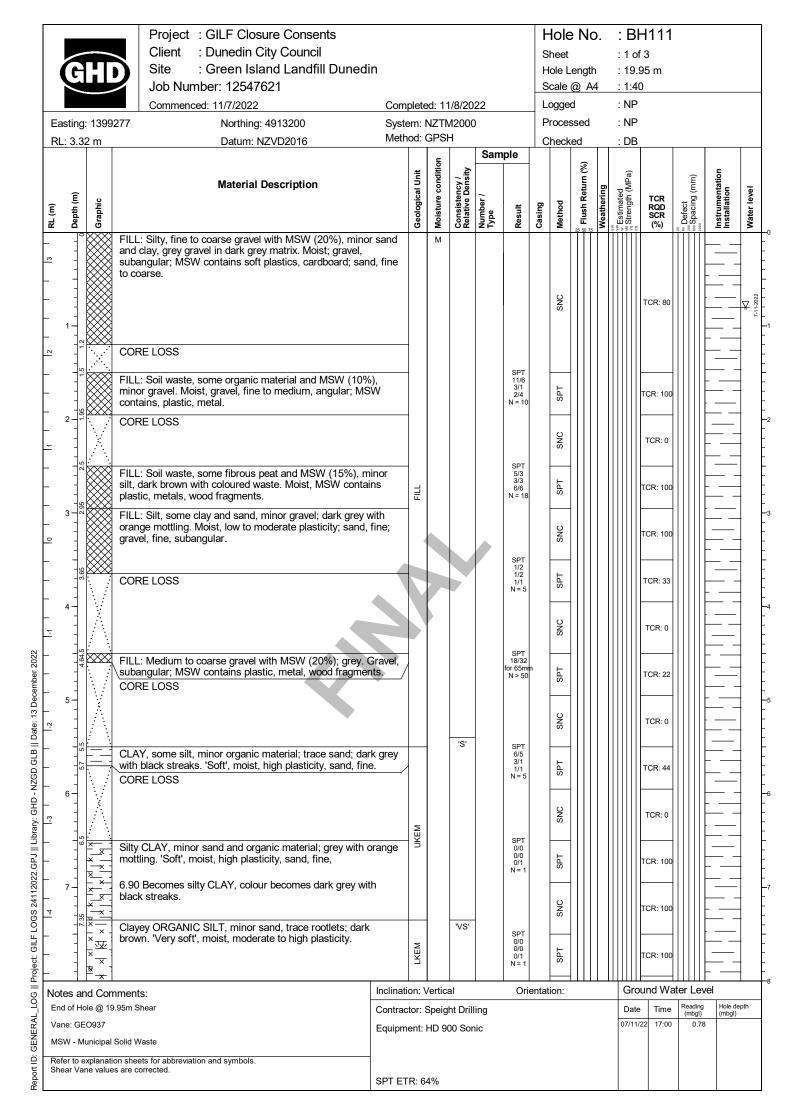
Shear Vane values are corrected.

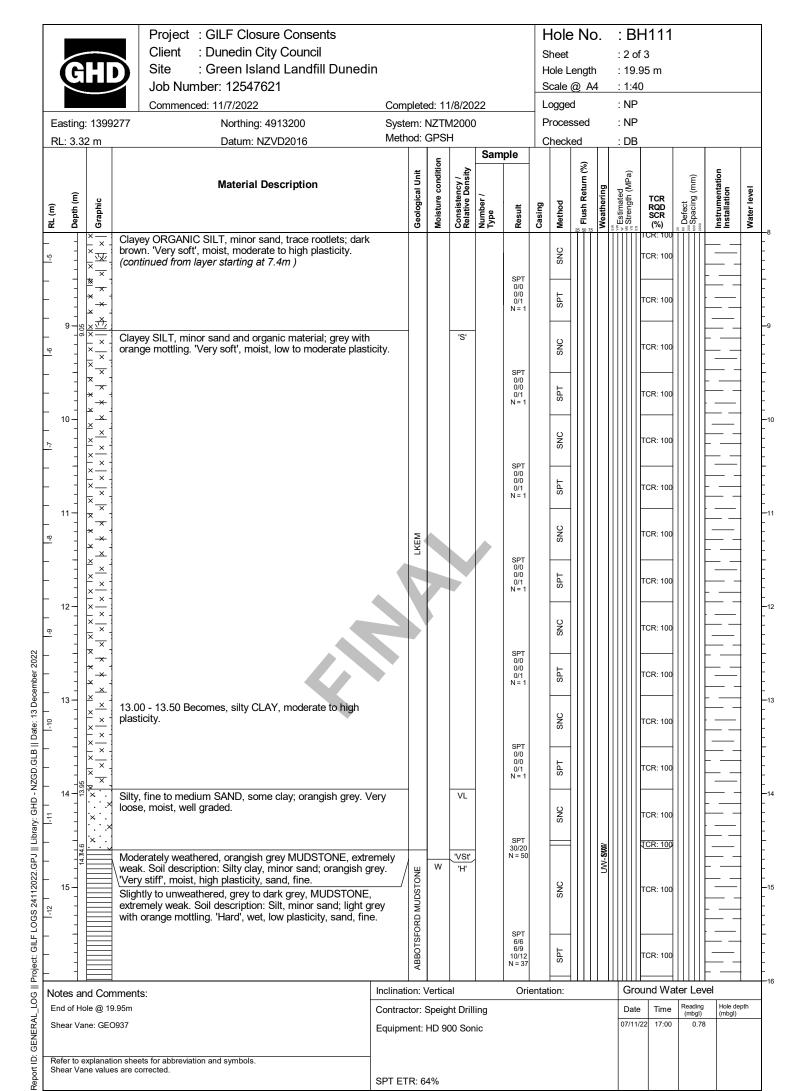




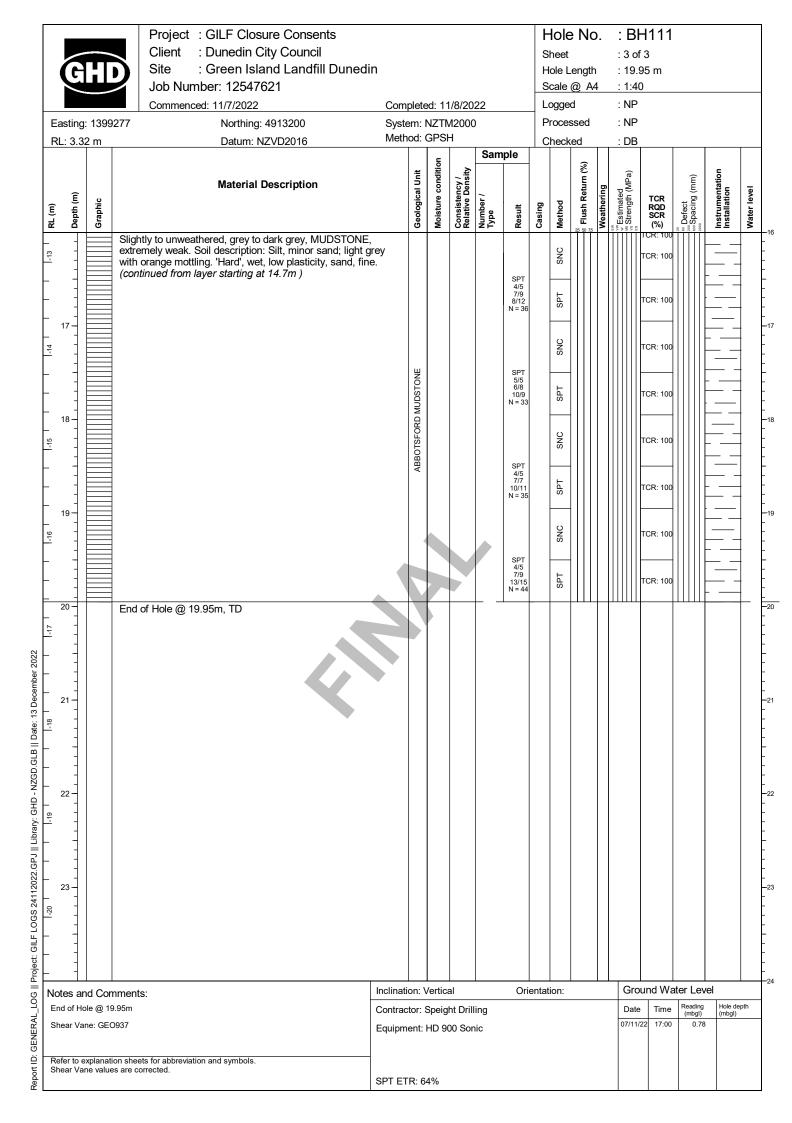








SPT ETR: 64%



Attachment C Analytical Results Tables



Lab Reference		22-40179-12	22-40179-13	22-40179-1	22-40179-2	22-40179-3	22-40179-4	22-40179-14	22-40179-15	22-40179-16	22-40179-17			
Sample Depth	m bgl	1.0	2.7	0.5	1.5	2.5	3.5	0.6	1.5	2.5	3.5	NESCS SCSs for Protection of	NESCS SCSs for Protection	
Borehole	_ ĭ		100		BH-				ВН	-102		Human Health based on a	of Human Health based on	Dunedin Background Soil
Sample Name		12547621-BH-100(1.0)	12547621-BH-100(2.7)	12547621-BH-101(0.5)	12547621-BH-101(1.5)	12547621-BH-101(2.5)	12547621-BH-101(3.5)	12547621-BH-102(0.6)	12547621-BH-102(1.5)	12547621-BH-102(2.5)	12547621-BH-102(3.5)	Commercial/Industrial land	a Recreational land use	Concentrations
Soil type		Fill/ sand/ waste	Fill/ wood in sand	Fill/ loam/waste	Peat/sandy silt	Silty sand	Silty sand	Fill/ sand/ waste	Fill/ peat/ waste	Sand	Sand	use	(mg/kg) 1 & 2	(mg/kg) ³
Sample Date		01/11/2022	01/11/2022	01/11/2022	01/11/2022	01/11/2022	01/11/2022	01/11/2022	01/11/2022	01/11/2022	01/11/2022	(mg/kg) ^{1 & 2}	(6)	
Heavy Metals														
Arsenic	mg/kg	2.8	7.8	9.7	12	7.1	5.5	4.7	29.7	11	11	70 ¹	80 ¹	11.77
Boron	mg/kg	1.5	9.2	6.9	27	8.4	11	2.1	32	9.9	9	>10,000 ¹	>10,000 ¹	
Cadmium	mg/kg	0.059	0.047	0.3	0.18	0.018	0.029	0.11	0.27	0.012	0.066	1,300 ¹	400 ¹	0.34
Chromium ⁴	mg/kg	21.6	28	29.3	55.2	21.8	18.8	46.2	125	23.7	22.5	6,300 ¹	2,700 ¹	80.15
Copper	mg/kg	18.5	6.6	14.2	18.2	4.1	4	27.3	34.2	6.5	5.5	>10,000 1	>10,000 1	60.85
Lead	mg/kg	14.8	5.01	60.1	50.8	4.9	4.2	44.8	98.5	6.24	4.7	3,300 ¹	880 ¹	44.34
-		0.055	<0.025	0.079	0.14	<0.025	<0.025	0.12	0.32	<0.025	<0.025			-
Mercury	mg/kg									+		4,200 1	1,800 1	44.96
Nickel	mg/kg	14.1	15.5	36.6	14.8	10.6	9.53	35.9	33.8	11.5	15.1	6,000 ²	1,200 2	
Zinc	mg/kg	52	47.4	183	119	36.7	29	105	467	49.4	45.9	400,000 ²	30,000 ²	182.8
Lab Reference	1	22-40179-8	22-40179-9	22-40179-10	22-40179-11	22-41184-5	22 41194 6	22-41184-7	22 40170 F	22-40179-6	22 40170 7			I
Sample Depth	m bgl	0.2	1.2	22-40179-10	3.5	22-41184-5 0.5	22-41184-6 1.0	22-41184-7	22-40179-5 0.5	1.5	22-40179-7 2.5	NESCS SCSs for Protection of	AUTOGO COO (
Borehole	III Dgi	0.2		-103	3.3	0.3	BH-104	2.0	0.3	BH-107	2.3	Human Health based on a	NESCS SCSs for Protection	Dunedin Background Soil
Sample Name		12547621-BH-103(0.2)	12547621-BH-103(1.2)	12547621-BH-103(2.5)	12547621-BH-103(3.5)	12547621-BH-104(0.5)	12547621-BH-104(1.0)	12547621-BH-104(2.0)	12547621-BH-107(0.5)	12547621-BH-107(1.5)	12547621-BH-107(2.5)	Commercial/Industrial land	of Human Health based on	Concentrations
Soil type		Topsoil/ organic silt	Fill/ organic silt	Clayey silt	Clayey silt	Fill/ clayey gravel	Fill/ silty clay	Fill/ waste material	Clayey silt/ waste	Gravelly silt/ bricks	Silt/ waste	use	a Recreational land use (mg/kg) ^{1 & 2}	(mg/kg) ³
Sample Date		01/11/2022	01/11/2022	01/11/2022	01/11/2022	09/11/2022	09/11/2022	09/11/2022	01/11/2022	01/11/2022	01/11/2022	(mg/kg) 1 & 2	(mg/kg)	
Heavy Metals					01/11/2022									
Arsenic	mg/kg	6.8	25.8	8.7	9.4	7.7	4.1	8.9	5	4.8	4.6	70 ¹	80 ¹	11.77
Beryllium	mg/kg	-	-	-	-	1.79	0.91	0.81	-	_	_	500 ²	90 ²	_
Boron	mg/kg	3.4	20	7.1	9.5	3.4	2.5	11	5.5	14	7.2	>10,000 ¹	>10,000 ¹	_
Cadmium	mg/kg	0.29	0.28	0.12	0.07	0.15	0.07	0.12	0.052	0.06	0.11	1,300 ¹	400 ¹	0.34
	mg/kg	20.1	156	19.8	19.1	46.6	27.3	27.6	18.4	39.2	24.2	6,300 ¹	2,700 ¹	80.15
Chromium *	mg/kg	20.3	31.6	6.4	6.7	38.3	18	16.9	8.14	17.7	16.2	· .	>10,000 ¹	60.85
Copper				+						17.7		>10,000 1	·	44.34
Lead	mg/kg	77.8	74.3	5.45	6.78	70.7	26.9	16.5	19.7		32.1	3,300 1	880 1	44.34
Mercury	mg/kg	0.1	0.3	0.029	<0.025	0.078	0.033	0.039	0.13	0.091	0.097	4,200 1	1,800 1	-
Nickel	mg/kg	19.1	30.7	13.2	12.2	35.8	19.3	18.9	10.2	25.6	9.59	6,000 ²	1,200 2	44.96
Zinc	mg/kg	120	210	42.7	49.3	115	68.1	68.3	37	67.7	53	400,000 2	30,000 ²	182.8
Lab Reference	1	22-41184-1	22-41184-2	22-41184-4*	22-41184-3	22-40820-1	22-40820-2	22-40820-3	22-40820-4**	1	1			I
Sample Depth	m bgl	0.5	1.9	1.9	3.3	0.5	1.5	3.2	3.2	1		NESCS SCSs for Protection of	NECOCOCO (D	
Borehole	48.	0.5		-108	5.5	0.5		-111	0.2	1		Human Health based on a	NESCS SCSs for Protection of Human Health based on	Dunedin Background Soil
Sample Name		12547621-BH-108(0.5)	12547621-BH-108(1.9)	12547621-DUP-2	12547621-BH-108(3.3)	12547621-BH-111(0.5)	12547621-BH-111(1.5)	12547621-BH-111(3.2)	12547621-Dup**	1		Commercial/Industrial land	a Recreational land use	Concentrations
Soil type		Fill/ clay	Fill/ clay	Fill/ clay	Fill/ waste/ clay	Fill/ gravel/ waste	Fill/waste/organics	Fill/ silt	Fill/ silt			use	(mg/kg) 1 & 2	(mg/kg) ³
Sample Date		09/11/2022	09/11/2022	09/11/2022	09/11/2022	07/11/2022	07/11/2022	07/11/2022	07/11/2022]		(mg/kg) 1 & 2	(1116/186)	
Heavy Metals														
Arsenic	mg/kg	2.4	7.2	7.1	9.1	1.7	8.9	9	9.6			70 ¹	80 ¹	11.77
Beryllium	mg/kg	0.48	1.42	1.33	3.15	1.2	1.3	1.35	1.75			500 ²	90 ²	-
Boron	mg/kg	1.3	7.6	7.9	40	8.4	20	14	15			>10,000 1	>10,000 1	-
Cadmium	mg/kg	0.0056	0.16	0.16	0.42	0.11	2.18	0.18	0.32			1,300 ¹	400 ¹	0.34
Chromium ⁴	mg/kg	11	50.1	49.4	34.7	90.5	34.6	52.7	58.3			6,300 ¹	2,700 ¹	80.15
Copper	mg/kg	3.3	20.9	20.8	54.2	59.1	40.3	25.2	34.3			>10,000 1	>10,000 1	60.85
Lead	mg/kg	11.7	52.9	42.8	316	49.1	179	32.9	62.2			3,300 ¹	880 ¹	44.34
Mercury	mg/kg	<0.025	0.28	0.27	0.21	<0.025	0.12	0.12	0.21			4,200 ¹	1,800 ¹	-
Nickel	mg/kg	3.9	38	35.9	31.6	105	32.7	45.4	49	†		6,000 ²	1,200 ²	44.96
Zinc	mg/kg	20.3	124	121	303	127	403	179	236	1		400,000 ²	30,000 ²	182.8
	···ь/ \ъ	20.5	147	121	555	14/	-03	1/3	230	I .	I	400,000	30,000	102.0

Notes:

Values shaded grey exceed the NESCS SCSs for Protection of Human Health (Commercial / Industrial land use) Values shaded yellow exceed the NESCS SCSs for Protection of Human Health (Recreational land use)

Red Text exceed the adopted Dunedin Background Soil Concentrations (mg/kg)

SCS - Soil contaminant standard

- Data not available or not analysed in this sample

All units are in mg/kg

m bgl - metres below ground level

* Field duplicate of 12547621-BH-108(1.9)

** Field duplicate of 12547621-BH-111(3.2)

- 1. Ministry for the Environment (2011). Resource Management (National Environmental Standard for Assessing and Managing Contaminants in Soil to Protect Human Health) Regulations 2011 (NESCS).
- 2. National Environment Protection Council (1999, revised 2013) National Environment Protection (Assessment of Site Contamination) Measure. Table 1A. (NEPM) Health Investigation Levels Commercial / Industrial D.
- 3. Landcare Research Limited (2006) PBC Predicted Background Soil Concentrations, New Zealand fill material (the highest value for each metal from the dominant soil groups surrounding Green Island Landfill) Dunedin , New Zealand

4. NESCS SCS criteria presented are for Chromium (VI)



Analytica Lab Reference		22-40179-12	22-40179-13	22-40179-1	22-40179-2	22-40179-3	22-40179-4	22-40179-14	22-42925-4	22-42925-5					
urofins Lab Reference		K22-No0015984	K22-No0015985	K22-Mo0015975	K22-No0015976	K22-No0015977	K22-No0015978	22-40179-14	-	-				Guidelines for Hydrocarbon	
ample Depth	m bgl	1.0	2.7	0.5	1.5	2.5	3.5	0.6	1.5	2.5	NESCS SCSs for Commercial /	NESCS SCSs for Recreational land	Guidelines for Hydrocarbon Contamination - Residential Use	Contamination -	Background Soil Concentrations, New
est Pit	•	ВН	-100		BH-	101	•		BH-102	•	Industrial land use 1	use ¹		Commercial/Industrial Use -	Zealand ³ (mg/kg)
Sample Name		12547621-BH-100(1.0)	12547621-BH-100(2.7)	12547621-BH-101(0.5)	12547621-BH-101(1.5)	12547621-BH-101(2.5)	12547621-BH-101(3.5)	12547621-BH-102(0.6)	12547621-BH-102(1.5)	12547621-BH-102(2.5)			All Pathways - Sand (<1m / 1-4m)	All Pathways -	
Soil Type		Fill/ sand/ waste	Fill/ wood in sand	Fill/ loam/waste	Peat/sandy silt	Silty sand	Silty sand	Fill/ sand/ waste	Fill/ peat/ waste	Sand	(mg/kg)	(mg/kg)		Sand (<1m / 1-4m)	
Sample Date		01/11/2022	01/11/2022	01/11/2022	01/11/2022	01/11/2022	01/11/2022	01/11/2022	01/11/2022	01/11/2022			(mg/kg) ²	(mg/kg) ²	
Acenaphthene	ma/ka	< 0.03	< 0.03	< 0.03	< 0.03	< 0.03	< 0.03	< 0.03	<0.010	<0.010					0.055
	mg/kg	0.04	< 0.03	< 0.03	< 0.03	< 0.03	< 0.03	< 0.03	<0.010	<0.010	-	-	-	-	
Acenaphthylene	mg/kg		< 0.03		< 0.03	< 0.03	< 0.03		<0.010	<0.010	-	-	-	-	0.069
Anthracene	mg/kg	0.04	1	< 0.03				< 0.03			-	-	-	-	0.113
Benz[a]anthracene	mg/kg	0.17	< 0.03	0.07	0.06	< 0.03	< 0.03	0.1	0.023	<0.020	#	#	-	-	0.47
Benzo[a]pyrene	mg/kg	0.23	< 0.03	0.06	0.05	< 0.03	< 0.03	0.1	0.043	<0.010	#	#	-	-	0.595
Benzo[b]&[j] fluoranthene	mg/kg	0.13	< 0.03	0.04	0.06	< 0.03	< 0.03	0.08	0.054	<0.020	#	#	-	-	0.947
Benzo[g,h,i]perylene	mg/kg	0.08	< 0.03	< 0.03	0.06	< 0.03	< 0.03	0.06	0.028	<0.020	-	-	-	-	0.459
Benzo[k]fluoranthene	mg/kg	0.14	< 0.03	< 0.03	0.06	< 0.03	< 0.03	0.09	0.015	<0.010	#	#	-	-	0.296
Chrysene	mg/kg	0.18	< 0.03	0.09	0.09	< 0.03	< 0.03	0.1	0.038	<0.010	#	#	-	-	0.539
Dibenz(a,h)anthracene	mg/kg	< 0.03	< 0.03	< 0.03	< 0.03	< 0.03	< 0.03	< 0.03	<0.010	<0.010	#	#	-	-	0.112
Fluoranthene	mg/kg	0.34	< 0.03	0.11	0.11	< 0.03	< 0.03	0.08	0.055	<0.020	#	#	-	-	1.345
Fluorene	mg/kg	< 0.03	< 0.03	< 0.03	< 0.03	< 0.03	< 0.03	< 0.03	< 0.010	< 0.010	-	-	-	-	0.06
Indeno(1,2,3-cd)pyrene	mg/kg	0.11	< 0.03	< 0.03	0.03	< 0.03	< 0.03	0.06	0.037	< 0.010	#	#	-	-	0.385
Naphthalene	mg/kg	< 0.1	< 0.1	< 0.1	< 0.1	< 0.1	< 0.1	< 0.1	<0.010	<0.010	-	-	58 ^(v) / 70 ^(v)	(190) (4,v) / (230) (4,v)	0.029
Phenanthrene	mg/kg	0.13	< 0.03	0.07	0.05	< 0.03	< 0.03	0.06	0.015	<0.010	-	-	-	-	0.703
Pyrene	mg/kg	0.33	< 0.03	0.13	0.11	< 0.03	< 0.03	0.08	0.06	<0.020	-	-	(1,600) (4,p) / NA (2)	NA (2) / NA (2)	1.362
Benzo[a]pyrene TEQ ⁴	mg/kg	0.32	0.08	0.11	0.1	0.08	0.08	0.16	0.07	0.03	35	40	-	-	0.922
Analytica Lab Reference		22-42925-2	22-42925-3	22-41184-5	22-41184-6	22-41184-7	22-40179-5	22-40179-6	22-42925-1				Guidelines for Hydrocarbon		

Analytica Lab Reference		22-42925-2	22-42925-3	22-41184-5	22-41184-6	22-41184-7	22-40179-5	22-40179-6	22-42925-1				Guidelines for Hydrocarbon	
Eurofins Lab Reference		-	-	-	-	-	K22-No0015979	K22-No0015980	-	NESCS SCSs for	NESCS SCSs for Recreational	Guidelines for Hydrocarbon	Contamination -	Background Soil
Sample Depth	m bgl	0.2	1.2	0.5	1.0	2.0	0.5	1.5	2.5	Commercial / Industrial	land use ¹	Contamination - Residential Use -	Commercial/Industrial Use - All	Concentrations, New Zealan
Test Pit		BH-	103		BH-104			BH-107		land use ¹	iand use	All Pathways -	Pathways -	3
Sample Name		12547621-BH-103(0.2)	12547621-BH-103(1.2)	12547621-BH-104(0.5)	12547621-BH-104(1.0)	12547621-BH-104(2.0)	12547621-BH-107(0.5)	12547621-BH-107(1.5)	12547621-BH-107(2.5)		((l)	Sand (<1m / 1-4m)	Sand (<1m / 1-4m)	
Soil Type		Topsoil/ organic silt	Fill/ organic silt	Fill/ clayey gravel	Fill/ silty clay	Fill/ waste material	Clayey silt/ waste	Gravelly silt/ bricks	Silt/ waste	(mg/kg)	(mg/kg)	(mg/kg) ²	(mg/kg) ²	(mg/kg)
Sample Date		01/11/2022	01/11/2022	09/11/2022	09/11/2022	09/11/2022	01/11/2022	01/11/2022	01/11/2022				(mg/kg)	
Acenaphthene	mg/kg	< 0.010	<0.010	12	0.01	0.019	< 0.03	0.61	0.033	-	-	-	-	0.055
Acenaphthylene	mg/kg	0.017	0.028	53	0.011	0.027	< 0.03	0.43	0.067	-	-	-	-	0.069
Anthracene	mg/kg	0.02	0.02	93	0.015	0.12	0.07	3.1	0.2	-	-	-	-	0.113
Benz[a]anthracene	mg/kg	0.092	0.14	320	0.039	0.28	0.48	5.4	0.53	#	#	-	-	0.47
Benzo[a]pyrene	mg/kg	0.12	0.23	290	0.058	0.32	0.58	4.5	0.74	#	#	-	-	0.595
Benzo[b]&[j] fluoranthene	mg/kg	0.12	0.27	500	0.058	0.33	0.3	4	0.73	#	#	-	-	0.947
Benzo[g,h,i]perylene	mg/kg	0.064	0.13	170	0.026	0.14	0.25	0.67	0.26	-	-	-	-	0.459
Benzo[k]fluoranthene	mg/kg	0.044	0.091	220	0.028	0.17	0.26	3.2	0.26	#	#	-	-	0.296
Chrysene	mg/kg	0.095	0.2	320	0.059	0.41	0.54	4.4	0.52	#	#	-	-	0.539
Dibenz(a,h)anthracene	mg/kg	0.012	0.023	50	< 0.010	0.032	0.07	0.35	0.061	#	#	-	-	0.112
Fluoranthene	mg/kg	0.17	0.31	420	0.076	0.65	0.67	13	1.5	#	#	-	-	1.345
Fluorene	mg/kg	< 0.010	<0.010	23	0.015	0.05	< 0.03	1.2	0.083	-	-	-	-	0.06
Indeno(1,2,3-cd)pyrene	mg/kg	0.082	0.17	220	0.029	0.16	0.23	1	0.35	#	#	-	-	0.385
Naphthalene	mg/kg	<0.010	< 0.010	2.3	0.083	0.19	< 0.1	0.5	0.053	-	-	58 ^(v) / 70 ^(v)	(190) (4,v) / (230) (4,v)	0.029
Phenanthrene	mg/kg	0.057	0.083	220	0.032	0.43	0.1	8.7	0.58	-	-	-	-	0.703
Pyrene	mg/kg	0.17	0.32	380	0.083	0.63	0.81	13	1.4	-	-	(1,600) ^(4,p) / NA ⁽²⁾	NA ⁽²⁾ / NA ⁽²⁾	1.362
Benzo[a]pyrene TEQ 4	mg/kg	0.17	0.32	470	0.08	0.46	0.79	6.4	1	35	40	<u> </u>	-	0.922

Lab Reference		22-41184-1	22-41184-2	22-41184-4	22-41184-3	22-40820-1	22-40820-2	22-40820-3			Guidelines for Hydrocarbon	Guidelines for Hydrocarbon	
Sample Depth	m bgl	0.5	1.9	1.9	3.3	0.5	1.5	3.2	NESCS SCSs for	NESCS SCSs for	Contamination Posidontial	Contamination -	Background Soil Concentrations,
Test Pit			BH-	108			BH-111		Commercial / Industrial	Recreational land use 1	Use - All Pathways -	Commercial/Industrial Use - All	New Zealand ³
Sample Name		12547621-BH-108(0.5)	12547621-BH-108(1.9)	12547621-DUP-2*	12547621-BH-108(3.3)	12547621-BH-111(0.5)	12547621-BH-111(1.5)	12547621-BH-111(3.2)	land use ¹		Sand (<1m / 1-4m)	Pathways -	
Soil Type		Fill/ clay	Fill/ clay	Fill/ clay	Fill/ waste/ clay	Fill/ gravel/ waste	Fill/waste/organics	Fill/ silt	(mg/kg)	(mg/kg)		Sand (<1m / 1-4m)	(mg/kg)
Sample Date		09/11/2022	09/11/2022	09/11/2022	09/11/2022	07/11/2022	07/11/2022	07/11/2022			(mg/kg) ²	(mg/kg) ²	
Acenaphthene	mg/kg	<0.010	< 0.010	0.026	0.033	< 0.010	0.038	<0.010	-	-	-	-	0.055
Acenaphthylene	mg/kg	<0.010	0.031	0.14	0.055	< 0.010	0.091	0.012	-	-	-	•	0.069
Anthracene	mg/kg	<0.010	0.054	1.5	0.16	< 0.010	0.19	0.019	-	-	-	•	0.113
Benz[a]anthracene	mg/kg	<0.020	0.22	4.6	0.73	<0.020	0.41	0.076	#	#	-	-	0.47
Benzo[a]pyrene	mg/kg	<0.010	0.35	2.7	0.75	< 0.010	0.66	0.1	#	#	-	•	0.595
Benzo[b]&[j] fluoranthene	mg/kg	<0.020	0.37	4	1.1	<0.020	0.78	0.11	#	#	-	-	0.947
Benzo[g,h,i]perylene	mg/kg	<0.020	0.21	1.2	0.43	<0.020	0.34	0.063	-	-	-	i	0.459
Benzo[k]fluoranthene	mg/kg	<0.010	0.18	1.8	0.55	< 0.010	0.29	0.038	#	#	-	•	0.296
Chrysene	mg/kg	<0.010	0.38	3.7	0.92	< 0.010	0.42	0.077	#	#	-	i	0.539
Dibenz(a,h)anthracene	mg/kg	<0.010	0.04	0.43	0.12	< 0.010	0.086	0.021	#	#	-	٠	0.112
Fluoranthene	mg/kg	<0.020	0.42	5.4	1.1	<0.020	0.63	0.18	#	#	-	-	1.345
Fluorene	mg/kg	<0.010	0.017	0.3	0.082	< 0.010	0.097	<0.010	-	-	-	٠	0.06
ndeno(1,2,3-cd)pyrene	mg/kg	<0.010	0.21	1.5	0.48	<0.010	0.48	0.078	#	#	-	•	0.385
Naphthalene	mg/kg	<0.010	<0.010	0.017	0.077	< 0.010	0.28	<0.010	-	-	58 ^(v) / 70 ^(v)	(190) (4,v) / (230) (4,v)	0.029
Phenanthrene	mg/kg	<0.010	0.15	4.8	0.58	< 0.010	0.66	0.066	-	-	-	-	0.703
Pyrene	mg/kg	<0.020	0.45	4.9	1.2	<0.020	0.52	0.19	-	-	(1,600) ^(4,p) / NA ⁽²⁾	NA (2) / NA (2)	1.362
Benzo[a]pyrene TEQ ⁴	mg/kg	0.03	0.49	4.4	1.2	0.03	0.96	0.16	35	40	-	-	0.922

Notes:
Values shaded grey exceed the NESCS SCSs for Commercial / Industrial land use
Value shaded yellow exceed the NESCS SCS for Recreational land use
Value shaded peach exceed the MIE Hydrocarbon Guidelines for Residential land use
Value shaded green exceed the MIE Hydrocarbon Guidelines for Commercial / Industrial land use
Red text exceed the ECan Background Concentrations of PAHs in Christchurch Urban Soils guideline values

* - Field duplicate of 12547621-BH-108(1.9)

Values shaded according according to the highest exceedance

A hyphen (-) indicates criterion less than the laboratory limit of reporting (LOR)

indicates criteria for these compounds are addressed using the Benzo(a)pyrene equivalence calculations provided in the NES SCS (refer to Note 1 & Note 6) Limiting pathways: (v) = volatilisation, (p) = Produce

(2) = Brackets denote values exceed threshold likely to correspond to formation of residual separate phase hydrocarbons

NA - indicates contaminant not limiting as estimated health-based criterion is significantly higher than that likely to be encountered on site.

- References
 1. Ministry for the Environment (2011). Resource Management (National Environmental Standard for Assessing and Managing Contaminants in Soil to Protect Human Health) Regulations 2011 (NESCS). Soil Contaminant Standards (SCS
 2. Ministry for the Environment (2011) Guidelines for Assessing and Managing Petroleum Hydrocarbon Contaminated Sites in New Zealand. Module 4 Tier 1 Soil screening criteria. Table 4.10. Residential land use and Table 4.11 Commercial / Industrial land use.
 3. Environment Canterbury (2007). Background Concentrations of polycyclic aromatic hydrocarbons in Christchurch urban soils. Report No. R07/19. Table 9.
 4. Benzo(a) pyrene equivalent concentration calculated as the sum of the carcinogenic PAHs in accordance with the methodology published in the NESCS.

Site Name: Green Island Landfill Table 3: Semi volatile organic compounds (SVOC) Analytical Results November 2022 Site Investigation



Analytica Lab Reference		22-40179-12	22-40179-13	22-40179-1	22-40179-2	22-40179-14	22-40179-15	22-40179-16					
Eurofins Lab reference		K22-No0015984	K22-No0015985	K22-No0015975	K22-No0015976	K22-No0015986	K22-No0015987	K22-No0015988				Codelines for the decrease	
Sample Depth	m bgl	1.0	2.7	0.5	1.5	0.6	1.5	2.5			Guidelines for Hydrocarbon	Guidelines for Hydrocarbon Contamination -	
Borehole		BH-	100	BH-	101		BH-102		NESCS SCSs ¹ for	NESCS SCSs 1 for	Contamination - Residential	Commercial/Industrial Use - All	Background Soil Concentrations, New Zealan
Sample Name		12547621-BH-100(1.0)	12547621-BH-100(2.7)	12547621-BH-101(0.5)	12547621-BH-101(1.5)	12547621-BH-102(0.6)	12547621-BH-102 1.2	12547621-BH-102 2.5	Commercial/Industrial land	Recreational land use	Use - All Pathways -	Pathways -	4,5,6
Soil Type		Fill/ sand/ waste	Fill/ wood in sand	Fill/ loam/waste	Peat/sandy silt	Fill/ sand/ waste			use (mg/kg)	(mg/kg)	Sand (<1m / 1-4m)	Sand (<1m / 1-4m)	(mg/kg)
Sample Date		01/11/2022	01/11/2022	01/11/2022	01/11/2022	01/11/2022	0.1/11/2022	1/11/2022	(mg/kg)		(mg/kg) ²	(mg/kg) ²	(8/8)
Semi volatile organic compounds (SVOCs)													
4.4'-DDD	mg/kg	< 0.01	< 0.01	< 0.01	< 0.01	< 0.01	<0.01	<0.01	•	-	-	-	0.00471
4.4'-DDE	mg/kg	< 0.01	< 0.01	< 0.01	< 0.01	< 0.01	<0.01	<0.01	-	-			0.0229
4.4'-DDT	mg/kg	< 0.01	< 0.01	< 0.01	< 0.01	< 0.01	<0.01	<0.01	1,000	400			0.0236
Acenaphthylene	mg/kg	0.04	< 0.03	< 0.03	< 0.03	< 0.03	<0.03	<0.03	-	-	-	-	0.069
Anthracene	mg/kg	0.04	< 0.03	< 0.03	< 0.03	< 0.03	<0.03	<0.03	-	-	-	-	0.113
Benz(a)anthracene	mg/kg	0.17	< 0.03	0.07	0.06	0.1	0.19	<0.03	-	-	-	-	0.47
Benzo(a)pyrene	mg/kg	0.23	< 0.03	0.06	0.05	0.1	0.30	<0.03	-	-	-	-	0.595
Benzo(a)pyrene TEQ (lower bound)* ³	mg/kg	0.29	< 0.03	0.07	0.07	0.13	0.46	<0.03	35	40	-	-	0.922
Benzo(a)pyrene TEQ (medium bound)* 3	mg/kg	0.31	0.04	0.09	0.09	0.15	0.46	<0.03	35	40	-	-	0.922
Benzo(a)pyrene TEQ (upper bound)* 3	mg/kg	0.32	0.08	0.11	0.1	0.16	0.46	0.08	35	40	-	-	0.922
Benzo(b&j)fluoranthene	mg/kg	0.13	< 0.03	0.04	0.06	0.08	0.27	<0.03	-	-	-	-	0.947
Benzo(g.h.i)perylene	mg/kg	0.08	< 0.03	< 0.03	0.06	0.06	0.25	<0.03	-	-	-	-	0.459
Benzo(k)fluoranthene	mg/kg	0.14	< 0.03	< 0.03	0.06	0.09	0.26	<0.03	-	-	-	-	0.296
Chrysene	mg/kg	0.18	< 0.03	0.09	0.09	0.1	0.27	<0.03	-	-	-	-	0.539
Dibenz(a.h)anthracene	mg/kg	< 0.03	< 0.03	< 0.03	< 0.03	< 0.03	0.06	<0.03	-	-	-	-	0.112
Fluoranthene	mg/kg	0.34	< 0.03	0.11	0.11	0.08	0.30	<0.03	-	-	-	·	1.345
ndeno(1.2.3-cd)pyrene	mg/kg	0.11	< 0.03	< 0.03	0.03	0.06	0.18	<0.03	-	-	-	-	0.385
Naphthalene	mg/kg	< 0.1	< 0.1	< 0.1	< 0.1	< 0.1	<0.1	<0.1	-	-	58 ^(v) / 70 ^(v)	(190) ^(4,v) / (230) ^(4,v)	0.029
Phenanthrene	mg/kg	0.13	< 0.03	0.07	0.05	0.06	0.13	<0.03	-	-	-	-	0.703
Pyrene	mg/kg	0.33	< 0.03	0.13	0.11	0.08	0.35	<0.03	-	-	(1,600) (4,p) / NA (2)	NA ⁽²⁾ / NA ⁽²⁾	1.362

Analytica Lab Reference		22-40179-8	22-40179-9	22-40179-6	22-40179-7	22-40820-2	22-40820-3			Guidelines for		
Eurofins Lab reference		K22-No0015982	K22-No0015983	K22-No0015980	K22-No0015981	-	-			Hydrocarbon	Guidelines for Hydrocarbon	
Sample Depth	m bgl	0.2	1.2	1.5	2.5	1.5	3.2	NESCS SCSs ¹ for	NESCS SCSs 1 for	Contamination -	Contamination -	Background Soil Concentrations, New
Borehole		ВН	-103	BH-	107	BH-:	111	Commercial/Industrial land	Recreational land use	Residential Use - All	Commercial/Industrial Use -	Zealand ^{4,5,6}
Sample Name		12547621-BH-103 0.2	12547621-BH-103 1.2	12547621-BH-107(1.5)	12547621-BH-107 2.5	12547621-BH-111(1.5)	12547621-BH-111(3.2)	use	(mg/kg)	Pathways -	All Pathways - Sand (<1m / 1-4m)	(mg/kg)
Soil Type				Gravelly silt/ bricks		Fill/waste/organics	Fill/ silt	(mg/kg)		Sand (<1m / 1-4m)	` _ `	
Sample Date		0.1/11/2022	1/11/2022	01/11/2022	1/11/2022	07/11/2022	07/11/2022			(mg/kg) ²	(mg/kg) ²	
Semi volatile organic compounds (SVOCs)												
4.4'-DDD	mg/kg	0.04	0.04	0.01	0.01	0.38	<0.30	-	-	-	-	0.00471
4.4'-DDE	mg/kg	0.04	<0.01	< 0.01	<0.01	<0.30	<0.30	-	-			0.0229
4.4'-DDT	mg/kg	0.13	0.06	< 0.01	<0.01	<0.50	<0.50	1,000	400			0.0236
Acenaphthene	mg/kg	< 0.03	<0.03	0.61	<0.03	<0.10	<0.10	-	-	-	-	0.055
Acenaphthylene	mg/kg	< 0.03	< 0.03	0.43	<0.03	0.14	<0.10	-	-	-	-	0.069
Anthracene	mg/kg	<0.03	<0.03	3.1	0.12	0.18	<0.10	-	=	-	-	0.113
Benz(a)anthracene	mg/kg	0.08	0.12	5.4	0.22	0.35	<0.10	-	-	-	-	0.47
Benzo(a)pyrene	mg/kg	0.09	0.20	4.5	0.28	0.62	0.24	-	-	-	-	0.595
Benzo(a)pyrene TEQ (lower bound)* 3	mg/kg	0.12	0.25	6.4	0.36	1.0	0.30	35	40	-	-	0.922
Benzo(a)pyrene TEQ (medium bound)* 3	mg/kg	0.13	0.27	6.4	0.37	-	-	35	40	-	-	0.922
Benzo(a)pyrene TEQ (upper bound)* 3	mg/kg	0.15	0.28	6.4	0.39	1.0	0.40	35	40	-	-	0.922
Benzo(b&j)fluoranthene	mg/kg	0.05	0.12	4.0	0.22	0.74	0.26	-	-	-	-	0.947
Benzo(g.h.i)perylene	mg/kg	0.08	0.12	0.67	0.13	0.36	0.15	-	-	-	-	0.459
Benzo(k)fluoranthene	mg/kg	0.06	0.14	3.2	0.15	0.31	0.13	-	-	-	-	0.296
Bis(2-ethylhexyl)phthalate	mg/kg	<0.5	<0.5	< 0.5	<0.5	0.50	<0.50	-	-	-	-	-
Chrysene	mg/kg	0.12	0.22	4.4	0.24	0.36	0.13	-	-	-	-	0.539
Dibenz(a.h)anthracene	mg/kg	<0.03	<0.03	0.35	<0.03	0.17	<0.10	-	-	-	-	0.112
Dibenzofuran	mg/kg	<0.5	<0.5	0.5	<0.5	<0.30	<0.030	-	•	-	-	-
Fluoranthene	mg/kg	0.14	0.30	13	0.51	0.96	0.33	-	-	-	-	1.345
Fluorene	mg/kg	<0.03	<0.03	1.2	0.05	0.13	<0.10	-	•	-	-	0.06
Indeno(1.2.3-cd)pyrene	mg/kg	0.05	0.10	1.0	0.12	0.48	0.19	-	-	-	-	0.385
Naphthalene	mg/kg	<0.1	<0.1	0.5	<0.1	0.12	<0.10	•	•	58 ^(v) / 70 ^(v)	(190) ^(4,v) / (230) ^(4,v)	0.029
Phenanthrene	mg/kg	0.04	0.12	8.7	0.31	0.58	0.14	-	-	-	-	0.703
Pyrene	mg/kg	0.13	0.34	13	0.53	0.87	0.3	-	-	(1,600) (4,p) / NA (2)	NA ⁽²⁾ / NA ⁽²⁾	1.362

Values shaded grey exceed the NESCS SCSs for Protection of Human Health (Commercial / Industrial land use) Values shaded yellow exceed the NESCS SCSs for Protection of Human Health (Recreational land use)

Value shaded peach exceed the MfE Hydrocarbon Guidelines for Residential land use

Value shaded green exceed the MfE Hydrocarbon Guidelines for Commercial / Industrial land use

Red text indicates values exceeds the adopted background concentration value

mg/kg - miligrams per kilogram m bgl - metres below ground level

A hyphen (-) indicates criterion not available for this analyte

< - reported at a concentration less than the laboratory limit of reporting (LOR)

- 1. Resource Management (National Environmental Standard for Assessing and Managing Contaminants in Soil to Protect Human Health) Regulations 2011 (NESCS). Soil contaminant standards (SCS) for Commercial / Industrial Land use
- 2. Ministry for the Environment (2011) Guidelines for Assessing and Managing Petroleum Hydrocarbon Contaminated Sites in New Zealand. Module 4 Tier 1 Soil screening criteria. Table 4.10. Residential land use and Table 4.11 Commercial / Industrial land use.
- 3. Benzo(a) pyrene equivalent concentration calculated as the sum of the carcinogenic PAHs in accordance with the methodology published in the NESCS.
- 3. US EPA Superfund EGV (June, 2013)
- 4. Ministry for the Environment (1998). Ambient concentrations of selected organochlorines in soils. Table F3 mean values.
- 5. Environment Canterbury (2007). Background Concentrations of polycyclic aromatic hydrocarbons in Christchurch urban soils. Report No. R07/19. Table 9.
 6. Landcare Research Limited (2015) Background soil concentrations of selected trace elements and organic contaminants in New Zealand. Table 21. Christchurch Soils.



Green Island Landfill Soil Analytical Results November 2022

Table 4: Total Petroleum Hydrocarbons (TPH)

Sample ID		22-40820-2		
Sample Depth	m bgl	1.5	Cuidalinas fan Hudusasukan	Cuidelines for Undersonbor
Borehole		BH-111	Guidelines for Hydrocarbon Contamination -Commercial / Industrial -	Guidelines for Hydrocarbon Contamination - Residential - All
Sample Name		12547621-BH-111(1.5)	•	
Soil Type		Fill/waste/organics	All Pathways - Sand (1-4m)	Pathways - Sand (1-4m)
Sample Date		07/11/2022	(mg/kg) ¹	(mg/kg) ¹
Total Petroleum Hy	drocarbons (TPH)			
C7-C9	mg/kg	<10	120 ^(m)	120 ^(m)
C10-C14	mg/kg	40	(1,900) ^(2,x)	(560) ^(2,x)
C15-C36	mg/kg	681	NA	NA
C7-C36 (Total)	mg/kg	721	-	-

Notes:

All units are in mg/kg

m bgl - metres below ground level

A hyphen (-) indicates criterion not available

< - reported at a concentration less than the laboratory limit of reporting (LOR)

Limiting pathways: (m) = Maintenance/Excavation, (2) = Brackets denote values exceed threshold likely to correspond to formation of residual separate phasehydrocarbons, (x) = PAH surrogate NA - indicates estimated criterion exceeds 20,000 mg/kg. At 20,000 mg/kg residual separate phase is expected to have formed in soil matrix. Some aesthetic impact may be noted.

References

1. Ministry for the Environment (2011) Guidelines for Assessing and Managing Petroleum Hydrocarbon Contaminated Sites in New Zealand. Module 4 - Tier 1 Soil screening criteria. Table 4.14. Commercial / Industrial land use.



Green Island Landfill Soil Analytical Results November 2022 Table 5: Ammonia in Soil

Lab Reference		22-40179	22-40179	22-40179	22-40179	22-40179	22-40179	22-40179	22-40179
Sample Depth	m bgl	1.0	2.7	0.5	1.5	2.5	1.2	3.5	0.5
Borehole		BH-10	0		BH-101		BH-	103	BH-107
Soil type		Fill/ sand/ waste	Fill/ wood in sand	Fill/ loam/waste	Peat/sandy silt	Silty sand	Fill/ organic silt	Clayey silt	Clayey silt/ waste
Sample Name		12547621-BH-100	12547621-BH-100(2.7)	12547621-BH-101(0.5)	12547621-BH-101(1.5)	12547621-BH-101(2.5)	12547621-BH-103	12547621-BH-103	12547621-BH-107(0.5)
Sample Date		01/11/2022	01/11/2022	01/11/2022	01/11/2022	01/11/2022	01/11/2022	01/11/2022	01/11/2022
Ammonia									
Ammonia	mg/kg	<5.00	5.27	<5.00	43.6	7.24	17.4	<5.00	<5.00

Notes:
All units are in mg/kg
m bgl - metres below ground level

GHD Project Number: 12547621 Page 1 of 1



Green Island Landfill Soil Analytical Results November 2022 Table 6: Asbestos in Soil

Laboratory ID		22-40290-10	22-40290-11	22-40290-1	22-40290-12	22-40290-13	22-40290-8	22-40968-5	
Borehole		BH	100	BH-101	BH-	102	BH-103	BH-104	BRANZ 1 - Guideline Values
GHD Sample Reference		12547621-BH-100(1.0)	12547621-BH-100(2.7)	12547621-BH-101(0.5)	12547621-BH-102(0.6)	12547621-BH-102(1.5)	12547621-BH-103(0.2)	12547621-BH-104 (0.5)	Commercial / Industrial
Sample Date		01/11/2022	01/11/2022	01/11/2022	01/11/2022	01/11/2022	01/11/2022	09/11/2022	
Depth (m bgl)	1	1.0	2.7	0.5	0.6	1.5	0.2	0.5	% w/w
oilType		Fill/ sand/ waste	Fill/ wood in sand	Fill/ loam/waste	Fill/ sand/ waste	Fill/ peat/ waste	Topsoil/ organic silt	Fill/ clayey gravel	
	>10mm fraction	-	-	-	-	-	-	-	-
Asbestos weight (g)	2-10mm fraction	-	-	-	-	-	-	-	-
	<2mm fraction	-	-	-	-	-	-	-	-
)4/ // 4/ h 4 0/	ACM	Absent	Absent	Absent	Absent	Absent	Absent	Absent	0.05
W/W asbestos %	AF/FA	Absent	Absent	Absent	Absent	Absent	Absent	Absent	0.001
					•				
boratory ID		22-40968-6	22-40968-7	22-40290-5	22-40968-1*	22-40290-6	22-40290-7	-	
	1	BH	104		BH-1	07			BRANZ 1 - Guideline Values
HD Sample Reference	1	12547621-BH-104 (1.0)	12547621-BH-104 (2.0)	12547621-BH-107(0.5)	12547621-BH-107 (0.5)	12547621-BH-107(1.5)	12547621-BH-107(2.5)	-	Commercial / Industrial
mple Date	1	09/11/2022	09/11/2022	01/11/2022	09/11/2022	01/11/2022	01/11/2022	-	·
epth (m bgl)	1	1.0	2.0	0.5	0.5	1.5	2,5	-	% w/w
oil Type	1	Fill/ silty clay	Fill/ waste material	Clayey silt/ waste	Clayey silt/ waste	Gravelly silt/ bricks	Silt/ waste	-	
	>10mm fraction	- '	-	-	None detected	-			-
Asbestos weight (g)	2-10mm fraction	-		-	None detected	-	-	-	-
	<2mm fraction	-	-		None detected		-	-	-
M/M/	ACM	Absent	Absent	Absent	Absent	Absent	Absent	-	0.05
W/W asbestos %	AF/FA	Absent	Absent	Present	Absent	Absent	Absent	-	0.001
boratory ID		22-40968-2	22-40968-3	22-40968-4	22-40546-1	22-40546-2	22-40546-3	-	
			BH-108			BH-111			BRANZ 1 - Guideline Values
HD Sample Reference	1	12547621-BH-108 (0.5)	12547621-BH-108 (1.9)	12547621-BH-108 (3.3)	12547621 - BH - 111 (0.5)	12547621 - BH - 111 (1.5)	12547621 - BH - 111 (3.2)	-	Commercial / Industrial
imple Date	1	09/11/2022	09/11/2022	09/11/2022	07/11/2022	07/11/2022	07/11/2022	-	
epth (m bgl)	1	0.5	1.9	3.3	0.5	1.5	3.2	-	% w/w
oil Type	1	Fill/ clay	Fill/ clay	Fill/ waste/ clay	Fill/ gravel/ waste	Fill/waste/organics	Fill/ silt	-	
	>10mm fraction	-	-	-	-	-	-	-	-
Asbestos weight (g)	2-10mm fraction	-	-	-	-	-	-	-	-
	<2mm fraction	-	-	-	-	-	-	-	-
14/04/bb 0/	ACM	Absent	Absent	Absent	Absent	Absent	Absent	-	0.05
W/W asbestos %	AF/FA	Absent	Absent	Absent	Absent	Absent	Absent	-	0.001

Notes:

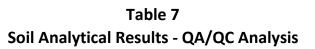
Peach shading indicates that asbestos was found present but at concentrations less than the guideline value Blue shading indicates asbestos present at concentrations above the adopted BRANZ guideline values

Reference

1. BRANZ (2017) New Zealand Guidelines for Assessing and Managing Asbestos in Soil. Table 5 - Commercial / Industrial guideline values

GHD project Number: 12547621 Page 1 of 1

^{*} This sample was a retest of the sample collected at the same location and depth on 1 November 2022





QA/QC TYPE		FIELD DUP		
Sample Name		12547621-BH-108(1.9)	2547621-BH-108(1.9) 12547621-DUP-2	
Sample Date		9/11/2	RPD	
Soil type		Fill / Clay	Fill / Clay	
Laboratory Number		22-41184-2	22-41184-4	
Depth (mbgl)		1.9		
Heavy Metals				
Arsenic	mg/kg	7.2	7.1	-1.4
Beryllium	mg/kg	1.42	1.33	-6.5
Boron	mg/kg	7.6	7.9	3.9
Cadmium	mg/kg	0.16	0.16	0.0
Chromium	mg/kg	50.1	49.4	-1.4
Copper	mg/kg	20.9	20.8	-0.5
Lead	mg/kg	52.9	42.8	-21.1
Mercury	mg/kg	0.28	0.27	-3.6
Nickel	mg/kg	38	35.9	-5.7
Zinc	mg/kg	124	121	-2.4

QA/QC TYPE		FIELD DUP	PLICATE		
Sample Name		12547621-BH-108(1.9)	12547621-DUP-2		
Sample Date		9/11/2	9/11/2022		
Laboratory Number		22-41184-2	22-41184-4		
Depth (mbgl)		1.9			
Acenaphthene	mg/kg	<0.010	0.026	-	
Acenaphthylene	mg/kg	0.031	0.14	127.5	
Anthracene	mg/kg	0.054	1.5	186.1	
Benz[a]anthracene	mg/kg	0.22	4.6	181.7	
Benzo[a]pyrene	mg/kg	0.35	2.7	154.1	
Benzo[b]&[j] fluoranthene	mg/kg	0.37	4	166.1	
Benzo[g,h,i]perylene	mg/kg	0.21	1.2	140.4	
Benzo[k]fluoranthene	mg/kg	0.18	1.8	163.6	
Chrysene	mg/kg	0.38	3.7	162.7	
Dibenz(a,h)anthracene	mg/kg	0.04	0.43	166.0	
Fluoranthene	mg/kg	0.42	5.4	171.1	
Fluorene	mg/kg	0.017	0.3	178.5	
Indeno(1,2,3-cd)pyrene	mg/kg	0.21	1.5	150.9	
Naphthalene	mg/kg	<0.010	0.017	-	
Phenanthrene	mg/kg	0.15	4.8	187.9	
Pyrene	mg/kg	0.45	4.9	166.4	
Benzo[a]pyrene TEQ ³	mg/kg	0.49	4.4	159.9	

QA/QC TYPE		FIELD D	UPLICATE	
Sample Name		12547621-BH- 111(3.2)	12547621-DUP	
Sample Date		7/11	1/2022	RPD
Soil type		Fill / Silt	Fill / Silt	
Laboratory Number		22-40820-3	22-40820-04	
Depth (mbgl)			3.2	
Heavy Metals				
Arsenic	mg/kg	9	9.6	6.5
Beryllium	mg/kg	1.35	1.75	25.8
Boron	mg/kg	14	15	6.9
Cadmium	mg/kg	0.18	0.32	56.0
Chromium	mg/kg	52.7	58.3	10.1
Copper	mg/kg	25.2	34.3	30.6
Lead	mg/kg	32.9	62.2	61.6
Mercury	mg/kg	0.12	0.21	54.5
Nickel	mg/kg	45.4	49	7.6
Zinc	mg/kg	179	236	27.5

Attachment D Laboratory Reports



Analytica Laboratories Limited Ruakura Research Centre 10 Bisley Road Hamilton sales@analytica.co.nz www.analytica.co.nz

Certificate of Analysis

GHD Ltd

Level 1, Bing Harris Building, 286 Princess Street, Dunedin

Dunedin 9016

Attention: Cecilia Gately Phone: 021 973 994

Email: danny.fitzgerald@ghd.com

Sampling Site: Green Island Landfill

Lab Reference: 22-40179

Submitted by: Danny Fitzgerald

Date Received: 04/11/2022
Testing Initiated: 4/11/2022
Date Completed: 22/11/2022

Order Number:

Reference: 12547621

Report Comments

Samples were collected by yourselves (or your agent) and analysed as received at Analytica Laboratories. Samples were in acceptable condition unless otherwise noted on this report.

Specific testing dates are available on request.

9 Heavy Metals in Soil

Client Sample ID		12547621-BH-101 0.5	12547621-BH-101 1.5	12547621-BH-101 2.5	12547621-BH-101 3.5	12547621- BH-107 0.5	
	Da	te Sampled	01/11/2022	01/11/2022	01/11/2022	01/11/2022	01/11/2022
Analyte	Unit	Reporting Limit	22-40179-1	22-40179-2	22-40179-3	22-40179-4	22-40179-5
Arsenic	mg/kg dry wt	0.125	9.7	12	7.1	5.5	5.0
Boron	mg/kg dry wt	1.25	6.9	27	8.4	11	5.5
Cadmium	mg/kg dry wt	0.005	0.30	0.18	0.018	0.029	0.052
Chromium	mg/kg dry wt	0.125	29.3	55.2	21.8	18.8	18.4
Copper	mg/kg dry wt	0.075	14.2	18.2	4.1	4.0	8.14
Lead	mg/kg dry wt	0.25	60.1	50.8	4.9	4.2	19.7
Mercury	mg/kg dry wt	0.025	0.079	0.14	<0.025	<0.025	0.13
Nickel	mg/kg dry wt	0.05	36.6	14.8	10.6	9.53	10.2
Zinc	mg/kg dry wt	0.05	183	119	36.7	29.0	37.0

9 Heavy Metals in Soil

Client Sample ID			12547621-BH-107 1.5	12547621-BH-107 2.5	12547621-BH-103 0.2	12547621-BH-103 1.2	12547621- BH-103 2.5
Date Sampled		01/11/2022	01/11/2022	01/11/2022	01/11/2022	01/11/2022	
Analyte	Unit	Reporting Limit	22-40179-6	22-40179-7	22-40179-8	22-40179-9	22-40179-10
Arsenic	mg/kg dry wt	0.125	4.8	4.6	6.8	25.8	8.7
Boron	mg/kg dry wt	1.25	14	7.2	3.4	20	7.1
Cadmium	mg/kg dry wt	0.005	0.060	0.11	0.29	0.28	0.12
Chromium	mg/kg dry wt	0.125	39.2	24.2	20.1	156	19.8
Copper	mg/kg dry wt	0.075	17.7	16.2	20.3	31.6	6.4

All tests reported herein have been performed in accordance with the laboratory's scope of accreditation with the exception of tests marked *, which are not accredited.





9 Heavy Metals in Soil

Client Sample ID		12547621-BH-107 1.5	12547621-BH-107 2.5	12547621-BH-103 0.2	12547621-BH-103 1.2	12547621- BH-103 2.5	
	Da	te Sampled	01/11/2022	01/11/2022	01/11/2022	01/11/2022	01/11/2022
Lead	mg/kg dry wt	0.25	17.7	32.1	77.8	74.3	5.45
Mercury	mg/kg dry wt	0.025	0.091	0.097	0.10	0.30	0.029
Nickel	mg/kg dry wt	0.05	25.6	9.59	19.1	30.7	13.2
Zinc	mg/kg dry wt	0.05	67.7	53.0	120	210	42.7

9 Heavy Metals in Soil

Client Sample ID		12547621-BH-103 3.5	12547621-BH-100 1.0	12547621-BH-100 2.7	12547621-BH-102 0.6	12547621- BH-102 1.5	
	Da	te Sampled	01/11/2022	01/11/2022	01/11/2022	01/11/2022	01/11/2022
Analyte	Unit	Reporting Limit	22-40179-11	22-40179-12	22-40179-13	22-40179-14	22-40179-15
Arsenic	mg/kg dry wt	0.125	9.4	2.8	7.8	4.7	29.7
Boron	mg/kg dry wt	1.25	9.5	1.5	9.2	2.1	32
Cadmium	mg/kg dry wt	0.005	0.070	0.059	0.047	0.11	0.27
Chromium	mg/kg dry wt	0.125	19.1	21.6	28.0	46.2	125
Copper	mg/kg dry wt	0.075	6.7	18.5	6.6	27.3	34.2
Lead	mg/kg dry wt	0.25	6.78	14.8	5.01	44.8	98.5
Mercury	mg/kg dry wt	0.025	<0.025	0.055	<0.025	0.12	0.32
Nickel	mg/kg dry wt	0.05	12.2	14.1	15.5	35.9	33.8
Zinc	mg/kg dry wt	0.05	49.3	52.0	47.4	105	467

9 Heavy Metals in Soil

	Clien	12547621-BH-102 2.5	12547621-BH-102 3.5	
	Da	01/11/2022	01/11/2022	
Analyte	Unit	Reporting Limit	22-40179-16	22-40179-17
Arsenic	mg/kg dry wt	0.125	11	11
Boron	mg/kg dry wt	1.25	9.9	9.0
Cadmium	mg/kg dry wt	0.005	0.012	0.066
Chromium	mg/kg dry wt	0.125	23.7	22.5
Copper	mg/kg dry wt	0.075	6.5	5.5
Lead	mg/kg dry wt	0.25	6.24	4.7
Mercury	mg/kg dry wt	0.025	<0.025	<0.025
Nickel	mg/kg dry wt	0.05	11.5	15.1
Zinc	mg/kg dry wt	0.05	49.4	45.9

Soil Aggregate Properties and Nutrients

Client Sample ID		12547621-BH-101 0.5	12547621-BH-101 1.5	12547621-BH-101 2.5	12547621-BH-107 0.5	12547621- BH-103 1.2	
Date Sampled		01/11/2022	01/11/2022	01/11/2022	01/11/2022	01/11/2022	
Analyte Unit Reporting Limit		22-40179-1	22-40179-2	22-40179-3	22-40179-5	22-40179-9	
Ammonia-N*	mg/kg dry wt	5	<5.00	43.6	7.24	<5.00	17.4

Soil Aggregate Properties and Nutrients

	Client Sample ID			12547621-BH-100 1.0	12547621-BH-100 2.7
	Date Sampled			01/11/2022	01/11/2022
Analyte Unit Reporting Limit		22-40179-11	22-40179-12	22-40179-13	
Ammonia-N*	mg/kg dry wt	5	<5.00	<5.00	5.27

Custom Job

Client Sample ID		12547621-BH-101 0.5	12547621-BH-101 1.5	12547621-BH-101 2.5	12547621-BH-101 3.5	12547621- BH-107 0.5
Date Sampled		01/11/2022	01/11/2022	01/11/2022	01/11/2022	01/11/2022
Analyte Unit	Reporting Limit	22-40179-1	22-40179-2	22-40179-3	22-40179-4	22-40179-5
Custom Job		Complete	Complete	Complete	Complete	Complete

Custom Job

Client Sample ID		12547621-BH-107 1.5	12547621-BH-107 2.5	12547621-BH-103 0.2	12547621-BH-103 1.2	12547621- BH-100 1.0
Date Sampled		01/11/2022	01/11/2022	01/11/2022	01/11/2022	01/11/2022
Analyte Unit Reporting Limit		22-40179-6	22-40179-7	22-40179-8	22-40179-9	22-40179-12
Custom Job		Complete	Complete	Complete	Complete	Complete

Custom Job

Clien	Client Sample ID		12547621-BH-102 0.6	12547621-BH-102 1.5	12547621-BH-102 2.5
Date Sampled		01/11/2022	01/11/2022	01/11/2022	01/11/2022
Analyte Unit	Reporting Limit	22-40179-13	22-40179-14	22-40179-15	22-40179-16
Custom Job		Complete	Complete	Complete	Complete

Method Summary

Elements in Soil Samples dried and passed through a 2 mm sieve followed by acid digestion and analysis by ICP-

MS. In accordance with in-house procedure based on US EPA method 200.8.

Ammonia-N in Soil 1:5 water extraction (NEPM, Schedule B3, Laboratory Analysis of Potentially Contaminated Soil, 2011) followed by colour-metric analysis (APHA 4500 NH₃ - F Online edition - modified - Discrete

Analyser).

Jarred Wilson, DipSci

Sandra Mathews, B.Eng.

Trace Elements Team Leader Technologist Adam Ang

Team Leader

Sharelle Frank, B.Sc. (Tech)

Technologist



Analytica Laboratories Limited 34 Brisbane Street Svdenham Christchurch sales@analytica.co.nz www.analytica.co.nz

Certificate of Analysis

GHD Ltd

Level 1, Bing Harris Building, 286 Princess Street, Dunedin

Dunedin 9016

Attention: Cecilia Gately Phone: 021 973 994

Email: danny.fitzgerald@ghd.com

Description of Work: PA - 12547621

Green Island Landfill Sampling Site:

Report Comments

Samples were collected by yourselves (or your agent) and analysed as received at Analytica Laboratories. Samples were in acceptable condition unless otherwise noted on this report.

Lab Reference:

Submitted by:

Date Received:

Testing Initiated:

Order Number: Reference:

Date Completed: 8/11/2022

22-40290

04/11/2022

7/11/2022

12547621

Danny Fitzgerald

Specific testing dates are available on request.

Asbestos in Soil (Qualitative)

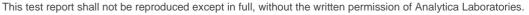
Sample Details

Campic Detai	10				
Laboratory ID	Client Sample ID	Sample Location	Sample Description	Date Sampled	Date Analysed
22-40290-1	12547621-BH-101 0.5		Soil	01/11/2022	07/11/2022
22-40290-5	12547621-BH-107 0.5		Soil	01/11/2022	07/11/2022
22-40290-6	12547621-BH-107 1.5		Soil	01/11/2022	07/11/2022
22-40290-7	12547621-BH-107 2.5		Soil	01/11/2022	08/11/2022
22-40290-8	12547621-BH-103 0.2		Soil	01/11/2022	08/11/2022
22-40290-10	12547621-BH-100 1.0		Soil	01/11/2022	08/11/2022
22-40290-11	12547621-BH-100 2.7		Soil	01/11/2022	08/11/2022
22-40290-12	12547621-BH-102 0.6		Soil	01/11/2022	08/11/2022
22-40290-13	12547621-BH-102 1.5		Soil	01/11/2022	08/11/2022

Information in the above table supplied by the client: Client Sample ID, Sample Location, Date Sampled.

Laboratory ID	Client Sample ID	Fibre Types	Trace Asbestos (Presence / Absence)	Asbestos (Presence / Absence)
	Units			
22-40290-1	12547621-BH-101 0.5	Asbestos NOT Detected. Organic Fibres	Absent	Absent
22-40290-5	12547621-BH-107 0.5	Chrysotile (White Asbestos) Organic Fibres	Absent	Present
22-40290-6	12547621-BH-107 1.5	Asbestos NOT Detected. Organic Fibres	Absent	Absent
22-40290-7	12547621-BH-107 2.5	Asbestos NOT Detected. Organic Fibres	Absent	Absent
22-40290-8	12547621-BH-103 0.2	Asbestos NOT Detected. Organic Fibres	Absent	Absent

All tests reported herein have been performed in accordance with the laboratory's scope of accreditation with the exception of tests marked *, which are not accredited.





Laboratory ID	Client Sample ID	Fibre Types	Trace Asbestos (Presence / Absence)	Asbestos (Presence / Absence)
	Units			
22-40290-10	12547621-BH-100 1.0	Asbestos NOT Detected. Organic Fibres	Absent	Absent
22-40290-11	12547621-BH-100 2.7	Asbestos NOT Detected. Organic Fibres	Absent	Absent
22-40290-12	12547621-BH-102 0.6	Asbestos NOT Detected. Organic Fibres	Absent	Absent
22-40290-13	12547621-BH-102 1.5	Asbestos NOT Detected. Organic Fibres	Absent	Absent

Information in the above table supplied by the client: Client Sample ID.

Asbestos in Soil (Qualitative) Approver:

Aleesha van Eeden, M.Sc. Technician

Method Summary

Asbestos Fibres in Soil (Qualitative)

Sample analysis was performed using polarised light microscopy with dispersion staining in accordance with AS4964-2004 Method for the qualitative identification of asbestos in bulk samples.

Note 1: The reporting limit for this analysis is 0.1g/kg (0.01%) by application of polarised light microscopy, dispersion staining and trace analysis techniques.

Note 2: Trace asbestos is indicative that freely liberated respirable fibres are present and dust control measures should be implemented or increased on site. This is not the sole indicator for the friable nature of the asbestos present.

Note 3: If mineral fibres of unknown type are detected, by PLM and dispersion staining, these may or may not be asbestos fibres. To confirm the identity of this fibre, another independent analytical technique such as XRD analysis is advised.

Note 4: The laboratory does not take responsibility for the sampling procedure or accuracy of sample location description.



Analytica Laboratories Limited 34 Brisbane Street Sydenham Christchurch sales@analytica.co.nz www.analytica.co.nz

Certificate of Analysis

GHD Ltd

Level 1, Bing Harris Building, 286 Princess Street, Dunedin

Dunedin 9016

Attention: Cecilia Gately Phone: 021973994

Email: danny.fitzgerald@ghd.com

Sampling Site: Green Island LF

Description of Work: Soil - Green Island LF

Lab Reference: 22-40546

Submitted by: Danny FitzGerald

Date Received: 07/11/2022 Testing Initiated: 7/11/2022 Date Completed: 9/11/2022

Order Number: N/A

Reference: 12547621

Report Comments

Samples were collected by yourselves (or your agent) and analysed as received at Analytica Laboratories. Samples were in acceptable condition unless otherwise noted on this report.

Specific testing dates are available on request.

AMENDED REPORT. This report replaces in full a previous version 22-40546[R00] sent on 09/11/2022. Changes have been made to the Client Sample ID's as per request of the client.

Asbestos in Soil (Qualitative)

Sample Details

Laboratory ID	Client Sample ID	Sample Location	Sample Description	Date Sampled	Date Analysed
22-40546-1	12547621 - BH - 111 (0.5)	N/A	Soil	07/11/2022	09/11/2022
22-40546-2	12547621 - BH - 111 (1.5)	N/A	Soil	07/11/2022	09/11/2022
22-40546-3	12547621 - BH - 111 (3.2)	N/A	Soil	07/11/2022	09/11/2022

Information in the above table supplied by the client: Client Sample ID, Sample Location, Date Sampled.

Laboratory ID	Client Sample ID	Fibre Types	Trace Asbestos (Presence / Absence)	Asbestos (Presence / Absence)
	Units			
22-40546-1	12547621 - BH - 111 (0.5)	Asbestos NOT Detected. Organic Fibres	Absent	Absent
22-40546-2	12547621 - BH - 111 (1.5)	Asbestos NOT Detected. Organic Fibres	Absent	Absent
22-40546-3	12547621 - BH - 111 (3.2)	Asbestos NOT Detected. Organic Fibres	Absent	Absent

Information in the above table supplied by the client: Client Sample ID.

Asbestos in Soil (Qualitative) Approver:

Aleesha van Eeden, M.Sc.

Technician

All tests reported herein have been performed in accordance with the laboratory's scope of accreditation with the exception of tests marked *, which are not accredited.

This test report shall not be reproduced except in full, without the written permission of Analytica Laboratories.



Method Summary

Asbestos Fibres in Soil (Qualitative)

Sample analysis was performed using polarised light microscopy with dispersion staining in accordance with AS4964-2004 Method for the qualitative identification of asbestos in bulk samples.

Note 1: The reporting limit for this analysis is 0.1g/kg (0.01%) by application of polarised light microscopy, dispersion staining and trace analysis techniques.

Note 2: Trace asbestos is indicative that freely liberated respirable fibres are present and dust control measures should be implemented or increased on site. This is not the sole indicator for the friable nature of the asbestos present.

Note 3: If mineral fibres of unknown type are detected, by PLM and dispersion staining, these may or may not be asbestos fibres. To confirm the identity of this fibre, another independent analytical technique such as XRD analysis is advised.

Note 4: The laboratory does not take responsibility for the sampling procedure or accuracy of sample location description.



Analytica Laboratories Limited Ruakura Research Centre 10 Bisley Road Hamilton sales@analytica.co.nz www.analytica.co.nz

Certificate of Analysis

GHD Ltd

Level 1, Bing Harris Building, 286 Princess Street, Dunedin

Dunedin 9016

Attention: Cecilia Gately Phone: 021 973 994

Email: danny.fitzgerald@ghd.com

Sampling Site:

Lab Reference: 22-40820

Submitted by: Danny FitzGerald

Date Received: 09/11/2022 Testing Initiated: 9/11/2022 Date Completed: 24/11/2022

Order Number:

Reference: 12547621

Report Comments

Samples were collected by yourselves (or your agent) and analysed as received at Analytica Laboratories. Samples were in acceptable condition unless otherwise noted on this report.

Specific testing dates are available on request.

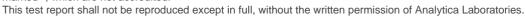
Heavy Metals in Soil

	Client Sample ID		12547621- BH-111(0.5) 0.5	12547621- BH-111(1.5) 1.5	12547621- BH-111(3.2) 3.2	12547621-Dup
	Da	te Sampled	07/11/2022	07/11/2022	07/11/2022	07/11/2022
Analyte	Unit	Reporting Limit	22-40820-1	22-40820-2	22-40820-3	22-40820-4
Arsenic	mg/kg dry wt	0.125	1.7	8.9	9.0	9.6
Beryllium	mg/kg dry wt	0.013	1.2	1.3	1.35	1.75
Boron	mg/kg dry wt	1.25	8.4	20	14	15
Cadmium	mg/kg dry wt	0.005	0.11	2.18	0.18	0.32
Chromium	mg/kg dry wt	0.125	90.5	34.6	52.7	58.3
Copper	mg/kg dry wt	0.075	59.1	40.3	25.2	34.3
Lead	mg/kg dry wt	0.25	49.1	179	32.9	62.2
Mercury	mg/kg dry wt	0.025	<0.025	0.12	0.12	0.21
Nickel	mg/kg dry wt	0.05	105	32.7	45.4	49.0
Zinc	mg/kg dry wt	0.05	127	403	179	236

Polycyclic Aromatic Hydrocarbons - Soil

	Client Sample ID			12547621- BH-111(1.5) 1.5	12547621- BH-111(3.2) 3.2
Date Sampled		07/11/2022	07/11/2022	07/11/2022	
Analyte	Unit	Reporting Limit	22-40820-1	22-40820-2	22-40820-3
1-Methylnaphthalene	mg/kg dry wt	0.01	<0.010	0.093	<0.010
2-Methylnaphthalene	mg/kg dry wt	0.01	<0.010	0.12	<0.010
Acenaphthene	mg/kg dry wt	0.01	<0.010	0.038	<0.010
Acenaphthylene	mg/kg dry wt	0.01	<0.010	0.091	0.012

All tests reported herein have been performed in accordance with the laboratory's scope of accreditation with the exception of tests marked *, which are not accredited.





Polycyclic Aromatic Hydrocarbons - Soil

	Client Sample ID		12547621- BH-111(0.5) 0.5	12547621- BH-111(1.5) 1.5	12547621- BH-111(3.2) 3.2
	Da	te Sampled	07/11/2022	07/11/2022	07/11/2022
Anthracene	mg/kg dry wt	0.01	<0.010	0.19	0.019
Benz[a]anthracene	mg/kg dry wt	0.02	<0.020	0.41	0.076
Benzo[a]pyrene	mg/kg dry wt	0.01	<0.010	0.66	0.10
Benzo[b]&[j] fluoranthene	mg/kg dry wt	0.02	<0.020	0.78	0.11
Benzo[g,h,i]perylene	mg/kg dry wt	0.02	<0.020	0.34	0.063
Benzo[k]fluoranthene	mg/kg dry wt	0.01	<0.010	0.29	0.038
Chrysene	mg/kg dry wt	0.01	<0.010	0.42	0.077
Dibenz(a,h)anthracene	mg/kg dry wt	0.01	<0.010	0.086	0.021
Fluoranthene	mg/kg dry wt	0.02	<0.020	0.63	0.18
Fluorene	mg/kg dry wt	0.01	<0.010	0.097	<0.010
Indeno(1,2,3-cd)pyrene	mg/kg dry wt	0.01	<0.010	0.48	0.078
Naphthalene	mg/kg dry wt	0.01	<0.010	0.28	<0.010
Phenanthrene	mg/kg dry wt	0.01	<0.010	0.66	0.066
Pyrene	mg/kg dry wt	0.02	<0.020	0.52	0.19
Benzo[a]pyrene TEQ (LOR)	mg/kg dry wt	0.03	0.030	0.96	0.16
Benzo[a]pyrene TEQ (Zero)	mg/kg dry wt	0.01	<0.010	0.96	0.16
Anthracene-d10 (Surrogate)	%	1	99	94	100

Total Petroleum Hydrocarbons - Soil

	Clien	12547621- BH-111(1.5) 1.5				
	Da	Date Sampled				
Analyte	Unit	Reporting Limit	22-40820-2			
C7-C9	mg/kg dry wt	10	<10			
C10-C14	mg/kg dry wt	15	40			
C15-C36	mg/kg dry wt	25	681			
C7-C36 (Total)	mg/kg dry wt	50	721			

Moisture Content

	Client Sample ID		12547621- BH-111(0.5) 0.5	12547621- BH-111(1.5) 1.5	12547621- BH-111(3.2) 3.2
	Date Sampled		07/11/2022	07/11/2022	07/11/2022
Analyte	Unit	Reporting Limit	22-40820-1	22-40820-2	22-40820-3
Moisture Content	%	1	14	49	24

Semivolatile Organic Compounds - Soil

	Clien	t Sample ID	12547621- BH-111(1.5) 1.5	12547621- BH-111(3.2) 3.2
	Da	te Sampled	07/11/2022	07/11/2022
Analyte	Unit	Reporting Limit	22-40820-2	22-40820-3
Phenol	mg/kg dry wt	0.3	<0.30	<0.30
2-Chlorophenol	mg/kg dry wt	0.3	<0.30	<0.30
2-Methylphenol	mg/kg dry wt	0.3	<0.30	<0.30
2-Nitrophenol	mg/kg dry wt	1.0	<1.0	<1.0

	Client	t Sample ID	12547621- BH-111(1.5) 1.5	12547621- BH-111(3.2) 3.2	
	Da	te Sampled	07/11/2022	07/11/2022	
2,4-Dimethylphenol	mg/kg dry wt	0.3	<0.30	<0.30	
2,4-Dichlorophenol	mg/kg dry wt	0.3	<0.30	<0.30	
2,6-Dichlorophenol	mg/kg dry wt	0.3	<0.30	<0.30	
4-Chloro-3- methylphenol	mg/kg dry wt	0.3	<0.30	<0.30	
2,4,5-Trichlorophenol	mg/kg dry wt	5	<5.0	<5.0	
2,4,6-Trichlorophenol	mg/kg dry wt	5	<5.0	<5.0	
2,3,4,6- Tetrachlorophenol	mg/kg dry wt	5	<5.0	<5.0	
4-Methylphenol	mg/kg dry wt	0.3	<0.30	<0.30	
Naphthalene	mg/kg dry wt	0.1	0.12	<0.10	
2-Methylnaphthalene	mg/kg dry wt	0.1	0.11	<0.10	
2-Chloronaphthalene	mg/kg dry wt	0.3	<0.30	<0.30	
Acenaphthene	mg/kg dry wt	0.1	<0.10	<0.10	
Acenaphthylene	mg/kg dry wt	0.1	0.14	<0.10	
Fluorene	mg/kg dry wt	0.1	0.13	<0.10	
Phenanthrene	mg/kg dry wt	0.1	0.58	0.14	
Anthracene	mg/kg dry wt	0.1	0.18	<0.10	
2-Phenylphenol	mg/kg dry wt	0.5	<0.50	<0.50	
Fluoranthene	mg/kg dry wt	0.1	0.96	0.33	
Benzo[a]anthracene	mg/kg dry wt	0.1	0.35	<0.10	
Chrysene	mg/kg dry wt	0.1	0.36	0.13	
Bis(2-ethylhexyl) adipate	mg/kg dry wt	0.5	<0.50	<0.50	
Benzo[b and i]fluoranthene	mg/kg dry wt	0.1	0.74	0.26	
Benzo[k]fluoranthene	mg/kg dry wt	0.1	0.31	0.13	
Benzo[a]pyrene	mg/kg dry wt	0.1	0.62	0.24	
Indeno(1,2,3-c,d)pyrene	mg/kg dry wt	0.1	0.48	0.19	
Dibenzo[a,h]anthracene	mg/kg dry wt	0.1	0.17	<0.10	
Benzo[g,h,i]perylene	mg/kg dry wt	0.1	0.36	0.15	
Pyrene	mg/kg dry wt	0.2	0.87	0.30	
Benzo[a]pyrene TEQ (LOR)	mg/kg dry wt	0.2	1.0	0.40	
Benzo[a]pyrene TEQ (Zero)	mg/kg dry wt	0.1	1.0	0.30	
4,4'-DDD	mg/kg dry wt	0.3	0.38	<0.30	
4,4'-DDE	mg/kg dry wt	0.3	<0.30	<0.30	
4,4'-DDT	mg/kg dry wt	0.5	<0.50	<0.50	
alpha-BHC	mg/kg dry wt	0.3	<0.30	<0.30	
beta-BHC	mg/kg dry wt	0.3	<0.30	<0.30	
gamma-BHC (Lindane)	mg/kg dry wt	0.3	<0.30	<0.30	
delta-BHC	mg/kg dry wt	0.3	<0.30	<0.30	
Aldrin	mg/kg dry wt	0.3	<0.30	<0.30	
cis-Chlordane	mg/kg dry wt	0.3	<0.30	<0.30	
trans-Chlordane	mg/kg dry wt	0.3	<0.30	<0.30	
Dieldrin	mg/kg dry wt	0.5	<0.50	<0.50	
Endosulfan I	mg/kg dry wt	0.3	<0.30	<0.30	
Endosulfan II	mg/kg dry wt	0.5	<0.50	<0.50	
Endosulfan sulfate	mg/kg dry wt	0.5	<0.50	<0.50	
Endrin	mg/kg dry wt	0.5	<0.50	<0.50	
Endrin aldehyde	mg/kg dry wt	0.5	<0.50	<0.50	
Endrin ketone	mg/kg dry wt	0.5	<0.50	<0.50	
Hexachlorobenzene	mg/kg dry wt	0.3	<0.30	<0.30	
	mg/kg dry wt	0.3	<0.30	<0.30	

	Client	: Sample ID	12547621- BH-111(1.5) 1.5	12547621- BH-111(3.2) 3.2
	Da	te Sampled	07/11/2022	07/11/2022
Heptachlor epoxide	mg/kg dry wt	0.3	<0.30	<0.30
Methoxychlor	mg/kg dry wt	0.5	<0.50	<0.50
Bis(2-ethylhexyl) phthalate	mg/kg dry wt	0.5	0.50	<0.50
Butyl benzyl phthalate	mg/kg dry wt	0.5	<0.50	<0.50
Di-n-butyl phthalate	mg/kg dry wt	1	<1.0	<1.0
Di-n-octyl phthalate	mg/kg dry wt	0.5	<0.50	<0.50
Diethyl phthalate	mg/kg dry wt	0.3	<0.30	<0.30
Dimethyl phthalate	mg/kg dry wt	0.3	<0.30	<0.30
N-Nitrosodiphenylamine	mg/kg dry wt	0.3	<0.30	<0.30
N-Nitrosodi-n- propylamine	mg/kg dry wt	0.3	<0.30	<0.30
2,4-Dinitrotoluene	mg/kg dry wt	0.3	<0.30	<0.30
2,6-Dinitrotoluene	mg/kg dry wt	0.3	<0.30	<0.30
Azobenzene	mg/kg dry wt	0.5	<0.50	<0.50
Isophorone	mg/kg dry wt	0.5	<0.50	<0.50
Nitrobenzene	mg/kg dry wt	0.3	<0.30	<0.30
4-Bromophenyl phenyl ether	mg/kg dry wt	0.3	<0.30	<0.30
4-Chlorophenyl phenyl ether	mg/kg dry wt	0.3	<0.30	<0.30
Bis(2-Chloroethyl) ether	mg/kg dry wt	0.3	<0.30	<0.30
Bis(2-Chloro-1- methylethyl) ether	mg/kg dry wt	0.3	<0.30	<0.30
Bis(2-Chloroethoxy) methane	mg/kg dry wt	0.3	<0.30	<0.30
1,2-Dichlorobenzene	mg/kg dry wt	0.3	<0.30	<0.30
1,3-Dichlorobenzene	mg/kg dry wt	0.3	<0.30	<0.30
1,4-Dichlorobenzene	mg/kg dry wt	0.3	<0.30	<0.30
Hexachlorobutadiene	mg/kg dry wt	0.3	<0.30	<0.30
Hexachlorocylopentadie ne	mg/kg dry wt	0.3	<0.30	<0.30
Hexachloroethane	mg/kg dry wt	0.3	<0.30	<0.30
4-Chloroaniline	mg/kg dry wt	1.0	<1.0	<1.0
2-Nitroaniline	mg/kg dry wt	0.3	<0.30	<0.30
3-Nitroaniline	mg/kg dry wt	0.5	<0.50	<0.50
Aniline	mg/kg dry wt	1.0	<1.0	<1.0
3,3'-Dichlorobenzidine	mg/kg dry wt	0.5	<0.50	<0.50
Dibenzofuran	mg/kg dry wt	0.3	<0.30	<0.30
Methyl methanesulfonate	mg/kg dry wt	1.0	<1.0	<1.0
Ethyl methanesulfonate	mg/kg dry wt	1	<1.0	<1.0
Benzyl alcohol	mg/kg dry wt	1	<1.0	<1.0
Phenol-d5 (Surrogate)	%	1	89	90
2-Fluorophenol (Surrogate)	%	1	91	93
2-Fluorobiphenyl (Surrogate)	%	1	98	100
2,4,6-Tribromophenol (Surrogate)	%	1	98	120
p-Terphenyl-d14 (Surrogate)	%	1	96	93
Nitrobenzene-d5 (Surrogate)	%	1	98	98

Method Summary

Elements in Soil Samples dried and passed through a 2 mm sieve followed by acid digestion and analysis by ICP-

MS. In accordance with in-house procedure based on US EPA method 200.8.

PAH in Soil Solvent extraction, silica cleanup, followed by GC-MS analysis.

> Benzo[a]pyrene TEQ (LOR): The most conservative TEQ estimate, where a result is reported as less than the limit of reporting (LOR) the LOR value is used to calculate the TEQ for that PAH. Benzo[a]pyrene TEQ (Zero): The least conservative TEQ estimate, PAHs reported as less than

the limit of reporting (LOR) are not included in the TEQ calculation.

Benzo[a]pyrene toxic equivalence (TEQ) is calculated according to 'Methodology for Deriving Standards for Contaminants in Soil to Protect Human Health'. Ministry for the Environment. 2011.

(In accordance with in-house procedure).

Solvent extraction, silica cleanup, followed by GC-FID analysis. (C7-C36). (In accordance with in-**TPH in Soil**

house procedure based on US EPA 8015).

Moisture Moisture content is determined gravimetrically by drying at 103 °C.

SVOC in Soil Solvent extraction, followed by GC-MS analysis.(In-house based on US EPA 8270).

Thara Samarasinghe, B.Sc.

Technician

Astra Southwood Lab Technician

Prianshu Chawla, B.Tech

Technologist

Senior Technologist



Analytica Laboratories Limited Unit 1, 30 Greenpark Road Penrose Auckland sales@analytica.co.nz www.analytica.co.nz

Certificate of Analysis

GHD Ltd

Level 1, Bing Harris Building, 286 Princess Street, Dunedin

Dunedin 9016

Attention: Danny Fitzgerald Phone: 021 973 994

Email: danny.fitzgerald@ghd.com

Sampling Site:

Description of Work: Soils - Green Island LF

Green Island LF

Report Comments

Samples were collected by yourselves (or your agent) and analysed as received at Analytica Laboratories. Samples were in acceptable condition unless otherwise noted on this report.

Lab Reference:

Date Received:

Order Number:

Reference:

Testing Initiated:

Date Completed: 14/11/2022

Submitted by:

22-40968

09/11/2022

9/11/2022

12547621

N/A

Danny Fitzgerald

Specific testing dates are available on request.

Asbestos in Soil (Semi-Quantitative)

Sample Details

Laboratory II	Client Sample ID	Sample Location	Sample Description	Date Sampled	Date Analysed
22-40968-1	12547621-BH-107 (0.5)	N/A	Soil	09/11/2022	14/11/2022

Information in the above table supplied by the client: Client Sample ID, Sample Location, Date Sampled

Analysis Results (Summary)

Laboratory ID	Client Sample ID	Asbestos	Sample Weight as Received	Moisture Content	Trace Asbestos (Presence / Absence)	Asbestos (Presence / Absence)
	Units		g	%		
22-40968-1	12547621-BH-107 (0.5)	Asbestos NOT Detected. Organic Fibres	1,152.0	18.2	Absent	Absent

Information in the above table supplied by the client: Client Sample ID



Analysis Results (Size Fraction Breakdown)

Laboratory ID	Client Sample ID	Fraction Size	Fraction Weight*	AF/FA Weight*	ACM Weight*	ACM Content*	Asbestos Matrix	Asbestos Weight*	W/W% Asbestos*
	Units Reporting Limit		g 0	g 0	g 0	%		g 0	
		>10mm	215.00	0.0000	0.0000	0	No Asbestos Detected	0.0000	<0.001
22-40968-1	12547621-BH-107 (0.5)	2-10mm	425.50	0.0000	-	-	No Asbestos Detected	0.0000	(ACM) <0.001
		<2mm	302.00	0.0000	-	-	No Asbestos Detected	0.0000	(AF/FA)

Information in the above table supplied by the client: Client Sample ID

Asbestos in Soil (Semi-Quantitative) Approver:

Emily Wang, M.Sc. Laboratory Technician

Method Summary

Asbestos Fibres in Soil (Semi-Quantitative)

Sample analysis was performed using polarised light microscopy with dispersion staining in accordance with AS4964-2004 Method for the qualitative identification of asbestos in soil samples.

Note 1: The reporting limit for this analysis is 0.1g/kg (0.01%) by application of polarised light microscopy, dispersion staining and trace analysis techniques.

Note 2: Trace asbestos is indicative that freely liberated respirable fibres are present and dust control measures should be implemented or increased on site. This is not the sole indicator for the friable nature of the asbestos present.

Note 3: If mineral fibres of unknown type are detected, by PLM and dispersion staining, these may or may not be asbestos fibres. To confirm the identity of this fibre, another independent analytical technique such as XRD analysis is advised.

Note 4: The laboratory does not take responsibility for the sampling procedure or accuracy of sample location description.



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Certificate of Analysis

GHD Ltd

Level 1, Bing Harris Building, 286 Princess Street, Dunedin

Dunedin 9016

Attention: Danny Fitzgerald Phone: 021 973 994

Email: danny.fitzgerald@ghd.com

Sampling Site: Green Island LF

Description of Work: Soils - Green Island LF

Lab Reference: 22-40968

Submitted by: Danny Fitzgerald

Date Received: 09/11/2022 Testing Initiated: 9/11/2022 Date Completed: 14/11/2022

Order Number: N/A

Reference: 12547621

Report Comments

Samples were collected by yourselves (or your agent) and analysed as received at Analytica Laboratories. Samples were in acceptable condition unless otherwise noted on this report.

Specific testing dates are available on request.

Asbestos in Soil (Qualitative)

Sample Details

Laboratory ID	Client Sample ID	Sample Location	Sample Description	Date Sampled	Date Analysed
22-40968-2	12547621-BH-108 (0.5)	N/A	Soil	09/11/2022	14/11/2022
22-40968-3	12547621-BH-108 (1.9)	N/A	Soil	09/11/2022	14/11/2022
22-40968-4	12547621-BH-108 (3.3)	N/A	Soil	09/11/2022	14/11/2022
22-40968-5	12547621-BH-104 (0.5)	N/A	Soil	09/11/2022	14/11/2022
22-40968-6	12547621-BH-104 (1.0)	N/A	Soil	09/11/2022	14/11/2022
22-40968-7	12547621-BH-104 (2.0)	N/A	Soil	09/11/2022	14/11/2022

Information in the above table supplied by the client: Client Sample ID, Sample Location, Date Sampled.

Laboratory ID	Client Sample ID	Fibre Types	Trace Asbestos (Presence / Absence)	Asbestos (Presence / Absence)
	Units			
22-40968-2	12547621-BH-108 (0.5)	Asbestos NOT Detected. Organic Fibres	Absent	Absent
22-40968-3	12547621-BH-108 (1.9)	Asbestos NOT Detected. Organic Fibres	Absent	Absent
22-40968-4	12547621-BH-108 (3.3)	Asbestos NOT Detected. Organic Fibres	Absent	Absent
22-40968-5	12547621-BH-104 (0.5)	Asbestos NOT Detected. Organic Fibres	Absent	Absent
22-40968-6	12547621-BH-104 (1.0)	Asbestos NOT Detected. Organic Fibres	Absent	Absent
22-40968-7	12547621-BH-104 (2.0)	Asbestos NOT Detected. Organic Fibres	Absent	Absent

Information in the above table supplied by the client: Client Sample ID.

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Asbestos in Soil (Qualitative) Approver:

Emily Wang, M.Sc. Laboratory Technician

Method Summary

Asbestos Fibres in Soil (Qualitative)

Sample analysis was performed using polarised light microscopy with dispersion staining in accordance with AS4964-2004 Method for the qualitative identification of asbestos in bulk samples.

Note 1: The reporting limit for this analysis is 0.1g/kg (0.01%) by application of polarised light microscopy, dispersion staining and trace analysis techniques.

Note 2: Trace asbestos is indicative that freely liberated respirable fibres are present and dust control measures should be implemented or increased on site. This is not the sole indicator for the friable nature of the asbestos present.

Note 3: If mineral fibres of unknown type are detected, by PLM and dispersion staining, these may or may not be asbestos fibres. To confirm the identity of this fibre, another independent analytical technique such as XRD analysis is advised.

Note 4: The laboratory does not take responsibility for the sampling procedure or accuracy of sample location description.



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Certificate of Analysis

GHD Ltd

Level 1, Bing Harris Building, 286 Princess Street, Dunedin

Dunedin 9016

Attention: Cecilia Gately Phone: 021 973 994

Email: danny.fitzgerald@ghd.com

Sampling Site: Green Island LF

Lab Reference: 22-41184

Submitted by: Danny FitzGerald

Date Received: 10/11/2022 Testing Initiated: 11/11/2022 Date Completed: 16/11/2022

Order Number:

Reference: 12547621

Report Comments

Samples were collected by yourselves (or your agent) and analysed as received at Analytica Laboratories. Samples were in acceptable condition unless otherwise noted on this report.

Specific testing dates are available on request.

Heavy Metals in Soil

	Client Sample ID		12547621 - BH - 108 0.5	12547621 - BH - 108 1.9	12547621 - BH - 108 3.3	12547621 - DUP-2	12547621 - BH - 104 0.5
	Da	te Sampled	09/11/2022	09/11/2022	09/11/2022	09/11/2022	09/11/2022
Analyte	Unit	Reporting Limit	22-41184-1	22-41184-2	22-41184-3	22-41184-4	22-41184-5
Arsenic	mg/kg dry wt	0.125	2.4	7.2	9.1	7.1	7.7
Beryllium	mg/kg dry wt	0.013	0.48	1.42	3.15	1.33	1.79
Boron	mg/kg dry wt	1.25	1.3	7.6	40	7.9	3.4
Cadmium	mg/kg dry wt	0.005	0.0056	0.16	0.42	0.16	0.15
Chromium	mg/kg dry wt	0.125	11	50.1	34.7	49.4	46.6
Copper	mg/kg dry wt	0.075	3.3	20.9	54.2	20.8	38.3
Lead	mg/kg dry wt	0.25	11.7	52.9	316	42.8	70.7
Mercury	mg/kg dry wt	0.025	<0.025	0.28	0.21	0.27	0.078
Nickel	mg/kg dry wt	0.05	3.9	38.0	31.6	35.9	35.8
Zinc	mg/kg dry wt	0.05	20.3	124	303	121	115

Heavy Metals in Soil

	Client	12547621 - BH - 104 1.0	12547621 - BH - 104 2.0	
	Da	te Sampled	09/11/2022	09/11/2022
Analyte	Unit	Reporting Limit	22-41184-6	22-41184-7
Arsenic	mg/kg dry wt	0.125	4.1	8.9
Beryllium	mg/kg dry wt	0.013	0.91	0.81
Boron	mg/kg dry wt	1.25	2.5	11
Cadmium	mg/kg dry wt	0.005	0.070	0.12

All tests reported herein have been performed in accordance with the laboratory's scope of accreditation with the exception of tests marked *, which are not accredited.





Heavy Metals in Soil

	Client	12547621 - BH - 104 1.0	12547621 - BH - 104 2.0	
	Da	te Sampled	09/11/2022	09/11/2022
Chromium	mg/kg dry wt	mg/kg dry wt 0.125		27.6
Copper	mg/kg dry wt	0.075	18.0	16.9
Lead	mg/kg dry wt	0.25	26.9	16.5
Mercury	mg/kg dry wt	0.025	0.033	0.039
Nickel	mg/kg dry wt	0.05	19.3	18.9
Zinc	mg/kg dry wt	0.05	68.1	68.3

Polycyclic Aromatic Hydrocarbons - Soil

	Clien	t Sample ID	12547621 - BH - 108 0.5	12547621 - BH - 108 1.9	12547621 - BH - 108 3.3	12547621 - DUP-2	12547621 - BH - 104 0.5
	Da	ite Sampled	09/11/2022	09/11/2022	09/11/2022	09/11/2022	09/11/2022
Analyte	Unit	Reporting Limit	22-41184-1	22-41184-2	22-41184-3	22-41184-4	22-41184-5
1-Methylnaphthalene	mg/kg dry wt	0.01	<0.010	<0.010	0.046	<0.010	2.1
2-Methylnaphthalene	mg/kg dry wt	0.01	<0.010	<0.010	0.051	0.011	1.7
Acenaphthene	mg/kg dry wt	0.01	<0.010	<0.010	0.033	0.026	12
Acenaphthylene	mg/kg dry wt	0.01	<0.010	0.031	0.055	0.14	53
Anthracene	mg/kg dry wt	0.01	<0.010	0.054	0.16	1.5	93
Benz[a]anthracene	mg/kg dry wt	0.02	<0.020	0.22	0.73	4.6	320
Benzo[a]pyrene	mg/kg dry wt	0.01	<0.010	0.35	0.75	2.7	290
Benzo[b]&[j] fluoranthene	mg/kg dry wt	0.02	<0.020	0.37	1.1	4.0	500
Benzo[g,h,i]perylene	mg/kg dry wt	0.02	<0.020	0.21	0.43	1.2	170
Benzo[k]fluoranthene	mg/kg dry wt	0.01	<0.010	0.18	0.55	1.8	220
Chrysene	mg/kg dry wt	0.01	<0.010	0.38	0.92	3.7	320
Dibenz(a,h)anthracene	mg/kg dry wt	0.01	<0.010	0.040	0.12	0.43	50
Fluoranthene	mg/kg dry wt	0.02	<0.020	0.42	1.1	5.4	420
Fluorene	mg/kg dry wt	0.01	<0.010	0.017	0.082	0.30	23
Indeno(1,2,3-cd)pyrene	mg/kg dry wt	0.01	<0.010	0.21	0.48	1.5	220
Naphthalene	mg/kg dry wt	0.01	<0.010	<0.010	0.077	0.017	2.3
Phenanthrene	mg/kg dry wt	0.01	<0.010	0.15	0.58	4.8	220
Pyrene	mg/kg dry wt	0.02	<0.020	0.45	1.2	4.9	380
Benzo[a]pyrene TEQ (LOR)	mg/kg dry wt	0.03	0.030	0.49	1.2	4.4	470
Benzo[a]pyrene TEQ (Zero)	mg/kg dry wt	0.01	<0.010	0.49	1.2	4.4	470
Anthracene-d10 (Surrogate)	%	1	95	99	110	120	110

Polycyclic Aromatic Hydrocarbons - Soil

	Clien	12547621 - BH - 104 1.0	12547621 - BH - 104 2.0	
	Da	09/11/2022	09/11/2022	
Analyte	Unit	22-41184-6	22-41184-7	
1-Methylnaphthalene	mg/kg dry wt	0.01	0.014	0.012
2-Methylnaphthalene	mg/kg dry wt	0.01	0.021	0.015
Acenaphthene	mg/kg dry wt	0.01	0.010	0.019
Acenaphthylene	mg/kg dry wt	0.01	0.011	0.027
Anthracene	mg/kg dry wt	0.01	0.015	0.12
Benz[a]anthracene	mg/kg dry wt	0.039	0.28	
Benzo[a]pyrene	mg/kg dry wt	0.01	0.058	0.32

Polycyclic Aromatic Hydrocarbons - Soil

	Client	t Sample ID	12547621 - BH - 104 1.0	12547621 - BH - 104 2.0
	Da	te Sampled	09/11/2022	09/11/2022
Benzo[b]&[j] fluoranthene	mg/kg dry wt	0.02	0.058	0.33
Benzo[g,h,i]perylene	mg/kg dry wt	0.02	0.026	0.14
Benzo[k]fluoranthene	mg/kg dry wt	0.01	0.028	0.17
Chrysene	mg/kg dry wt	0.01	0.059	0.41
Dibenz(a,h)anthracene	mg/kg dry wt	0.01	<0.010	0.032
Fluoranthene	mg/kg dry wt	0.02	0.076	0.65
Fluorene	mg/kg dry wt	0.01	0.015	0.050
Indeno(1,2,3-cd)pyrene	mg/kg dry wt	0.01	0.029	0.16
Naphthalene	mg/kg dry wt	0.01	0.083	0.19
Phenanthrene	mg/kg dry wt	0.01	0.032	0.43
Pyrene	mg/kg dry wt	0.02	0.083	0.63
Benzo[a]pyrene TEQ (LOR)	mg/kg dry wt	0.03	0.080	0.46
Benzo[a]pyrene TEQ (Zero)	mg/kg dry wt	0.01	0.070	0.46
Anthracene-d10 (Surrogate)	%	1	100	110

Moisture Content

CI	Client Sample ID		12547621 - BH - 108 1.9	12547621 - BH - 108 3.3	12547621 - DUP-2	12547621 - BH - 104 0.5
Date Sampled		09/11/2022	09/11/2022	09/11/2022	09/11/2022	09/11/2022
Analyte Unit Reporting Limit		22-41184-1	22-41184-2	22-41184-3	22-41184-4	22-41184-5
Moisture Content	6 1	19	23	15	24	19

Moisture Content

	Clien	t Sample ID	12547621 - BH - 104 1.0	12547621 - BH - 104 2.0
	Da	te Sampled	09/11/2022	09/11/2022
Analyte	Unit	Reporting Limit	22-41184-6	22-41184-7
Moisture Content	%	1	19	26

Method Summary

Elements in Soil Samples dried and passed through a 2 mm sieve followed by acid digestion and analysis by ICP-

MS. In accordance with in-house procedure based on US EPA method 200.8.

PAH in Soil Solvent extraction, silica cleanup, followed by GC-MS analysis.

Benzo[a]pyrene TEQ (LOR): The most conservative TEQ estimate, where a result is reported as less than the limit of reporting (LOR) the LOR value is used to calculate the TEQ for that PAH. **Benzo[a]pyrene TEQ (Zero)**: The least conservative TEQ estimate, PAHs reported as less than

the limit of reporting (LOR) are not included in the TEQ calculation.

Benzo[a]pyrene toxic equivalence (TEQ) is calculated according to 'Methodology for Deriving Standards for Contaminants in Soil to Protect Human Health'. Ministry for the Environment. 2011.

(In accordance with in-house procedure).

Moisture Moisture content is determined gravimetrically by drying at 103 °C.

Jarred Wilson, DipSci Trace Elements Team Leader

Astra Southwood Lab Technician Prianshu Chawla, B.Tech Technologist Brent Boynes Lab Technician



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Certificate of Analysis

GHD Ltd

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Dunedin 9016

Attention: Cecilia Gately Phone: 021 973 994

Email: danny.fitzgerald@ghd.com

Sampling Site: Green Island Landfill

Lab Reference: 22-42925

Submitted by: Danny Fitzgerald

Date Received: 24/11/2022 Testing Initiated: 24/11/2022 Date Completed: 29/11/2022

Order Number:

Reference: 12547621

Report Comments

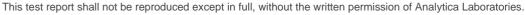
Samples were collected by yourselves (or your agent) and analysed as received at Analytica Laboratories. Samples were in acceptable condition unless otherwise noted on this report.

Specific testing dates are available on request.

Polycyclic Aromatic Hydrocarbons - Soil

	Client Sample ID 12547621-BH-107 2.5 12547621-BH-103 125476		12547621-BH-103 1.2	12547621-BH-102 1.5	12547621- BH-102 2.5			
	Da	ite Sampled	01/11/2022	01/11/2022	01/11/2022	01/11/2022	01/11/2022	
Analyte	Unit	Reporting Limit	22-42925-1	22-42925-2	22-42925-3	22-42925-4	22-42925-5	
1-Methylnaphthalene	mg/kg dry wt	0.01	0.028	<0.010	<0.010	<0.010	<0.010	
2-Methylnaphthalene	mg/kg dry wt	0.01	0.027	<0.010	<0.010	<0.010	<0.010	
Acenaphthene	mg/kg dry wt	0.01	0.033	<0.010	<0.010	<0.010	<0.010	
Acenaphthylene	mg/kg dry wt	0.01	0.067	0.017	0.028	<0.010	<0.010	
Anthracene	mg/kg dry wt	0.01	0.20	0.020	0.020	<0.010	<0.010	
Benz[a]anthracene	mg/kg dry wt	0.02	0.53	0.092	0.14	0.023	<0.020	
Benzo[a]pyrene	mg/kg dry wt	0.01	0.74	0.12	0.23	0.043	<0.010	
Benzo[b]&[j] fluoranthene	mg/kg dry wt	0.02	0.73	0.12	0.27	0.054	<0.020	
Benzo[g,h,i]perylene	mg/kg dry wt	0.02	0.26	0.064	0.13	0.028	<0.020	
Benzo[k]fluoranthene	mg/kg dry wt	0.01	0.26	0.044	0.091	0.015	<0.010	
Chrysene	mg/kg dry wt	0.01	0.52	0.095	0.20	0.038	<0.010	
Dibenz(a,h)anthracene	mg/kg dry wt	0.01	0.061	0.012	0.023	<0.010	<0.010	
Fluoranthene	mg/kg dry wt	0.02	1.5	0.17	0.31	0.055	<0.020	
Fluorene	mg/kg dry wt	0.01	0.083	<0.010	<0.010	<0.010	<0.010	
Indeno(1,2,3-cd)pyrene	mg/kg dry wt	0.01	0.35	0.082	0.17	0.037	<0.010	
Naphthalene	mg/kg dry wt	0.01	0.053	<0.010	<0.010	<0.010	<0.010	
Phenanthrene	mg/kg dry wt	0.01	0.58	0.057	0.083	0.015	<0.010	
Pyrene	mg/kg dry wt	0.02	1.4	0.17	0.32	0.060	<0.020	
Benzo[a]pyrene TEQ (LOR)	mg/kg dry wt	0.03	1.0	0.17	0.32	0.070	0.030	
Benzo[a]pyrene TEQ (Zero)	mg/kg dry wt	0.01	1.0	0.17	0.32	0.060	<0.010	

All tests reported herein have been performed in accordance with the laboratory's scope of accreditation with the exception of tests marked *, which are not accredited.





Polycyclic Aromatic Hydrocarbons - Soil

	Client	: Sample ID	12547621-BH-107 2.5	12547621-BH-103 0.2	12547621-BH-103 1.2	12547621-BH-102 1.5	12547621- BH-102 2.5
	Date Sampled 01/11/2022		01/11/2022	01/11/2022	01/11/2022	01/11/2022	
Anthracene-d10 (Surrogate)	%	1	100	110	110	110	100

Moisture Content

	Clien	Client Sample ID 12547621-BH-107 2.5		12547621-BH-103 0.2	12547621-BH-103 1.2	12547621-BH-102 1.5	12547621- BH-102 2.5
Date Sampled		01/11/2022	01/11/2022	01/11/2022	01/11/2022	01/11/2022	
Analyte	Unit Reporting 22-42925-1		22-42925-2	22-42925-3	22-42925-4	22-42925-5	
Moisture Content	%	1	21	23	53	48	27

Method Summary

PAH in Soil

Solvent extraction, silica cleanup, followed by GC-MS analysis.

Benzo[a]pyrene TEQ (LOR): The most conservative TEQ estimate, where a result is reported as less than the limit of reporting (LOR) the LOR value is used to calculate the TEQ for that PAH. **Benzo[a]pyrene TEQ (Zero)**: The least conservative TEQ estimate, PAHs reported as less than

the limit of reporting (LOR) are not included in the TEQ calculation.

Benzo[a]pyrene toxic equivalence (TEQ) is calculated according to 'Methodology for Deriving Standards for Contaminants in Soil to Protect Human Health'. Ministry for the Environment. 2011.

(In accordance with in-house procedure).

Moisture

Moisture content is determined gravimetrically by drying at 103 °C.

Prianshu Chawla, B.Tech

Technologist

Brent Boynes Lab Technician



Analytica Laboratories LTD 10 Bisley Road Hamilton New Zealand 3214



All tests reported herein have been performed in accordance with the laboratory's scope of accreditation

Attention: Customer Service

Report 938712-S

Project name

Project ID 22-40179
Received Date Nov 08, 2022

Client Sample ID			22-40179-1	22-40179-2	22-40179-3	22-40179-4
Sample Matrix			Soil	Soil	Soil	Soil
			K22-	K22-	K22-	K22-
Eurofins Sample No.			No0015975	No0015976	No0015977	No0015978
Date Sampled			Nov 02, 2022	Nov 02, 2022	Nov 02, 2022	Nov 02, 2022
Test/Reference	LOR	Unit				
Polycyclic Aromatic Hydrocarbons (NZ MfE)						
Acenaphthene	0.03	mg/kg	< 0.03	< 0.03	< 0.03	< 0.03
Acenaphthylene	0.03	mg/kg	< 0.03	< 0.03	< 0.03	< 0.03
Anthracene	0.03	mg/kg	< 0.03	< 0.03	< 0.03	< 0.03
Benz(a)anthracene	0.03	mg/kg	0.07	0.06	< 0.03	< 0.03
Benzo(a)pyrene	0.03	mg/kg	0.06	0.05	< 0.03	< 0.03
Benzo(a)pyrene TEQ (lower bound)*	0.03	mg/kg	0.07	0.07	< 0.03	< 0.03
Benzo(a)pyrene TEQ (medium bound)*	0.03	mg/kg	0.09	0.09	0.04	0.04
Benzo(a)pyrene TEQ (upper bound)*	0.03	mg/kg	0.11	0.10	0.08	0.08
Benzo(b&j)fluoranthene ^{N07}	0.03	mg/kg	0.04	0.06	< 0.03	< 0.03
Benzo(g.h.i)perylene	0.03	mg/kg	< 0.03	0.06	< 0.03	< 0.03
Benzo(k)fluoranthene	0.03	mg/kg	< 0.03	0.06	< 0.03	< 0.03
Chrysene	0.03	mg/kg	0.09	0.09	< 0.03	< 0.03
Dibenz(a.h)anthracene	0.03	mg/kg	< 0.03	< 0.03	< 0.03	< 0.03
Fluoranthene	0.03	mg/kg	0.11	0.11	< 0.03	< 0.03
Fluorene	0.03	mg/kg	< 0.03	< 0.03	< 0.03	< 0.03
Indeno(1.2.3-cd)pyrene	0.03	mg/kg	< 0.03	0.03	< 0.03	< 0.03
Naphthalene	0.1	mg/kg	< 0.1	< 0.1	< 0.1	< 0.1
Phenanthrene	0.03	mg/kg	0.07	0.05	< 0.03	< 0.03
Pyrene	0.03	mg/kg	0.13	0.11	< 0.03	< 0.03
p-Terphenyl-d14 (surr.)	1	%	83	79	73	77
2-Fluorobiphenyl (surr.)	1	%	76	68	62	61
Semivolatile Organics						
1-Chloronaphthalene	0.5	mg/kg	< 0.5	< 0.5	-	-
1-Naphthylamine	0.5	mg/kg	< 0.5	< 0.5	-	-
1.2-Dichlorobenzene	0.5	mg/kg	< 0.5	< 0.5	-	-
1.2.3-Trichlorobenzene	0.5	mg/kg	< 0.5	< 0.5	-	-
1.2.3.4-Tetrachlorobenzene	0.5	mg/kg	< 0.5	< 0.5	-	-
1.2.3.5-Tetrachlorobenzene	0.5	mg/kg	< 0.5	< 0.5	-	-
1.2.4-Trichlorobenzene	0.5	mg/kg	< 0.5	< 0.5	-	-
1.2.4.5-Tetrachlorobenzene	0.5	mg/kg	< 0.5	< 0.5	-	-
1.3-Dichlorobenzene	0.5	mg/kg	< 0.5	< 0.5	-	-
1.3.5-Trichlorobenzene	0.5	mg/kg	< 0.5	< 0.5	-	-
1.4-Dichlorobenzene	0.5	mg/kg	< 0.5	< 0.5	-	-
2-Chloronaphthalene	0.5	mg/kg	< 0.5	< 0.5	-	-
2-Chlorophenol	0.5	mg/kg	< 0.5	< 0.5	-	-



Client Sample ID			22-40179-1	22-40179-2	22-40179-3	22-40179-4
Sample Matrix			Soil	Soil	Soil	Soil
Eurofins Sample No.			K22- No0015975	K22- No0015976	K22- No0015977	K22- No0015978
Date Sampled			Nov 02, 2022	Nov 02, 2022	Nov 02, 2022	Nov 02, 2022
Test/Reference	LOR	Unit	1107 02, 2022	1107 02, 2022	1101 02, 2022	1107 02, 2022
Semivolatile Organics	LOR	Offic				
		0/	70			
2-Fluorobiphenyl (surr.)	1 5	%	76	68	-	-
2-Methyl-4.6-dinitrophenol	5	mg/kg	< 5	< 5	-	-
2-Methylnaphthalene	0.5	mg/kg	< 0.5	< 0.5	-	-
2-Methylphenol (o-Cresol)	0.2	mg/kg	< 0.2	< 0.2	-	-
2-Naphthylamine	0.5	mg/kg	< 0.5	< 0.5	-	-
2-Nitroaniline	0.5	mg/kg	< 0.5	< 0.5	-	-
2-Nitrophenol	1	mg/kg	< 1	< 1	-	-
2-Picoline	0.5	mg/kg	< 0.5	< 0.5	-	-
2.3.4.6-Tetrachlorophenol	5	mg/kg	< 5	< 5	-	-
2.4-Dichlorophenol	0.5	mg/kg	< 0.5	< 0.5	-	-
2.4-Dimethylphenol	0.5	mg/kg	< 0.5	< 0.5	=	-
2.4-Dinitrophenol	5	mg/kg	< 5	< 5	-	-
2.4-Dinitrotoluene	0.5	mg/kg	< 0.5	< 0.5	-	-
2.4.5-Trichlorophenol	1	mg/kg	< 1	< 1	-	-
2.4.6-Tribromophenol (surr.)	1	%	92	85	-	-
2.4.6-Trichlorophenol	1	mg/kg	< 1	< 1	-	-
2.6-Dichlorophenol	0.5	mg/kg	< 0.5	< 0.5	-	-
2.6-Dinitrotoluene	0.5	mg/kg	< 0.5	< 0.5	-	-
3&4-Methylphenol (m&p-Cresol)	0.4	mg/kg	< 0.4	< 0.4	-	-
3-Methylcholanthrene	0.5	mg/kg	< 0.5	< 0.5	-	-
3.3'-Dichlorobenzidine	0.5	mg/kg	< 0.5	< 0.5	-	-
4-Aminobiphenyl	0.5	mg/kg	< 0.5	< 0.5	-	-
4-Bromophenyl phenyl ether	0.5	mg/kg	< 0.5	< 0.5	-	-
4-Chloro-3-methylphenol	1	mg/kg	< 1	< 1	-	-
4-Chlorophenyl phenyl ether	0.5	mg/kg	< 0.5	< 0.5	-	-
4-Nitrophenol	5	mg/kg	< 5	< 5	-	-
4.4'-DDD	0.01	mg/kg	< 0.01	< 0.01	-	-
4.4'-DDE	0.01	mg/kg	< 0.01	< 0.01	-	-
4.4'-DDT	0.01	mg/kg	< 0.01	< 0.01	-	-
7.12-Dimethylbenz(a)anthracene	0.5	mg/kg	< 0.5	< 0.5	-	-
a-HCH	0.01	mg/kg	< 0.01	< 0.01	-	-
Acenaphthene	0.03	mg/kg	< 0.03	< 0.03	-	-
Acenaphthylene	0.03	mg/kg	< 0.03	< 0.03	-	-
Acetophenone	0.5	mg/kg	< 0.5	< 0.5	-	-
Aldrin	0.01	mg/kg	< 0.01	< 0.01	-	-
Aniline	0.5	mg/kg	< 0.5	< 0.5	-	-
Anthracene	0.03	mg/kg	< 0.03	< 0.03	-	-
b-HCH	0.01	mg/kg	< 0.01	< 0.01	-	-
Benz(a)anthracene	0.03	mg/kg	0.07	0.06	-	-
Benzo(a)pyrene	0.03	mg/kg	0.06	0.05	-	-
Benzo(a)pyrene TEQ (lower bound)*	0.03	mg/kg	0.07	0.07	-	-
Benzo(a)pyrene TEQ (medium bound)*	0.03	mg/kg	0.09	0.09	-	-
Benzo(a)pyrene TEQ (upper bound)*	0.03	mg/kg	0.11	0.10	-	-
Benzo(b&j)fluoranthene ^{N07}	0.03	mg/kg	0.04	0.06	_	_
Benzo(g.h.i)perylene	0.03	mg/kg	< 0.03	0.06	_	_
Benzo(k)fluoranthene	0.03	mg/kg	< 0.03	0.06	_	_
Benzyl chloride	0.5	mg/kg	< 0.5	< 0.5	_	_
Bis(2-chloroethoxy)methane	0.5	mg/kg	< 0.5	< 0.5	-	_
Bis(2-chloroisopropyl)ether	0.5	mg/kg	< 0.5	< 0.5	-	_

Report Number: 938712-S



Client Sample ID			22-40179-1	22-40179-2	22-40179-3	22-40179-4
Sample Matrix			Soil	Soil	Soil	Soil
Eurofins Sample No.			K22- No0015975	K22- No0015976	K22- No0015977	K22- No0015978
Date Sampled			Nov 02, 2022	Nov 02, 2022	Nov 02, 2022	Nov 02, 2022
Test/Reference	LOR	Unit				
Semivolatile Organics		•				
Bis(2-ethylhexyl)phthalate	0.5	mg/kg	< 0.5	< 0.5	-	-
Butyl benzyl phthalate	0.5	mg/kg	< 0.5	< 0.5	-	-
Chrysene	0.03	mg/kg	0.09	0.09	-	-
d-HCH	0.01	mg/kg	< 0.01	< 0.01	-	-
Di-n-butyl phthalate	0.5	mg/kg	< 0.5	< 0.5	-	-
Di-n-octyl phthalate	0.5	mg/kg	< 0.5	< 0.5	-	-
Dibenz(a.h)anthracene	0.03	mg/kg	< 0.03	< 0.03	-	-
Dibenz(a.j)acridine	0.5	mg/kg	< 0.5	< 0.5	-	-
Dibenzofuran	0.5	mg/kg	< 0.5	< 0.5	-	-
Dieldrin	0.01	mg/kg	< 0.01	< 0.01	-	-
Diethyl phthalate	0.5	mg/kg	< 0.5	< 0.5	-	-
Dimethyl phthalate	0.5	mg/kg	< 0.5	< 0.5	-	-
Dimethylaminoazobenzene	0.5	mg/kg	< 0.5	< 0.5	-	-
Diphenylamine	0.5	mg/kg	< 0.5	< 0.5	-	-
Endosulfan I	0.01	mg/kg	< 0.01	< 0.01	-	-
Endosulfan II	0.01	mg/kg	< 0.01	< 0.01	-	-
Endosulfan sulphate	0.01	mg/kg	< 0.01	< 0.01	-	-
Endrin	0.01	mg/kg	< 0.01	< 0.01	-	-
Endrin aldehyde	0.01	mg/kg	< 0.01	< 0.01	-	-
Endrin ketone	0.01	mg/kg	< 0.01	< 0.01	-	-
Fluoranthene	0.03	mg/kg	0.11	0.11	-	-
Fluorene	0.03	mg/kg	< 0.03	< 0.03	-	-
g-HCH (Lindane)	0.01	mg/kg	< 0.01	< 0.01	-	-
Heptachlor	0.01	mg/kg	< 0.01	< 0.01	-	-
Heptachlor epoxide	0.01	mg/kg	< 0.01	< 0.01	-	-
Hexachlorobenzene	0.01	mg/kg	< 0.01	< 0.01	-	-
Hexachlorobutadiene	0.5	mg/kg	< 0.5	< 0.5	-	-
Hexachlorocyclopentadiene	0.5	mg/kg	< 0.5	< 0.5	-	-
Hexachloroethane	0.5	mg/kg	< 0.5	< 0.5	-	-
Indeno(1.2.3-cd)pyrene	0.03	mg/kg	< 0.03	0.03	-	-
Methoxychlor	0.01	mg/kg	< 0.01	< 0.01	-	-
N-Nitrosodibutylamine	0.5	mg/kg	< 0.5	< 0.5	-	-
N-Nitrosodipropylamine	0.5	mg/kg	< 0.5	< 0.5	-	-
N-Nitrosopiperidine	0.5	mg/kg	< 0.5	< 0.5	-	-
Naphthalene	0.1	mg/kg	< 0.1	< 0.1	-	-
Nitrobenzene	0.5	mg/kg	< 0.5	< 0.5	-	-
Nitrobenzene-d5 (surr.)	1	%	89	81	-	-
Pentachlorobenzene	0.5	mg/kg	< 0.5	< 0.5	-	-
Pentachloronitrobenzene	0.5	mg/kg	< 0.5	< 0.5	-	-
Pentachlorophenol	1	mg/kg	< 1	< 1	-	-
Phenanthrene	0.03	mg/kg	0.07	0.05	-	-
Phenol	0.5	mg/kg	< 0.5	< 0.5	-	-
Phenol-d6 (surr.)	1	%	70	69	-	-
Pronamide	0.5	mg/kg	< 0.5	< 0.5	-	-
Pyrene	0.03	mg/kg	0.13	0.11	-	-
Trifluralin	0.5	mg/kg	< 0.5	< 0.5	-	-



Client Sample ID			22-40179-5	22-40179-6	22-40179-7	22-40179-8
Sample Matrix			Soil	Soil	Soil	Soil
Eurofins Sample No.			K22- No0015979	K22- No0015980	K22- No0015981	K22- No0015982
Date Sampled			Nov 02, 2022	Nov 02, 2022	Nov 02, 2022	Nov 02, 2022
Test/Reference	LOR	Unit				
Polycyclic Aromatic Hydrocarbons (NZ MfE)						
Acenaphthene	0.03	mg/kg	< 0.03	0.61	_	_
Acenaphthylene	0.03	mg/kg	< 0.03	0.43	_	_
Anthracene	0.03	mg/kg	0.07	3.1	_	_
Benz(a)anthracene	0.03	mg/kg	0.48	5.4	_	_
Benzo(a)pyrene	0.03	mg/kg	0.58	4.5	-	-
Benzo(a)pyrene TEQ (lower bound)*	0.03	mg/kg	0.79	6.4	-	-
Benzo(a)pyrene TEQ (medium bound)*	0.03	mg/kg	0.79	6.4	_	-
Benzo(a)pyrene TEQ (upper bound)*	0.03	mg/kg	0.79	6.4	_	-
Benzo(b&j)fluoranthene ^{N07}	0.03	mg/kg	0.30	4.0	_	-
Benzo(g.h.i)perylene	0.03	mg/kg	0.25	0.67	_	-
Benzo(k)fluoranthene	0.03	mg/kg	0.26	3.2	-	-
Chrysene	0.03	mg/kg	0.54	4.4	-	-
Dibenz(a.h)anthracene	0.03	mg/kg	0.07	0.35	-	-
Fluoranthene	0.03	mg/kg	0.67	13	-	-
Fluorene	0.03	mg/kg	< 0.03	1.2	_	-
Indeno(1.2.3-cd)pyrene	0.03	mg/kg	0.23	1.0	-	-
Naphthalene	0.1	mg/kg	< 0.1	0.5	-	-
Phenanthrene	0.03	mg/kg	0.10	8.7	-	-
Pyrene	0.03	mg/kg	0.81	13	-	-
p-Terphenyl-d14 (surr.)	1	%	74	79	-	-
2-Fluorobiphenyl (surr.)	1	%	65	71	-	-
Semivolatile Organics						
1-Chloronaphthalene	0.5	mg/kg	-	< 0.5	< 0.5	< 0.5
1-Naphthylamine	0.5	mg/kg	-	< 0.5	< 0.5	< 0.5
1.2-Dichlorobenzene	0.5	mg/kg	-	< 0.5	< 0.5	< 0.5
1.2.3-Trichlorobenzene	0.5	mg/kg	-	< 0.5	< 0.5	< 0.5
1.2.3.4-Tetrachlorobenzene	0.5	mg/kg	-	< 0.5	< 0.5	< 0.5
1.2.3.5-Tetrachlorobenzene	0.5	mg/kg	-	< 0.5	< 0.5	< 0.5
1.2.4-Trichlorobenzene	0.5	mg/kg	-	< 0.5	< 0.5	< 0.5
1.2.4.5-Tetrachlorobenzene	0.5	mg/kg	-	< 0.5	< 0.5	< 0.5
1.3-Dichlorobenzene	0.5	mg/kg	_	< 0.5	< 0.5	< 0.5
1.3.5-Trichlorobenzene	0.5	mg/kg	-	< 0.5	< 0.5	< 0.5
1.4-Dichlorobenzene	0.5	mg/kg	_	< 0.5	< 0.5	< 0.5
2-Chloronaphthalene	0.5	mg/kg	-	< 0.5	< 0.5	< 0.5
2-Chlorophenol	0.5	mg/kg	_	< 0.5	< 0.5	< 0.5
2-Fluorobiphenyl (surr.)	1	%	_	71	INT	56
2-Methyl-4.6-dinitrophenol	5	mg/kg	-	< 5	< 5	< 5
2-Methylnaphthalene	0.5	mg/kg	-	< 0.5	< 0.5	< 0.5
2-Methylphenol (o-Cresol)	0.2	mg/kg	-	< 0.2	< 0.2	< 0.2
2-Naphthylamine	0.5	mg/kg	-	< 0.5	< 0.5	< 0.5
2-Nitroaniline	0.5	mg/kg	-	< 0.5	< 0.5	< 0.5
2-Nitrophenol	1	mg/kg	-	< 1	< 1	< 1
2-Picoline	0.5	mg/kg	-	< 0.5	< 0.5	< 0.5
2.3.4.6-Tetrachlorophenol	5	mg/kg	-	< 5	< 5	< 5
2.4-Dichlorophenol	0.5	mg/kg	-	< 0.5	< 0.5	< 0.5
2.4-Dimethylphenol	0.5	mg/kg	-	< 0.5	< 0.5	< 0.5
2.4-Dinitrophenol	5	mg/kg	-	< 5	< 5	< 5
2.4-Dinitrotoluene	0.5	mg/kg	-	< 0.5	< 0.5	< 0.5
2.4.5-Trichlorophenol	1	mg/kg	-	< 1	< 1	< 1



Client Sample ID			22-40179-5	22-40179-6	22-40179-7	22-40179-8
Sample Matrix			Soil	Soil	Soil	Soil
Eurofins Sample No.			K22- No0015979	K22- No0015980	K22- No0015981	K22- No0015982
Date Sampled			Nov 02, 2022	Nov 02, 2022	Nov 02, 2022	Nov 02, 2022
Test/Reference	LOR	Unit				
Semivolatile Organics						
2.4.6-Tribromophenol (surr.)	1	%	_	93	71	53
2.4.6-Trichlorophenol	1	mg/kg	_	< 1	< 1	< 1
2.6-Dichlorophenol	0.5	mg/kg	_	< 0.5	< 0.5	< 0.5
2.6-Dinitrotoluene	0.5	mg/kg	-	< 0.5	< 0.5	< 0.5
3&4-Methylphenol (m&p-Cresol)	0.4	mg/kg	_	< 0.4	< 0.4	< 0.4
3-Methylcholanthrene	0.5	mg/kg	-	< 0.5	< 0.5	< 0.5
3.3'-Dichlorobenzidine	0.5	mg/kg	_	< 0.5	< 0.5	< 0.5
4-Aminobiphenyl	0.5	mg/kg	_	< 0.5	< 0.5	< 0.5
4-Bromophenyl phenyl ether	0.5	mg/kg	_	< 0.5	< 0.5	< 0.5
4-Chloro-3-methylphenol	1	mg/kg	-	< 1	< 1	< 1
4-Chlorophenyl phenyl ether	0.5	mg/kg	-	< 0.5	< 0.5	< 0.5
4-Nitrophenol	5	mg/kg	-	< 5	< 5	< 5
4.4'-DDD	0.01	mg/kg	-	0.01	0.01	0.04
4.4'-DDE	0.01	mg/kg	-	< 0.01	< 0.01	0.04
4.4'-DDT	0.01	mg/kg	-	< 0.01	< 0.01	0.13
7.12-Dimethylbenz(a)anthracene	0.5	mg/kg	_	< 0.5	< 0.5	< 0.5
a-HCH	0.01	mg/kg	-	< 0.01	< 0.01	< 0.01
Acenaphthene	0.03	mg/kg	-	0.61	< 0.03	< 0.03
Acenaphthylene	0.03	mg/kg	-	0.43	< 0.03	< 0.03
Acetophenone	0.5	mg/kg	-	< 0.5	< 0.5	< 0.5
Aldrin	0.01	mg/kg	-	< 0.01	< 0.01	< 0.01
Aniline	0.5	mg/kg	-	< 0.5	< 0.5	< 0.5
Anthracene	0.03	mg/kg	-	3.1	0.12	< 0.03
b-HCH	0.01	mg/kg	-	< 0.01	< 0.01	< 0.01
Benz(a)anthracene	0.03	mg/kg	-	5.4	0.22	0.08
Benzo(a)pyrene	0.03	mg/kg	-	4.5	0.28	0.09
Benzo(a)pyrene TEQ (lower bound)*	0.03	mg/kg	-	6.4	0.36	0.12
Benzo(a)pyrene TEQ (medium bound)*	0.03	mg/kg	-	6.4	0.37	0.13
Benzo(a)pyrene TEQ (upper bound)*	0.03	mg/kg	-	6.4	0.39	0.15
Benzo(b&j)fluoranthene ^{N07}	0.03	mg/kg	-	4.0	0.22	0.05
Benzo(g.h.i)perylene	0.03	mg/kg	-	0.67	0.13	0.08
Benzo(k)fluoranthene	0.03	mg/kg	-	3.2	0.15	0.06
Benzyl chloride	0.5	mg/kg	-	< 0.5	< 0.5	< 0.5
Bis(2-chloroethoxy)methane	0.5	mg/kg	-	< 0.5	< 0.5	< 0.5
Bis(2-chloroisopropyl)ether	0.5	mg/kg	-	< 0.5	< 0.5	< 0.5
Bis(2-ethylhexyl)phthalate	0.5	mg/kg	-	< 0.5	< 0.5	< 0.5
Butyl benzyl phthalate	0.5	mg/kg	-	< 0.5	< 0.5	< 0.5
Chrysene	0.03	mg/kg	-	4.4	0.24	0.12
d-HCH	0.01	mg/kg	-	< 0.01	< 0.01	< 0.01
Di-n-butyl phthalate	0.5	mg/kg	-	< 0.5	< 0.5	< 0.5
Di-n-octyl phthalate	0.5	mg/kg	-	< 0.5	< 0.5	< 0.5
Dibenz(a.h)anthracene	0.03	mg/kg	-	0.35	< 0.03	< 0.03
Dibenz(a.j)acridine	0.5	mg/kg	-	< 0.5	< 0.5	< 0.5
Dibenzofuran	0.5	mg/kg	-	0.5	< 0.5	< 0.5
Dieldrin	0.01	mg/kg	-	< 0.01	< 0.01	< 0.01
Diethyl phthalate	0.5	mg/kg	-	< 0.5	< 0.5	< 0.5
Dimethyl phthalate	0.5	mg/kg	-	< 0.5	< 0.5	< 0.5
Dimethylaminoazobenzene	0.5	mg/kg	-	< 0.5	< 0.5	< 0.5
Diphenylamine	0.5	mg/kg	-	< 0.5	< 0.5	< 0.5



Client Sample ID			22-40179-5	22-40179-6	22-40179-7	22-40179-8
Sample Matrix			Soil	Soil	Soil	Soil
Eurofins Sample No.			K22- No0015979	K22- No0015980	K22- No0015981	K22- No0015982
Date Sampled			Nov 02, 2022	Nov 02, 2022	Nov 02, 2022	Nov 02, 2022
•	LOD	l lasia	1407 02, 2022	1407 02, 2022	1407 02, 2022	1400 02, 2022
Test/Reference Semivolatile Organics	LOR	Unit				
•	0.04			0.04	0.04	0.04
Endosulfan I	0.01	mg/kg	-	< 0.01	< 0.01	< 0.01
Endosulfan II	0.01	mg/kg	-	< 0.01	< 0.01	< 0.01
Endosulfan sulphate	0.01	mg/kg	-	< 0.01	< 0.01	< 0.01
Endrin	0.01	mg/kg	-	< 0.01	< 0.01	< 0.01
Endrin aldehyde	0.01	mg/kg	-	< 0.01	< 0.01	< 0.01
Endrin ketone	0.01	mg/kg	-	< 0.01	< 0.01	< 0.01
Fluoranthene	0.03	mg/kg	-	13	0.51	0.14
Fluorene	0.03	mg/kg	-	1.2	0.05	< 0.03
g-HCH (Lindane)	0.01	mg/kg	-	< 0.01	< 0.01	< 0.01
Heptachlor	0.01	mg/kg	-	< 0.01	< 0.01	< 0.01
Heptachlor epoxide	0.01	mg/kg	-	< 0.01	< 0.01	< 0.01
Hexachlorobenzene	0.01	mg/kg	-	< 0.01	< 0.01	< 0.01
Hexachlorobutadiene	0.5	mg/kg	-	< 0.5	< 0.5	< 0.5
Hexachlorocyclopentadiene	0.5	mg/kg	-	< 0.5	< 0.5	< 0.5
Hexachloroethane	0.5	mg/kg	-	< 0.5	< 0.5	< 0.5
Indeno(1.2.3-cd)pyrene	0.03	mg/kg	-	1.0	0.12	0.05
Methoxychlor	0.01	mg/kg	-	< 0.01	< 0.01	< 0.01
N-Nitrosodibutylamine	0.5	mg/kg	-	< 0.5	< 0.5	< 0.5
N-Nitrosodipropylamine	0.5	mg/kg	-	< 0.5	< 0.5	< 0.5
N-Nitrosopiperidine	0.5	mg/kg	-	< 0.5	< 0.5	< 0.5
Naphthalene	0.1	mg/kg	-	0.5	< 0.1	< 0.1
Nitrobenzene	0.5	mg/kg	-	< 0.5	< 0.5	< 0.5
Nitrobenzene-d5 (surr.)	1	%	-	62	INT	61
Pentachlorobenzene	0.5	mg/kg	-	< 0.5	< 0.5	< 0.5
Pentachloronitrobenzene	0.5	mg/kg	-	< 0.5	< 0.5	< 0.5
Pentachlorophenol	1	mg/kg	-	< 1	< 1	< 1
Phenanthrene	0.03	mg/kg	-	8.7	0.31	0.04
Phenol	0.5	mg/kg	-	< 0.5	< 0.5	< 0.5
Phenol-d6 (surr.)	1	%	-	57	51	57
Pronamide	0.5	mg/kg	_	< 0.5	< 0.5	< 0.5
Pyrene	0.03	mg/kg	-	13	0.53	0.13
Trifluralin	0.5	mg/kg	_	< 0.5	< 0.5	< 0.5
Timurumi		i iiig/ikg		V 0.0	V 0.0	
% Moisture	1	%	31	8.9	22	22

Client Sample ID			22-40179-9	22-40179-12	22-40179-13	22-40179-14
Sample Matrix			Soil	Soil	Soil	Soil
Eurofins Sample No.			K22- No0015983	K22- No0015984	K22- No0015985	K22- No0015986
Date Sampled			Nov 02, 2022	Nov 02, 2022	Nov 02, 2022	Nov 02, 2022
Test/Reference	LOR	Unit				
Polycyclic Aromatic Hydrocarbons (NZ MfE)						
Acenaphthene	0.03	mg/kg	-	< 0.03	< 0.03	< 0.03
Acenaphthylene	0.03	mg/kg	-	0.04	< 0.03	< 0.03
Anthracene	0.03	mg/kg	-	0.04	< 0.03	< 0.03
Benz(a)anthracene	0.03	mg/kg	-	0.17	< 0.03	0.10
Benzo(a)pyrene	0.03	mg/kg	-	0.23	< 0.03	0.10
Benzo(a)pyrene TEQ (lower bound)*	0.03	mg/kg	-	0.29	< 0.03	0.13



Client Sample ID			22-40179-9	22-40179-12	22-40179-13	22-40179-14
Sample Matrix			Soil	Soil	Soil	Soil
			K22-	K22-	K22-	K22-
Eurofins Sample No.			No0015983	No0015984	No0015985	No0015986
Date Sampled			Nov 02, 2022	Nov 02, 2022	Nov 02, 2022	Nov 02, 2022
Test/Reference	LOR	Unit				
Polycyclic Aromatic Hydrocarbons (NZ MfE)						
Benzo(a)pyrene TEQ (medium bound)*	0.03	mg/kg	-	0.31	0.04	0.15
Benzo(a)pyrene TEQ (upper bound)*	0.03	mg/kg	-	0.32	0.08	0.16
Benzo(b&j)fluoranthene ^{N07}	0.03	mg/kg	-	0.13	< 0.03	0.08
Benzo(g.h.i)perylene	0.03	mg/kg	-	0.08	< 0.03	0.06
Benzo(k)fluoranthene	0.03	mg/kg	-	0.14	< 0.03	0.09
Chrysene	0.03	mg/kg	-	0.18	< 0.03	0.10
Dibenz(a.h)anthracene	0.03	mg/kg	-	< 0.03	< 0.03	< 0.03
Fluoranthene	0.03	mg/kg	-	0.34	< 0.03	0.08
Fluorene	0.03	mg/kg	-	< 0.03	< 0.03	< 0.03
Indeno(1.2.3-cd)pyrene	0.03	mg/kg	-	0.11	< 0.03	0.06
Naphthalene	0.1	mg/kg	-	< 0.1	< 0.1	< 0.1
Phenanthrene	0.03	mg/kg	-	0.13	< 0.03	0.06
Pyrene	0.03	mg/kg	-	0.33	< 0.03	0.08
p-Terphenyl-d14 (surr.)	1	%	-	81	94	101
2-Fluorobiphenyl (surr.)	1	%	-	68	62	70
Semivolatile Organics						
1-Chloronaphthalene	0.5	mg/kg	< 0.5	< 0.5	< 0.5	< 0.5
1-Naphthylamine	0.5	mg/kg	< 0.5	< 0.5	< 0.5	< 0.5
1.2-Dichlorobenzene	0.5	mg/kg	< 0.5	< 0.5	< 0.5	< 0.5
1.2.3-Trichlorobenzene	0.5	mg/kg	< 0.5	< 0.5	< 0.5	< 0.5
1.2.3.4-Tetrachlorobenzene	0.5	mg/kg	< 0.5	< 0.5	< 0.5	< 0.5
1.2.3.5-Tetrachlorobenzene	0.5	mg/kg	< 0.5	< 0.5	< 0.5	< 0.5
1.2.4-Trichlorobenzene	0.5	mg/kg	< 0.5	< 0.5	< 0.5	< 0.5
1.2.4.5-Tetrachlorobenzene	0.5	mg/kg	< 0.5	< 0.5	< 0.5	< 0.5
1.3-Dichlorobenzene	0.5	mg/kg	< 0.5	< 0.5	< 0.5	< 0.5
1.3.5-Trichlorobenzene	0.5	mg/kg	< 0.5	< 0.5	< 0.5	< 0.5
1.4-Dichlorobenzene	0.5	mg/kg	< 0.5	< 0.5	< 0.5	< 0.5
2-Chloronaphthalene	0.5	mg/kg	< 0.5	< 0.5	< 0.5	< 0.5
2-Chlorophenol	0.5	mg/kg	< 0.5	< 0.5	< 0.5	< 0.5
2-Fluorobiphenyl (surr.)	1	%	67	68	62	70
2-Methyl-4.6-dinitrophenol	5	mg/kg	< 5	< 5	< 5	< 5
2-Methylnaphthalene	0.5	mg/kg	< 0.5	< 0.5	< 0.5	< 0.5
2-Methylphenol (o-Cresol)	0.2	mg/kg	< 0.2	< 0.2	< 0.2	< 0.2
2-Naphthylamine	0.5	mg/kg	< 0.5	< 0.5	< 0.5	< 0.5
2-Nitroaniline	0.5	mg/kg	< 0.5	< 0.5	< 0.5	< 0.5
2-Nitrophenol	1	mg/kg	< 1	< 1	< 1	< 1
2-Picoline	0.5	mg/kg	< 0.5	< 0.5	< 0.5	< 0.5
2.3.4.6-Tetrachlorophenol	5	mg/kg	< 5	< 5	< 5	< 5
2.4-Dichlorophenol	0.5	mg/kg	< 0.5	< 0.5	< 0.5	< 0.5
2.4-Dimethylphenol	0.5	mg/kg	< 0.5	< 0.5	< 0.5	< 0.5
2.4-Dinitrophenol	5	mg/kg	< 5	< 5	< 5	< 5
2.4-Dinitrotoluene	0.5	mg/kg	< 0.5	< 0.5	< 0.5	< 0.5
2.4.5-Trichlorophenol	1	mg/kg	< 1	< 1	< 1	< 1
2.4.6-Tribromophenol (surr.)	1	%	92	69	58	INT
2.4.6-Trichlorophenol	1	mg/kg	< 1	< 1	< 1	< 1
2.6-Dichlorophenol	0.5	mg/kg	< 0.5	< 0.5	< 0.5	< 0.5
2.6-Dinitrotoluene	0.5	mg/kg	< 0.5	< 0.5	< 0.5	< 0.5
3&4-Methylphenol (m&p-Cresol)	0.4	mg/kg	< 0.4	< 0.4	< 0.4	< 0.4
3-Methylcholanthrene	0.5	mg/kg	< 0.5	< 0.5	< 0.5	< 0.5

Report Number: 938712-S



Olivery Occurring ID			T	T	T	T
Client Sample ID			22-40179-9	22-40179-12	22-40179-13	22-40179-14
Sample Matrix			Soil	Soil	Soil	Soil
Eurofins Sample No.			K22- No0015983	K22- No0015984	K22- No0015985	K22- No0015986
Date Sampled			Nov 02, 2022	Nov 02, 2022	Nov 02, 2022	Nov 02, 2022
Test/Reference	LOR	Unit				
Semivolatile Organics	•	•				
3.3'-Dichlorobenzidine	0.5	mg/kg	< 0.5	< 0.5	< 0.5	< 0.5
4-Aminobiphenyl	0.5	mg/kg	< 0.5	< 0.5	< 0.5	< 0.5
4-Bromophenyl phenyl ether	0.5	mg/kg	< 0.5	< 0.5	< 0.5	< 0.5
4-Chloro-3-methylphenol	1	mg/kg	< 1	< 1	< 1	< 1
4-Chlorophenyl phenyl ether	0.5	mg/kg	< 0.5	< 0.5	< 0.5	< 0.5
4-Nitrophenol	5	mg/kg	< 5	< 5	< 5	< 5
4.4'-DDD	0.01	mg/kg	0.04	< 0.01	< 0.01	< 0.01
4.4'-DDE	0.01	mg/kg	< 0.01	< 0.01	< 0.01	< 0.01
4.4'-DDT	0.01	mg/kg	0.06	< 0.01	< 0.01	< 0.01
7.12-Dimethylbenz(a)anthracene	0.5	mg/kg	< 0.5	< 0.5	< 0.5	< 0.5
a-HCH	0.01	mg/kg	< 0.01	< 0.01	< 0.01	< 0.01
Acenaphthene	0.03	mg/kg	< 0.03	< 0.03	< 0.03	< 0.03
Acenaphthylene	0.03	mg/kg	< 0.03	0.04	< 0.03	< 0.03
Acetophenone	0.5	mg/kg	< 0.5	< 0.5	< 0.5	< 0.5
Aldrin	0.01	mg/kg	< 0.01	< 0.01	< 0.01	< 0.01
Aniline	0.5	mg/kg	< 0.5	< 0.5	< 0.5	< 0.5
Anthracene	0.03	mg/kg	< 0.03	0.04	< 0.03	< 0.03
b-HCH	0.01	mg/kg	< 0.01	< 0.01	< 0.01	< 0.01
Benz(a)anthracene	0.03	mg/kg	0.12	0.17	< 0.03	0.10
Benzo(a)pyrene	0.03	mg/kg	0.20	0.23	< 0.03	0.10
Benzo(a)pyrene TEQ (lower bound)*	0.03	mg/kg	0.25	0.29	< 0.03	0.13
Benzo(a)pyrene TEQ (medium bound)*	0.03	mg/kg	0.27	0.31	0.04	0.15
Benzo(a)pyrene TEQ (upper bound)*	0.03	mg/kg	0.28	0.32	0.08	0.16
Benzo(b&j)fluoranthene ^{N07}	0.03	mg/kg	0.12	0.13	< 0.03	0.08
Benzo(g.h.i)perylene	0.03	mg/kg	0.12	0.08	< 0.03	0.06
Benzo(k)fluoranthene	0.03	mg/kg	0.14	0.14	< 0.03	0.09
Benzyl chloride	0.5	mg/kg	< 0.5	< 0.5	< 0.5	< 0.5
Bis(2-chloroethoxy)methane	0.5	mg/kg	< 0.5	< 0.5	< 0.5	< 0.5
Bis(2-chloroisopropyl)ether	0.5	mg/kg	< 0.5	< 0.5	< 0.5	< 0.5
Bis(2-ethylhexyl)phthalate	0.5	mg/kg	< 0.5	< 0.5	< 0.5	< 0.5
Butyl benzyl phthalate	0.5	mg/kg	< 0.5	< 0.5	< 0.5	< 0.5
Chrysene	0.03	mg/kg	0.22	0.18	< 0.03	0.10
d-HCH	0.01	mg/kg	< 0.01	< 0.01	< 0.01	< 0.01
Di-n-butyl phthalate	0.5	mg/kg	< 0.5	< 0.5	< 0.5	< 0.5
Di-n-octyl phthalate	0.5	mg/kg	< 0.5	< 0.5	< 0.5	< 0.5
Dibenz(a.h)anthracene	0.03	mg/kg	< 0.03	< 0.03	< 0.03	< 0.03
Dibenz(a.j)acridine	0.5	mg/kg	< 0.5	< 0.5	< 0.5	< 0.5
Dibenzofuran	0.5	mg/kg	< 0.5	< 0.5	< 0.5	< 0.5
Dieldrin	0.01	mg/kg	< 0.01	< 0.01	< 0.01	< 0.01
Diethyl phthalate	0.5	mg/kg	< 0.5	< 0.5	< 0.5	< 0.5
Dimethyl phthalate	0.5	mg/kg	< 0.5	< 0.5	< 0.5	< 0.5
Dimethylaminoazobenzene	0.5	mg/kg	< 0.5	< 0.5	< 0.5	< 0.5
Diphenylamine	0.5	mg/kg	< 0.5	< 0.5	< 0.5	< 0.5
Endosulfan I	0.01	mg/kg	< 0.01	< 0.01	< 0.01	< 0.01
Endosulfan II	0.01	mg/kg	< 0.01	< 0.01	< 0.01	< 0.01
Endosulfan sulphate	0.01	mg/kg	< 0.01	< 0.01	< 0.01	< 0.01
Endrin	0.01	mg/kg	< 0.01	< 0.01	< 0.01	< 0.01
Endrin aldehyde	0.01	mg/kg	< 0.01	< 0.01	< 0.01	< 0.01
Endrin ketone	0.01	mg/kg	< 0.01	< 0.01	< 0.01	< 0.01

Report Number: 938712-S



Client Sample ID			22-40179-9	22-40179-12	22-40179-13	22-40179-14
Sample Matrix			Soil	Soil	Soil	Soil
Eurofins Sample No.			K22- No0015983	K22- No0015984	K22- No0015985	K22- No0015986
Date Sampled			Nov 02, 2022	Nov 02, 2022	Nov 02, 2022	Nov 02, 2022
Test/Reference	LOR	Unit				
Semivolatile Organics	•	•				
Fluoranthene	0.03	mg/kg	0.30	0.34	< 0.03	0.08
Fluorene	0.03	mg/kg	< 0.03	< 0.03	< 0.03	< 0.03
g-HCH (Lindane)	0.01	mg/kg	< 0.01	< 0.01	< 0.01	< 0.01
Heptachlor	0.01	mg/kg	< 0.01	< 0.01	< 0.01	< 0.01
Heptachlor epoxide	0.01	mg/kg	< 0.01	< 0.01	< 0.01	< 0.01
Hexachlorobenzene	0.01	mg/kg	< 0.01	< 0.01	< 0.01	< 0.01
Hexachlorobutadiene	0.5	mg/kg	< 0.5	< 0.5	< 0.5	< 0.5
Hexachlorocyclopentadiene	0.5	mg/kg	< 0.5	< 0.5	< 0.5	< 0.5
Hexachloroethane	0.5	mg/kg	< 0.5	< 0.5	< 0.5	< 0.5
Indeno(1.2.3-cd)pyrene	0.03	mg/kg	0.10	0.11	< 0.03	0.06
Methoxychlor	0.01	mg/kg	< 0.01	< 0.01	< 0.01	< 0.01
N-Nitrosodibutylamine	0.5	mg/kg	< 0.5	< 0.5	< 0.5	< 0.5
N-Nitrosodipropylamine	0.5	mg/kg	< 0.5	< 0.5	< 0.5	< 0.5
N-Nitrosopiperidine	0.5	mg/kg	< 0.5	< 0.5	< 0.5	< 0.5
Naphthalene	0.1	mg/kg	< 0.1	< 0.1	< 0.1	< 0.1
Nitrobenzene	0.5	mg/kg	< 0.5	< 0.5	< 0.5	< 0.5
Nitrobenzene-d5 (surr.)	1	%	72	61	56	60
Pentachlorobenzene	0.5	mg/kg	< 0.5	< 0.5	< 0.5	< 0.5
Pentachloronitrobenzene	0.5	mg/kg	< 0.5	< 0.5	< 0.5	< 0.5
Pentachlorophenol	1	mg/kg	< 1	< 1	< 1	< 1
Phenanthrene	0.03	mg/kg	0.12	0.13	< 0.03	0.06
Phenol	0.5	mg/kg	< 0.5	< 0.5	< 0.5	< 0.5
Phenol-d6 (surr.)	1	%	65	57	71	70
Pronamide	0.5	mg/kg	< 0.5	< 0.5	< 0.5	< 0.5
Pyrene	0.03	mg/kg	0.34	0.33	< 0.03	0.08
Trifluralin	0.5	mg/kg	< 0.5	< 0.5	< 0.5	< 0.5
	T.				1	
% Moisture	1	%	50	17	32	15

Client Sample ID Sample Matrix Eurofins Sample No.			22-40179-15 Soil K22- No0015987	22-40179-16 Soil K22- No0015988
Date Sampled			Nov 02, 2022	Nov 02, 2022
Test/Reference	LOR	Unit		
Semivolatile Organics	·			
1-Chloronaphthalene	0.5	mg/kg	< 0.5	< 0.5
1-Naphthylamine	0.5	mg/kg	< 0.5	< 0.5
1.2-Dichlorobenzene	0.5	mg/kg	< 0.5	< 0.5
1.2.3-Trichlorobenzene	0.5	mg/kg	< 0.5	< 0.5
1.2.3.4-Tetrachlorobenzene	0.5	mg/kg	< 0.5	< 0.5
1.2.3.5-Tetrachlorobenzene	0.5	mg/kg	< 0.5	< 0.5
1.2.4-Trichlorobenzene	0.5	mg/kg	< 0.5	< 0.5
1.2.4.5-Tetrachlorobenzene	0.5	mg/kg	< 0.5	< 0.5
1.3-Dichlorobenzene	0.5	mg/kg	< 0.5	< 0.5
1.3.5-Trichlorobenzene	0.5	mg/kg	< 0.5	< 0.5
1.4-Dichlorobenzene	0.5	mg/kg	< 0.5	< 0.5
2-Chloronaphthalene	0.5	mg/kg	< 0.5	< 0.5



Client Sample ID			22-40179-15	22-40179-16
Sample Matrix			Soil	Soil
Eurofins Sample No.			K22- No0015987	K22- No0015988
Date Sampled			Nov 02, 2022	Nov 02, 2022
Test/Reference	LOR	Unit		
Semivolatile Organics		J 0		
2-Chlorophenol	0.5	mg/kg	< 0.5	< 0.5
2-Fluorobiphenyl (surr.)	1	%	75	65
2-Methyl-4.6-dinitrophenol	5	mg/kg	< 5	< 5
2-Methylnaphthalene	0.5	mg/kg	< 0.5	< 0.5
2-Methylphenol (o-Cresol)	0.2	mg/kg	< 0.2	< 0.2
2-Naphthylamine	0.5	mg/kg	< 0.5	< 0.5
2-Nitroaniline	0.5	mg/kg	< 0.5	< 0.5
2-Nitrophenol	1	mg/kg	< 1	< 1
2-Picoline	0.5	mg/kg	< 0.5	< 0.5
2.3.4.6-Tetrachlorophenol	5	mg/kg	< 5	< 5
2.4-Dichlorophenol	0.5	mg/kg	< 0.5	< 0.5
2.4-Dimethylphenol	0.5	mg/kg	< 0.5	< 0.5
2.4-Dinitrophenol	5	mg/kg	< 5	< 5
2.4-Dinitrotoluene	0.5	mg/kg	< 0.5	< 0.5
2.4.5-Trichlorophenol	1	mg/kg	< 1	< 1
2.4.6-Tribromophenol (surr.)	1	%	92	64
2.4.6-Trichlorophenol	1	mg/kg	< 1	< 1
2.6-Dichlorophenol	0.5	mg/kg	< 0.5	< 0.5
2.6-Dinitrotoluene	0.5	mg/kg	< 0.5	< 0.5
3&4-Methylphenol (m&p-Cresol)	0.4	mg/kg	< 0.4	< 0.4
3-Methylcholanthrene	0.5	mg/kg	< 0.5	< 0.5
3.3'-Dichlorobenzidine	0.5	mg/kg	< 0.5	< 0.5
4-Aminobiphenyl	0.5	mg/kg	< 0.5	< 0.5
4-Bromophenyl phenyl ether	0.5	mg/kg	< 0.5	< 0.5
4-Chloro-3-methylphenol	1	mg/kg	< 1	< 1
4-Chlorophenyl phenyl ether	0.5	mg/kg	< 0.5	< 0.5
4-Nitrophenol	5	mg/kg	< 5	< 5
4.4'-DDD	0.01	mg/kg	< 0.01	< 0.01
4.4'-DDE	0.01	mg/kg	< 0.01	< 0.01
4.4'-DDT	0.01	mg/kg	< 0.01	< 0.01
7.12-Dimethylbenz(a)anthracene	0.5	mg/kg	< 0.5	< 0.5
a-HCH	0.01	mg/kg	< 0.01	< 0.01
Acenaphthene	0.03	mg/kg	< 0.03	< 0.03
Acenaphthylene	0.03	mg/kg	< 0.03	< 0.03
Acetophenone	0.5	mg/kg	< 0.5	< 0.5
Aldrin	0.01	mg/kg	< 0.01	< 0.01
Anthropono	0.5	mg/kg	< 0.5	< 0.5
Anthracene	0.03	mg/kg	< 0.03	< 0.03
b-HCH Ronz/a)anthracene	0.01	mg/kg	< 0.01	< 0.01
Benzo(a)pyrene	0.03	mg/kg	0.19	< 0.03
Benzo(a)pyrene Benzo(a)pyrene TEQ (lower bound)*	0.03	mg/kg	0.30 0.46	< 0.03 < 0.03
Benzo(a)pyrene TEQ (nedium bound)*	0.03	mg/kg mg/kg	0.46	0.04
Benzo(a)pyrene TEQ (medium bound)*	0.03	mg/kg	0.46	0.04
Benzo(b&j)fluoranthene ^{N07}	0.03	mg/kg	0.46	< 0.08
Benzo(g.h.i)perylene	0.03	mg/kg	0.27	< 0.03
Benzo(k)fluoranthene	0.03	mg/kg	0.25	< 0.03
Benzyl chloride	0.05	mg/kg	< 0.5	< 0.05
Bis(2-chloroethoxy)methane	0.5	mg/kg	< 0.5	< 0.5



Client Sample ID			22-40179-15	22-40179-16
Sample Matrix			Soil	Soil
Eurofins Sample No.			K22- No0015987	K22- No0015988
Date Sampled			Nov 02, 2022	Nov 02, 2022
Test/Reference	LOR	Unit		
Semivolatile Organics				
Bis(2-chloroisopropyl)ether	0.5	mg/kg	< 0.5	< 0.5
Bis(2-ethylhexyl)phthalate	0.5	mg/kg	< 0.5	< 0.5
Butyl benzyl phthalate	0.5	mg/kg	< 0.5	< 0.5
Chrysene	0.03	mg/kg	0.27	< 0.03
d-HCH	0.01	mg/kg	< 0.01	< 0.01
Di-n-butyl phthalate	0.5	mg/kg	< 0.5	< 0.5
Di-n-octyl phthalate	0.5	mg/kg	< 0.5	< 0.5
Dibenz(a.h)anthracene	0.03	mg/kg	0.06	< 0.03
Dibenz(a.j)acridine	0.5	mg/kg	< 0.5	< 0.5
Dibenzofuran	0.5	mg/kg	< 0.5	< 0.5
Dieldrin	0.01	mg/kg	< 0.01	< 0.01
Diethyl phthalate	0.5	mg/kg	< 0.5	< 0.5
Dimethyl phthalate	0.5	mg/kg	< 0.5	< 0.5
Dimethylaminoazobenzene	0.5	mg/kg	< 0.5	< 0.5
Diphenylamine	0.5	mg/kg	< 0.5	< 0.5
Endosulfan I	0.01	mg/kg	< 0.01	< 0.01
Endosulfan II	0.01	mg/kg	< 0.01	< 0.01
Endosulfan sulphate	0.01	mg/kg	< 0.01	< 0.01
Endrin	0.01	mg/kg	< 0.01	< 0.01
Endrin aldehyde	0.01	mg/kg	< 0.01	< 0.01
Endrin ketone	0.01	mg/kg	< 0.01	< 0.01
Fluoranthene	0.03	mg/kg	0.30	< 0.03
Fluorene	0.03	mg/kg	< 0.03	< 0.03
g-HCH (Lindane)	0.01	mg/kg	< 0.01	< 0.01
Heptachlor	0.01	mg/kg	< 0.01	< 0.01
Heptachlor epoxide	0.01	mg/kg	< 0.01	< 0.01
Hexachlorobenzene	0.01	mg/kg	< 0.01	< 0.01
Hexachlorobutadiene	0.5	mg/kg	< 0.5	< 0.5
Hexachlorocyclopentadiene	0.5	mg/kg	< 0.5	< 0.5
Hexachloroethane	0.5	mg/kg	< 0.5	< 0.5
Indeno(1.2.3-cd)pyrene	0.03	mg/kg	0.18	< 0.03
Methoxychlor	0.01	mg/kg	< 0.01	< 0.01
N-Nitrosodibutylamine	0.5	mg/kg	< 0.5	< 0.5
N-Nitrosodipropylamine	0.5	mg/kg	< 0.5	< 0.5
N-Nitrosopiperidine	0.5	mg/kg	< 0.5	< 0.5
Naphthalene	0.1	mg/kg	< 0.1	< 0.1
Nitrobenzene	0.5	mg/kg	< 0.5	< 0.5
Nitrobenzene-d5 (surr.)	1	%	63	63
Pentachlorobenzene	0.5	mg/kg	< 0.5	< 0.5
Pentachloronitrobenzene	0.5	mg/kg	< 0.5	< 0.5
Pentachlorophenol	1	mg/kg	< 1	< 1
Phenanthrene	0.03	mg/kg	0.13	< 0.03
Phenol	0.5	mg/kg	< 0.5	< 0.5
Phenol-d6 (surr.)	1	%	73	INT
Pronamide	0.5	mg/kg	< 0.5	< 0.5
Pyrene	0.03	mg/kg	0.35	< 0.03
Trifluralin	0.5	mg/kg	< 0.5	< 0.5
		%		



Sample History

Where samples are submitted/analysed over several days, the last date of extraction is reported.

If the date and time of sampling are not provided, the Laboratory will not be responsible for compromised results should testing be performed outside the recommended holding time.

Description	Testing Site	Extracted	Holding Time
Polycyclic Aromatic Hydrocarbons (NZ MfE)	Auckland	Nov 08, 2022	14 Days
- Method: LTM-ORG-2130 PAH and Phenols in Soil and Water by GC MSMS			
Semivolatile Organics	Auckland	Nov 08, 2022	14 Days
- Method: LTM-ORG-2190 SVOC in Water & Soil by GC-MS			
% Moisture	Auckland	Nov 08, 2022	14 Days

⁻ Method: LTM-GEN-7080 Moisture Content in Soil by Gravimetry



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Eurofins ARL Pty Ltd

Company Name:

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Analytica Laboratories LTD

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Order No.: Report #: Phone:

938712 (07) 9744740

22-40179

Fax:

Received: Nov 8, 2022 12:19 PM Due: Nov 15, 2022

Priority: 5 Day

Contact Name: Customer Service

Project Name:

Project ID: 22-40179

Eurofins Analytical Services Manager: Karishma Patel

		Sa	imple Detail			Moisture Set	Polycyclic Aromatic Hydrocarbons (NZ MfE)	Semivolatile Organics
Aucl	kland Laborato	ry - IANZ# 1327				Х	Х	Х
Chri	stchurch Labor	atory - IANZ# 1	290					
Exte	rnal Laboratory	/		1				
No	Sample ID	Sample Date	Sampling Time	Matrix	LAB ID			
1	22-40179-1	Nov 02, 2022		Soil	K22-No0015975	Х	Х	Х
2	22-40179-2	Nov 02, 2022		Soil	K22-No0015976	Х	Х	Х
3	22-40179-3	Nov 02, 2022		Soil	K22-No0015977	Х	Х	
4	22-40179-4	Nov 02, 2022		Soil	K22-No0015978	Х	Х	
5	22-40179-5	Nov 02, 2022		Soil	K22-No0015979	Х	Х	
6	22-40179-6	Nov 02, 2022		Soil	K22-No0015980	Х	Х	Х
7	22-40179-7	Nov 02, 2022		Soil	K22-No0015981	Х		Х
8	22-40179-8	Nov 02, 2022		Soil	K22-No0015982	Х		Х
9	22-40179-9	Nov 02, 2022		Soil	K22-No0015983	Х		Х
10	22-40179-12	Nov 02, 2022		Soil	K22-No0015984	Х	Х	Х
11	22-40179-13	Nov 02, 2022		Soil	K22-No0015985	Х	Х	Х
12	22-40179-14	Nov 02, 2022		Soil	K22-No0015986	Х	Х	Х



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Received: Nov 8, 2022 12:19 PM Due: Nov 15, 2022

Priority: 5 Day

Contact Name: Customer Service

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		Sa	mple Detail			Moisture Set	Polycyclic Aromatic Hydrocarbons (NZ MfE)	Semivolatile Organics
Auck	dand Laborator	y - IANZ# 1327				Χ	Х	Χ
Chris	stchurch Labora	atory - IANZ# 1	290					
Exte	rnal Laboratory							
13	22-40179-15	Nov 02, 2022		Soil	K22-No0015987	Х		Х
14	22-40179-16	Nov 02, 2022		Soil	K22-No0015988	Х		Х
Test	Counts					14	9	11



Internal Quality Control Review and Glossary

General

- Laboratory QC results for Method Blanks, Duplicates, Matrix Spikes, and Laboratory Control Samples follows guidelines delineated in the National Environment Protection (Assessment of Site Contamination) Measure 1999, as amended May 2013 and are included in this QC report where applicable. Additional QC data may be available on request.
- 2. All soil/sediment/solid results are reported on a dry basis, unless otherwise stated.
- 3. All biota/food results are reported on a wet weight basis on the edible portion, unless otherwise stated.
- 4. Actual LORs are matrix dependant. Quoted LORs may be raised where sample extracts are diluted due to interferences.
- 5. Results are uncorrected for matrix spikes or surrogate recoveries except for PFAS compounds.
- 6. SVOC analysis on waters are performed on homogenised, unfiltered samples, unless noted otherwise
- 7. Samples were analysed on an 'as received' basis
- 8. Information identified on this report with blue colour, indicates data provided by customer that may have an impact on the results.
- 9. This report replaces any interim results previously issued.

Holding Times

Please refer to 'Sample Preservation and Container Guide' for holding times (QS3001).

For samples received on the last day of holding time, notification of testing requirements should have been received at least 6 hours prior to sample receipt deadlines as stated on the SRA

If the Laboratory did not receive the information in the required timeframe, and regardless of any other integrity issues, suitably qualified results may still be reported.

Holding times apply from the date of sampling, therefore compliance to these may be outside the laboratory's control.

For VOCs containing vinyl chloride, styrene and 2-chloroethyl vinyl ether the holding time is 7 days however for all other VOCs such as BTEX or C6-10 TRH then the holding time is 14 days.

Units

mg/kg: milligrams per kilogram mg/L: micrograms per litre µg/L: micrograms per litre

ppm: parts per million **ppb**: parts per billion
%: Percentage

org/100 mL: Organisms per 100 millilitres NTU: Nephelometric Turbidity Units MPN/100 mL: Most Probable Number of organisms per 100 millilitres

Terms

APHA American Public Health Association

COC Chain of Custody

CP Client Parent - QC was performed on samples pertaining to this report

CRM Certified Reference Material (ISO17034) - reported as percent recovery.

Dry Where a moisture has been determined on a solid sample the result is expressed on a dry basis

Duplicate A second piece of analysis from the same sample and reported in the same units as the result to show comparison.

LOR Limit of Reporting.

Laboratory Control Sample - reported as percent recovery.

Method Blank In the case of solid samples these are performed on laboratory certified clean sands and in the case of water samples these are performed on de-ionised water.

NCP Non-Client Parent - QC performed on samples not pertaining to this report, QC is representative of the sequence or batch that client samples were analysed within.

RPD Relative Percent Difference between two Duplicate pieces of analysis.

SPIKE Addition of the analyte to the sample and reported as percentage recovery

SRA Sample Receipt Advice

Surr - Surrogate The addition of a like compound to the analyte target and reported as percentage recovery.

TBTO Tributyltin oxide (bis-tributyltin oxide) - individual tributyltin compounds cannot be identified separately in the environment however free tributyltin was measured

and its values were converted stoichiometrically into tributyltin oxide for comparison with regulatory limits.

TCLP Toxicity Characteristic Leaching Procedure
TEQ Toxic Equivalency Quotient or Total Equivalence

QSM US Department of Defense Quality Systems Manual Version 5.4

US EPA United States Environmental Protection Agency

WA DWER Sum of PFBA, PFPeA, PFHxA, PFHpA, PFOA, PFBS, PFHxS, PFOS, 6:2 FTSA, 8:2 FTSA

QC - Acceptance Criteria

The acceptance criteria should be used as a guide only and may be different when site specific Sampling Analysis and Quality Plan (SAQP) have been implemented

RPD Duplicates: Global RPD Duplicates Acceptance Criteria is 30% however the following acceptance guidelines are equally applicable:

Results <10 times the LOR: No Limit

Results between 10-20 times the LOR: RPD must lie between 0-50%

Results >20 times the LOR: RPD must lie between 0-30% NOTE: pH duplicates are reported as a range not as RPD

Surrogate Recoveries: Recoveries must lie between 20-130% for Speciated Phenols & 50-150% for PFAS

PFAS field samples that contain surrogate recoveries in excess of the QC limit designated in QSM 5.4 where no positive PFAS results have been reported have been reviewed and no data was affected.

QC Data General Comments

- 1. Where a result is reported as a less than (<), higher than the nominated LOR, this is due to either matrix interference, extract dilution required due to interferences or contaminant levels within the sample, high moisture content or insufficient sample provided.
- 2. Duplicate data shown within this report that states the word "BATCH" is a Batch Duplicate from outside of your sample batch, but within the laboratory sample batch at a 1:10 ratio. The Parent and Duplicate data shown is not data from your samples.
- 3. pH and Free Chlorine analysed in the laboratory Analysis on this test must begin within 30 minutes of sampling. Therefore, laboratory analysis is unlikely to be completed within holding time. Analysis will begin as soon as possible after sample receipt.
- 4. Recovery Data (Spikes & Surrogates) where chromatographic interference does not allow the determination of recovery the term "INT" appears against that analyte.
- 5. For Matrix Spikes and LCS results a dash "-" in the report means that the specific analyte was not added to the QC sample.
- 6. Duplicate RPDs are calculated from raw analytical data thus it is possible to have two sets of data



Quality Control Results

Test	Units	Result 1	Acceptance Limits	Pass Limits	Qualifying Code
Method Blank					
Polycyclic Aromatic Hydrocarbons (NZ MfE)					
Acenaphthene	mg/kg	< 0.03	0.03	Pass	
Acenaphthylene	mg/kg	< 0.03	0.03	Pass	
Anthracene	mg/kg	< 0.03	0.03	Pass	
Benz(a)anthracene	mg/kg	< 0.03	0.03	Pass	
Benzo(a)pyrene	mg/kg	< 0.03	0.03	Pass	
Benzo(b&j)fluoranthene	mg/kg	< 0.03	0.03	Pass	
Benzo(g.h.i)perylene	mg/kg	< 0.03	0.03	Pass	
Benzo(k)fluoranthene	mg/kg	< 0.03	0.03	Pass	
Chrysene	mg/kg	< 0.03	0.03	Pass	
Dibenz(a.h)anthracene	mg/kg	< 0.03	0.03	Pass	
Fluoranthene	mg/kg	< 0.03	0.03	Pass	
Fluorene	mg/kg	< 0.03	0.03	Pass	
Indeno(1.2.3-cd)pyrene	mg/kg	< 0.03	0.03	Pass	
Naphthalene	mg/kg	< 0.1	0.1	Pass	
Phenanthrene	mg/kg	< 0.03	0.03	Pass	
Pyrene	mg/kg	< 0.03	0.03	Pass	
Method Blank		1 0.00	0.00	1 400	
Semivolatile Organics				Ī	
1-Chloronaphthalene	mg/kg	< 0.5	0.5	Pass	
1-Naphthylamine	mg/kg	< 0.5	0.5	Pass	
1.2-Dichlorobenzene	mg/kg	< 0.5	0.5	Pass	
1.2.3-Trichlorobenzene		< 0.5	0.5	Pass	
1.2.3.4-Tetrachlorobenzene	mg/kg mg/kg	< 0.5	0.5	Pass	
1.2.3.5-Tetrachlorobenzene	mg/kg	< 0.5	0.5	Pass	
1.2.4-Trichlorobenzene		< 0.5	0.5	Pass	
1.2.4.5-Tetrachlorobenzene	mg/kg	< 0.5	0.5	Pass	
	mg/kg	1			
1.3-Dichlorobenzene	mg/kg	< 0.5	0.5	Pass	
1.3.5-Trichlorobenzene	mg/kg	< 0.5	0.5	Pass	
1.4-Dichlorobenzene	mg/kg	< 0.5	0.5	Pass	
2-Chloronaphthalene	mg/kg	< 0.5	0.5	Pass	
2-Chlorophenol	mg/kg	< 0.5	0.5	Pass	
2-Methyl-4.6-dinitrophenol	mg/kg	< 5	5	Pass	
2-Methylnaphthalene	mg/kg	< 0.5	0.5	Pass	
2-Methylphenol (o-Cresol)	mg/kg	< 0.2	0.2	Pass	
2-Naphthylamine	mg/kg	< 0.5	0.5	Pass	
2-Nitroaniline	mg/kg	< 0.5	0.5	Pass	
2-Nitrophenol	mg/kg	< 1	1	Pass	
2-Picoline	mg/kg	< 0.5	0.5	Pass	
2.3.4.6-Tetrachlorophenol	mg/kg	< 5	5	Pass	
2.4-Dichlorophenol	mg/kg	< 0.5	0.5	Pass	
2.4-Dimethylphenol	mg/kg	< 0.5	0.5	Pass	
2.4-Dinitrophenol	mg/kg	< 5	5	Pass	
2.4-Dinitrotoluene	mg/kg	< 0.5	0.5	Pass	
2.4.5-Trichlorophenol	mg/kg	< 1	1	Pass	
2.4.6-Trichlorophenol	mg/kg	< 1	1	Pass	
2.6-Dichlorophenol	mg/kg	< 0.5	0.5	Pass	
2.6-Dinitrotoluene	mg/kg	< 0.5	0.5	Pass	
3&4-Methylphenol (m&p-Cresol)	mg/kg	< 0.4	0.4	Pass	
3-Methylcholanthrene	mg/kg	< 0.5	0.5	Pass	
3.3'-Dichlorobenzidine	mg/kg	< 0.5	0.5	Pass	



Test	Units	Result 1	Acceptance Limits	Pass Limits	Qualifying Code
4-Aminobiphenyl	mg/kg	< 0.5	0.5	Pass	
4-Bromophenyl phenyl ether	mg/kg	< 0.5	0.5	Pass	
4-Chloro-3-methylphenol	mg/kg	<1	1	Pass	
4-Chlorophenyl phenyl ether	mg/kg	< 0.5	0.5	Pass	
4-Nitrophenol	mg/kg	< 5	5	Pass	
4.4'-DDD	mg/kg	< 0.01	0.01	Pass	
4.4'-DDE	mg/kg	< 0.01	0.01	Pass	
4.4'-DDT	mg/kg	< 0.01	0.01	Pass	
7.12-Dimethylbenz(a)anthracene	mg/kg	< 0.5	0.5	Pass	
a-HCH	mg/kg	< 0.01	0.01	Pass	
Acetophenone	mg/kg	< 0.5	0.5	Pass	
Aldrin	mg/kg	< 0.01	0.01	Pass	
Aniline	mg/kg	< 0.5	0.5	Pass	
b-HCH	mg/kg	< 0.01	0.01	Pass	
Benzyl chloride	mg/kg	< 0.5	0.5	Pass	
Bis(2-chloroethoxy)methane	mg/kg	< 0.5	0.5	Pass	
Bis(2-chloroisopropyl)ether	mg/kg	< 0.5	0.5	Pass	
Bis(2-ethylhexyl)phthalate	mg/kg	< 0.5	0.5	Pass	
Butyl benzyl phthalate	mg/kg	< 0.5	0.5	Pass	
d-HCH	mg/kg	< 0.01	0.01	Pass	
Di-n-butyl phthalate	mg/kg	< 0.5	0.5	Pass	
Di-n-octyl phthalate	mg/kg	< 0.5	0.5	Pass	
Dibenz(a.j)acridine	mg/kg	< 0.5	0.5	Pass	
Dibenzofuran	mg/kg	< 0.5	0.5	Pass	
Dieldrin	mg/kg	< 0.01	0.01	Pass	
Diethyl phthalate	mg/kg	< 0.5	0.5	Pass	
Dimethyl phthalate	mg/kg	< 0.5	0.5	Pass	
Dimethylaminoazobenzene	mg/kg	< 0.5	0.5	Pass	
Diphenylamine	mg/kg	< 0.5	0.5	Pass	
Endosulfan I	mg/kg	< 0.01	0.01	Pass	
Endosulfan II	mg/kg	< 0.01	0.01	Pass	
Endosulfan sulphate	mg/kg	< 0.01	0.01	Pass	
Endrin Endrin	mg/kg	< 0.01	0.01	Pass	
Endrin aldehyde		0.01	0.01	Pass	
Endrin ketone	mg/kg mg/kg	< 0.01	0.01	Pass	
			0.01	Pass	
g-HCH (Lindane) Heptachlor	mg/kg	< 0.01 < 0.01	0.01	Pass	
Heptachlor epoxide	mg/kg	< 0.01	0.01	Pass	
Hexachlorobenzene	mg/kg	< 0.01	0.01	Pass	
	mg/kg				
Hexachlorobutadiene Hexachlorocyclopentadiene	mg/kg	< 0.5	0.5	Pass	
Hexachloroethane	mg/kg	< 0.5	0.5 0.5	Pass	
	mg/kg	< 0.5		Pass	
Methoxychlor N Nitropodibutylamina	mg/kg	< 0.01	0.01	Pass	
N-Nitrosodibutylamine	mg/kg	< 0.5	0.5	Pass	
N-Nitrosodipropylamine	mg/kg	< 0.5	0.5	Pass	
N-Nitrosopiperidine	mg/kg	< 0.5	0.5	Pass	
Nitrobenzene	mg/kg	< 0.5	0.5	Pass	
Pentachlorobenzene Dentachloropitrahenzene	mg/kg	< 0.5	0.5	Pass	
Pentachloronitrobenzene Dentachloronitrobenzene	mg/kg	< 0.5	0.5	Pass	
Pentachlorophenol	mg/kg	<1	1	Pass	
Phenol	mg/kg	< 0.5	0.5	Pass	
Pronamide	mg/kg	< 0.5	0.5	Pass	
Trifluralin LCS - % Recovery	mg/kg	< 0.5	0.5	Pass	



Test	Units	Result 1	Acceptance Limits	Pass Limits	Qualifying Code
Polycyclic Aromatic Hydrocarbons (NZ MfE)					
Acenaphthene	%	102	70-130	Pass	
Acenaphthylene	%	101	70-130	Pass	
Anthracene	%	89	70-130	Pass	
Benz(a)anthracene	%	96	70-130	Pass	
Benzo(a)pyrene	%	113	70-130	Pass	
Benzo(b&j)fluoranthene	%	80	70-130	Pass	
Benzo(g.h.i)perylene	%	71	70-130	Pass	
Benzo(k)fluoranthene	%	104	70-130	Pass	
Chrysene	%	100	70-130	Pass	
Dibenz(a.h)anthracene	%	75	70-130	Pass	
Fluoranthene	%	86	70-130	Pass	
Fluorene	%	91	70-130	Pass	
Indeno(1.2.3-cd)pyrene	%	77	70-130	Pass	
Naphthalene	%	106	70-130	Pass	
Phenanthrene	%	86	70-130	Pass	
Pyrene	%	89	70-130	Pass	
LCS - % Recovery	70	1 00	70 100	1 455	
Semivolatile Organics		Т			
1-Chloronaphthalene	%	128	70-130	Pass	
1-Naphthylamine	%	107	70-130	Pass	
1.2-Dichlorobenzene	%	78	70-130	Pass	
		1			
1.2.3-Trichlorobenzene	%	77	70-130	Pass	
1.2.3.4-Tetrachlorobenzene	%	82	70-130	Pass	
1.2.3.5-Tetrachlorobenzene	%	80	70-130	Pass	
1.3-Dichlorobenzene	%	86	70-130	Pass	
1.3.5-Trichlorobenzene	%	71	70-130	Pass	
1.4-Dichlorobenzene	%	72	70-130	Pass	
2-Chloronaphthalene	%	125	70-130	Pass	
2-Chlorophenol	%	64	25-130	Pass	
2-Methyl-4.6-dinitrophenol	%	48	25-130	Pass	
2-Methylnaphthalene	%	72	70-130	Pass	
2-Methylphenol (o-Cresol)	%	59	25-130	Pass	
2-Naphthylamine	%	78	70-130	Pass	
2-Nitrophenol	%	69	25-130	Pass	
2.4-Dichlorophenol	%	51	25-130	Pass	
2.4-Dimethylphenol	%	56	25-130	Pass	
2.4-Dinitrophenol	%	73	25-130	Pass	
2.4.5-Trichlorophenol	%	66	25-130	Pass	
2.4.6-Trichlorophenol	%	96	25-130	Pass	
2.6-Dichlorophenol	%	64	25-130	Pass	
3&4-Methylphenol (m&p-Cresol)	%	56	25-130	Pass	
3-Methylcholanthrene	%	78	70-130	Pass	
4-Aminobiphenyl	%	79	70-130	Pass	
4-Chloro-3-methylphenol	%	57	25-130	Pass	
4-Nitrophenol	%	113	25-130	Pass	
4.4'-DDD	%	101	70-130	Pass	
4.4'-DDE	%	85	70-130	Pass	
4.4'-DDT	%	124	70-130	Pass	
7.12-Dimethylbenz(a)anthracene	%	81	70-130	Pass	
a-HCH	%	83	70-130	Pass	
Acetophenone	%	81	70-130	Pass	
Aldrin	%	80	70-130	Pass	
b-HCH	%	96	70-130	Pass	



Test			Units	Result 1		Acceptance Limits	Pass Limits	Qualifying Code
Benzyl chloride			%	77		70-130	Pass	
Bis(2-chloroethoxy)methane			%	70		70-130	Pass	
Bis(2-chloroisopropyl)ether			%	81		70-130	Pass	
Bis(2-ethylhexyl)phthalate			%	125		70-130	Pass	
Butyl benzyl phthalate			%	119		70-130	Pass	
d-HCH			%	79		70-130	Pass	
Di-n-butyl phthalate			%	73		70-130	Pass	
Di-n-octyl phthalate			%	86		70-130	Pass	
Dieldrin			%	84		70-130	Pass	
Diethyl phthalate			%	70		70-130	Pass	
Dimethyl phthalate			%	74		70-130	Pass	
Dimethylaminoazobenzene			%	78		70-130	Pass	
Diphenylamine			%	73		70-130	Pass	
Endosulfan I			%	95		70-130	Pass	
Endosulfan II			%	105		70-130	Pass	
Endosulfan sulphate			%	92		70-130	Pass	
Endrin			%	81		70-130	Pass	
Endrin aldehyde			%	92		70-130	Pass	
Endrin ketone			%	97		70-130	Pass	
g-HCH (Lindane)			%	115		70-130	Pass	
Heptachlor			%	82		70-130	Pass	
Heptachlor epoxide			%	75		70-130	Pass	
Hexachlorobenzene			%	82		70-130	Pass	
Hexachlorobutadiene			%	72		70-130	Pass	
Methoxychlor			%	104		70-130	Pass	
N-Nitrosopiperidine			%	71		70-130	Pass	
Nitrobenzene			%	100		70-130	Pass	
Pentachlorobenzene			%	70		70-130	Pass	
Pentachloronitrobenzene			%	82		70-130	Pass	
Pentachlorophenol			%	100		25-130	Pass	
Phenol			%	58		25-130	Pass	
Trifluralin			%	80		70-130	Pass	
Test	Lab Sample ID	QA Source	Units	Result 1		Acceptance Limits	Pass Limits	Qualifying Code
Spike - % Recovery				1	1			
Polycyclic Aromatic Hydrocarbons	i ' ' ' ' ' ' ' ' ' ' ' ' ' ' ' ' ' ' '			Result 1				
Acenaphthylene	K22-No0023937	NCP	%	72		70-130	Pass	
Benzo(b&j)fluoranthene	K22-No0023937	NCP	%	70		70-130	Pass	
Benzo(g.h.i)perylene	K22-No0019326	NCP	%	74		70-130	Pass	
Chrysene	K22-No0019326	NCP	%	74		70-130	Pass	
Indeno(1.2.3-cd)pyrene	K22-No0019326	NCP	%	73		70-130	Pass	
Spike - % Recovery				l	T T			
Semivolatile Organics				Result 1			_	
a-HCH	K22-No0001195	NCP	%	73		70-130	Pass	
Aldrin	K22-No0001195	NCP	%	76		70-130	Pass	
Endosulfan II	K22-No0023937	NCP	%	70		70-130	Pass	
Endrin	K22-No0007117	NCP	%	75		70-130	Pass	
Heptachlor	K22-No0007117	NCP	%	77		70-130	Pass	
Heptachlor epoxide	K22-No0007117	NCP	%	87		70-130	Pass	
Spike - % Recovery	(1.1.							
Polycyclic Aromatic Hydrocarbons	, ,	0-		Result 1			_	
Acenaphthene	K22-No0015976	CP	%	78		70-130	Pass	
Anthracene	K22-No0015976	CP	%	70		70-130	Pass	
Benz(a)anthracene	K22-No0015976	CP	%	71		70-130	Pass	
Benzo(a)pyrene	K22-No0015976	CP	%	81		70-130	Pass	



Test	Lab Sample ID	QA Source	Units	Result 1			Acceptance Limits	Pass Limits	Qualifying Code
Benzo(k)fluoranthene	K22-No0015976	CP	%	71			70-130	Pass	
Dibenz(a.h)anthracene	K22-No0015976	CP	%	73			70-130	Pass	
Fluoranthene	K22-No0015976	CP	%	70			70-130	Pass	
Fluorene	K22-No0015976	CP	%	70			70-130	Pass	
Naphthalene	K22-No0015976	CP	%	76			70-130	Pass	
Phenanthrene	K22-No0015976	CP	%	70			70-130	Pass	
Pyrene	K22-No0015976	CP	%	73			70-130	Pass	
Spike - % Recovery									
Semivolatile Organics				Result 1					
1-Chloronaphthalene	K22-No0015976	СР	%	125			70-130	Pass	
2.3.4.6-Tetrachlorophenol	K22-No0015976	СР	%	87			70-130	Pass	
4.4'-DDD	K22-No0015976	СР	%	80			70-130	Pass	
4.4'-DDE	K22-No0015976	СР	%	75			70-130	Pass	
4.4'-DDT	K22-No0015976	CP	%	123			70-130	Pass	
b-HCH	K22-No0015976	CP	%	71			70-130	Pass	
Benzyl chloride	K22-No0015976	CP	%	85			70-130	Pass	
Bis(2-chloroisopropyl)ether	K22-No0015976	CP	%	104			70-130	Pass	
Bis(2-ethylhexyl)phthalate	K22-No0015976	CP	%	91			70-130	Pass	
Butyl benzyl phthalate	K22-No0015976	CP	%	83			70-130	Pass	
d-HCH	K22-No0015976	CP	%	70			70-130	Pass	
Di-n-butyl phthalate	K22-No0015976	CP	%	74			70-130	Pass	
Di-n-octyl phthalate	K22-No0015976	CP	%	92			70-130	Pass	
Dieldrin	K22-No0015976	CP	%	71			70-130	Pass	
		CP		71			70-130		
Diethyl phthalate	K22-No0015976		%					Pass	
Dimethylaminoazobenzene	K22-No0015976	CP	%	70			70-130	Pass	
Endosulfan I	K22-No0015976	CP	%	76			70-130	Pass	
Endosulfan sulphate	K22-No0015976	CP	%	78			70-130	Pass	
Endrin aldehyde	K22-No0015976	CP	%	85			70-130	Pass	
Endrin ketone	K22-No0015976	CP	%	78			70-130	Pass	
g-HCH (Lindane)	K22-No0015976	CP	%	107			70-130	Pass	
Hexachlorobenzene	K22-No0015976	CP	%	123			70-130	Pass	
Methoxychlor	K22-No0015976	CP	%	85			70-130	Pass	
Pentachloronitrobenzene	K22-No0015976	CP	%	77			70-130	Pass	
Pentachlorophenol	K22-No0015976	CP	%	106			70-130	Pass	
Trifluralin	K22-No0015976	CP	%	71			70-130	Pass	
Spike - % Recovery				I	I		ı		
Semivolatile Organics				Result 1					
1-Naphthylamine	K22-No0005712	NCP	%	71			70-130	Pass	
1.2.3.4-Tetrachlorobenzene	K22-No0005712	NCP	%	72			70-130	Pass	
1.3-Dichlorobenzene	K22-No0005712	NCP	%	71			70-130	Pass	
2.4.5-Trichlorophenol	K22-No0005712	NCP	%	91			70-130	Pass	
3-Methylcholanthrene	K22-No0005712	NCP	%	89	1		70-130	Pass	
4-Aminobiphenyl	K22-No0005712	NCP	%	83			70-130	Pass	
Dibenz(a.j)acridine	K22-No0001205	NCP	%	2.4			70-130	Fail	
Dimethyl phthalate	K22-No0005712	NCP	%	73			70-130	Pass	
Test	Lab Sample ID	QA Source	Units	Result 1			Acceptance Limits	Pass Limits	Qualifying Code
Duplicate									
Polycyclic Aromatic Hydrocarbon	ns (NZ MfE)			Result 1	Result 2	RPD			
Acenaphthene	K22-No0015975	СР	mg/kg	< 0.03	< 0.03	<1	30%	Pass	
Acenaphthylene	K22-No0015975	СР	mg/kg	< 0.03	< 0.03	<1	30%	Pass	
Anthracene	K22-No0015975	СР	mg/kg	< 0.03	< 0.03	<1	30%	Pass	
Benz(a)anthracene	K22-No0015975	СР	mg/kg	0.07	0.04	40	30%	Fail	Q15
Benzo(a)pyrene	K22-No0015975	CP	mg/kg	0.06	0.04	26	30%	Pass	
Benzo(b&j)fluoranthene	K22-No0015975	CP	mg/kg	0.04	< 0.03	21	30%	Pass	



Test	Lab Sample ID	QA Source	Units	Result 1			Acceptance Limits	Pass Limits	Qualifying Code
Duplicate									0000
Polycyclic Aromatic Hydrocarbon	ns (NZ MfE)			Result 1	Result 2	RPD			
Benzo(g.h.i)perylene	K22-No0015975	СР	mg/kg	< 0.03	0.04	11	30%	Pass	
Benzo(k)fluoranthene	K22-No0015975	СР	mg/kg	< 0.03	< 0.03	<1	30%	Pass	
Chrysene	K22-No0015975	СР	mg/kg	0.09	0.07	25	30%	Pass	
Dibenz(a.h)anthracene	K22-No0015975	СР	mg/kg	< 0.03	< 0.03	<1	30%	Pass	
Fluoranthene	K22-No0015975	СР	mg/kg	0.11	0.07	39	30%	Fail	Q15
Fluorene	K22-No0015975	CP	mg/kg	< 0.03	< 0.03	<1	30%	Pass	
Indeno(1.2.3-cd)pyrene	K22-No0015975	СР	mg/kg	< 0.03	< 0.03	<1	30%	Pass	
Naphthalene	K22-No0015975	СР	mg/kg	< 0.1	< 0.1	<1	30%	Pass	
Phenanthrene	K22-No0015975	СР	mg/kg	0.07	< 0.03	110	30%	Fail	Q15
Pyrene	K22-No0015975	СР	mg/kg	0.13	0.07	59	30%	Fail	Q15
Duplicate									
Semivolatile Organics				Result 1	Result 2	RPD			
1-Chloronaphthalene	K22-No0015975	СР	mg/kg	< 0.5	< 0.5	<1	30%	Pass	
1-Naphthylamine	K22-No0015975	СР	mg/kg	< 0.5	< 0.5	<1	30%	Pass	
1.2-Dichlorobenzene	K22-No0015975	CP	mg/kg	< 0.5	< 0.5	<1	30%	Pass	
1.2.3-Trichlorobenzene	K22-No0015975	CP	mg/kg	< 0.5	< 0.5	<1	30%	Pass	
1.2.3.4-Tetrachlorobenzene	K22-No0015975	CP	mg/kg	< 0.5	< 0.5	<1	30%	Pass	
1.2.3.5-Tetrachlorobenzene	K22-No0015975	CP	mg/kg	< 0.5	< 0.5	<1	30%	Pass	
1.2.4-Trichlorobenzene	K22-No0015975	CP	mg/kg	< 0.5	< 0.5	<1	30%	Pass	
1.2.4.5-Tetrachlorobenzene	K22-No0015975	CP	mg/kg	< 0.5	< 0.5	<1	30%	Pass	
1.3-Dichlorobenzene	K22-No0015975	CP	mg/kg	< 0.5	< 0.5	<1	30%	Pass	
1.3.5-Trichlorobenzene	K22-No0015975	CP	mg/kg	< 0.5	< 0.5	<1	30%	Pass	
1.4-Dichlorobenzene	K22-No0015975	CP	mg/kg	< 0.5	< 0.5	<1	30%	Pass	
2-Chloronaphthalene	K22-No0015975	CP	mg/kg	< 0.5	< 0.5	<1	30%	Pass	
2-Chlorophenol	K22-No0015975	CP	mg/kg	< 0.5	< 0.5	<1	30%	Pass	
2-Methyl-4.6-dinitrophenol	K22-No0015975	CP	mg/kg	< 5	< 5	<1	30%	Pass	
2-Methylnaphthalene	K22-No0015975	CP	mg/kg	< 0.5	< 0.5	<1	30%	Pass	
2-Methylphenol (o-Cresol)	K22-No0015975	CP	mg/kg	< 0.2	< 0.2	<1	30%	Pass	
2-Naphthylamine	K22-No0015975	CP	mg/kg	< 0.5	< 0.5	<1	30%	Pass	
2-Nitroaniline	K22-No0015975	CP	mg/kg	< 0.5	< 0.5	<1	30%	Pass	
2-Nitrophenol	K22-No0015975	CP	mg/kg	< 1	< 1	<1	30%	Pass	
2-Picoline	K22-No0015975	CP	mg/kg	< 0.5	< 0.5	<1	30%	Pass	
2.3.4.6-Tetrachlorophenol	K22-No0015975	CP	mg/kg	< 5	< 5	<1	30%	Pass	
2.4-Dichlorophenol	K22-No0015975	CP	mg/kg	< 0.5	< 0.5	<1	30%	Pass	
2.4-Dimethylphenol	K22-No0015975	CP	mg/kg	< 0.5	< 0.5	<1	30%	Pass	
2.4-Dinitrophenol	K22-No0015975	CP	mg/kg	< 5	< 5	<1	30%	Pass	
2.4-Dinitrotoluene	K22-No0015975	CP	mg/kg	< 0.5	< 0.5	<1	30%	Pass	
2.4.5-Trichlorophenol	K22-No0015975	CP	mg/kg	< 1	< 1	<1	30%	Pass	
2.4.6-Trichlorophenol	K22-No0015975	CP	mg/kg	< 1	< 1	<1	30%	Pass	
2.6-Dichlorophenol	K22-No0015975	CP	mg/kg	< 0.5	< 0.5	<1	30%	Pass	
2.6-Dinitrotoluene	K22-No0015975	CP	mg/kg	< 0.5	< 0.5	<1	30%	Pass	
3&4-Methylphenol (m&p-Cresol)	K22-No0015975	CP	mg/kg	< 0.4	< 0.4	<1	30%	Pass	
3-Methylcholanthrene	K22-No0015975	CP	mg/kg	< 0.5	< 0.5	<1	30%	Pass	
3.3'-Dichlorobenzidine	K22-No0015975	CP	mg/kg	< 0.5	< 0.5	<1	30%	Pass	
4-Aminobiphenyl	K22-No0015975	CP	mg/kg	< 0.5	< 0.5	<1	30%	Pass	
4-Bromophenyl phenyl ether	K22-No0015975	CP	mg/kg	< 0.5	< 0.5	<1	30%	Pass	
4-Chloro-3-methylphenol	K22-No0015975	CP	mg/kg	< 1	< 1	<1	30%	Pass	
4-Chlorophenyl phenyl ether	K22-No0015975	CP	mg/kg	< 0.5	< 0.5	<1	30%	Pass	
4-Nitrophenol	K22-No0015975	CP	mg/kg	< 5	< 5	<1	30%	Pass	
4.4'-DDD	K22-No0015975	CP	mg/kg	< 0.01	< 0.01	<1	30%	Pass	
4.4'-DDE	K22-No0015975		mg/kg	< 0.01	< 0.01	<1	30%	Pass	
4.4'-DDT	K22-No0015975	CP	mg/kg	< 0.01	< 0.01	<1	30%	Pass	



Duplicate									
Semivolatile Organics				Result 1	Result 2	RPD			
7.12-Dimethylbenz(a)anthracene	K22-No0015975	CP	mg/kg	< 0.5	< 0.5	<1	30%	Pass	
a-HCH	K22-No0015975	CP	mg/kg	< 0.01	< 0.01	<1	30%	Pass	
Acetophenone	K22-No0015975	CP	mg/kg	< 0.5	< 0.5	<1	30%	Pass	
Aldrin	K22-No0015975	CP	mg/kg	< 0.01	< 0.01	<1	30%	Pass	
Aniline	K22-No0015975	CP	mg/kg	< 0.5	< 0.5	<1	30%	Pass	
b-HCH	K22-No0015975	CP	mg/kg	< 0.01	< 0.01	<1	30%	Pass	
Benzyl chloride	K22-No0015975	CP	mg/kg	< 0.5	< 0.5	<1	30%	Pass	
Bis(2-chloroethoxy)methane	K22-No0015975	CP	mg/kg	< 0.5	< 0.5	<1	30%	Pass	
Bis(2-chloroisopropyl)ether	K22-No0015975	CP	mg/kg	< 0.5	< 0.5	<1	30%	Pass	
Bis(2-ethylhexyl)phthalate	K22-No0015975	CP	mg/kg	< 0.5	< 0.5	<1	30%	Pass	
Butyl benzyl phthalate	K22-No0015975	CP	mg/kg	< 0.5	< 0.5	<1	30%	Pass	
d-HCH	K22-No0015975	CP	mg/kg	< 0.01	< 0.01	<1	30%	Pass	
Di-n-butyl phthalate	K22-No0015975	CP	mg/kg	< 0.5	< 0.5	<1	30%	Pass	
, ,		CP							
Di-n-octyl phthalate	K22-No0015975	CP CP	mg/kg	< 0.5	< 0.5	<1	30%	Pass	
Dibenz(a.j)acridine	K22-No0015975	CP CP	mg/kg	< 0.5	< 0.5	<1	30%	Pass	
Dibenzofuran Dioldrin	K22-No0015975	CP CP	mg/kg	< 0.5	< 0.5	<1	30%	Pass Fail	
Dieldrin Diethyl abthalata	K22-No0015975		mg/kg	< 0.01	0.01	58	30%		
Diethyl phthalate	K22-No0015975	CP	mg/kg	< 0.5	< 0.5	<1	30%	Pass	
Dimethyl phthalate	K22-No0015975	CP	mg/kg	< 0.5	< 0.5	<1	30%	Pass	
Dimethylaminoazobenzene	K22-No0015975	CP	mg/kg	< 0.5	< 0.5	<1	30%	Pass	
Diphenylamine	K22-No0015975	CP	mg/kg	< 0.5	< 0.5	<1	30%	Pass	
Endosulfan I	K22-No0015975	CP	mg/kg	< 0.01	< 0.01	<1	30%	Pass	
Endosulfan II	K22-No0015975	CP	mg/kg	< 0.01	< 0.01	<1	30%	Pass	
Endosulfan sulphate	K22-No0015975	CP	mg/kg	< 0.01	< 0.01	<1	30%	Pass	
Endrin	K22-No0015975	CP	mg/kg	< 0.01	< 0.01	<1	30%	Pass	
Endrin aldehyde	K22-No0015975	CP	mg/kg	< 0.01	< 0.01	<1	30%	Pass	
Endrin ketone	K22-No0015975	CP	mg/kg	< 0.01	< 0.01	<1	30%	Pass	
g-HCH (Lindane)	K22-No0015975	CP	mg/kg	< 0.01	< 0.01	<1	30%	Pass	
Heptachlor	K22-No0015975	CP	mg/kg	< 0.01	< 0.01	<1	30%	Pass	
Heptachlor epoxide	K22-No0015975	CP	mg/kg	< 0.01	< 0.01	<1	30%	Pass	
Hexachlorobenzene	K22-No0015975	CP	mg/kg	< 0.01	< 0.01	<1	30%	Pass	
Hexachlorobutadiene	K22-No0015975	CP	mg/kg	< 0.5	< 0.5	<1	30%	Pass	
Hexachlorocyclopentadiene	K22-No0015975	CP	mg/kg	< 0.5	< 0.5	<1	30%	Pass	
Hexachloroethane	K22-No0015975	CP	mg/kg	< 0.5	< 0.5	<1	30%	Pass	
Methoxychlor	K22-No0015975	CP	mg/kg	< 0.01	< 0.01	<1	30%	Pass	
N-Nitrosodibutylamine	K22-No0015975	CP	mg/kg	< 0.5	< 0.5	<1	30%	Pass	
N-Nitrosodipropylamine	K22-No0015975	CP	mg/kg	< 0.5	< 0.5	<1	30%	Pass	
N-Nitrosopiperidine	K22-No0015975	CP	mg/kg	< 0.5	< 0.5	<1	30%	Pass	
Nitrobenzene	K22-No0015975	CP	mg/kg	< 0.5	< 0.5	<1	30%	Pass	
Pentachlorobenzene	K22-No0015975	CP	mg/kg	< 0.5	< 0.5	<1	30%	Pass	
Pentachloronitrobenzene	K22-No0015975	CP	mg/kg	< 0.5	< 0.5	<1	30%	Pass	
Pentachlorophenol	K22-No0015975	CP	mg/kg	< 1	< 1	<1	30%	Pass	
Phenol	K22-No0015975	CP	mg/kg	< 0.5	< 0.5	<1	30%	Pass	
Pronamide	K22-No0015975	CP	mg/kg	< 0.5	< 0.5	<1	30%	Pass	
Trifluralin	K22-No0015975	CP	mg/kg	< 0.5	< 0.5	<1	30%	Pass	
Duplicate									
				Result 1	Result 2	RPD			
% Moisture	K22-No0015975	CP	%	30	30	<1	30%	Pass	
Duplicate									
Duplicate				Result 1	Result 2	RPD			



Duplicate									
Polycyclic Aromatic Hydrocarbons	: (NZ MfF)			Result 1	Result 2	RPD			
	K22-No0015988	CP	mg/kg	< 0.03	< 0.03	<1	30%	Pass	
Acenaphthene Acenaphthylene	K22-N00015988	CP CP		< 0.03	< 0.03	<1	30%	Pass	
· •		CP CP	mg/kg						
Anthracene	K22-No0015988	CP CP	mg/kg	< 0.03	< 0.03	<1	30%	Pass	
Benza(a)anthracene	K22-No0015988	CP CP	mg/kg	< 0.03	< 0.03	<1	30%	Pass	
Benzo(a)pyrene	K22-No0015988		mg/kg	< 0.03	< 0.03	<1	30%	Pass	
Benzo(b&j)fluoranthene	K22-No0015988	CP	mg/kg	< 0.03	< 0.03	<1	30%	Pass	
Benzo(g.h.i)perylene	K22-No0015988	CP	mg/kg	< 0.03	< 0.03	<1	30%	Pass	
Benzo(k)fluoranthene	K22-No0015988	CP	mg/kg	< 0.03	< 0.03	<1	30%	Pass	
Chrysene	K22-No0015988	CP	mg/kg	< 0.03	< 0.03	<1	30%	Pass	
Dibenz(a.h)anthracene	K22-No0015988	CP	mg/kg	< 0.03	< 0.03	<1	30%	Pass	
Fluoranthene	K22-No0015988	CP	mg/kg	< 0.03	< 0.03	<1	30%	Pass	
Fluorene	K22-No0015988	CP	mg/kg	< 0.03	< 0.03	<1	30%	Pass	
Indeno(1.2.3-cd)pyrene	K22-No0015988	CP	mg/kg	< 0.03	< 0.03	<1	30%	Pass	
Naphthalene	K22-No0015988	CP	mg/kg	< 0.1	< 0.1	<1	30%	Pass	
Phenanthrene	K22-No0015988	CP	mg/kg	< 0.03	< 0.03	<1	30%	Pass	
Pyrene	K22-No0015988	CP	mg/kg	< 0.03	< 0.03	<1	30%	Pass	
Duplicate				I					
Semivolatile Organics			1	Result 1	Result 2	RPD			
1-Chloronaphthalene	K22-No0015988	CP	mg/kg	< 0.5	< 0.5	<1	30%	Pass	
1-Naphthylamine	K22-No0015988	CP	mg/kg	< 0.5	< 0.5	<1	30%	Pass	
1.2-Dichlorobenzene	K22-No0015988	CP	mg/kg	< 0.5	< 0.5	<1	30%	Pass	
1.2.3-Trichlorobenzene	K22-No0015988	CP	mg/kg	< 0.5	< 0.5	<1	30%	Pass	
1.2.3.4-Tetrachlorobenzene	K22-No0015988	CP	mg/kg	< 0.5	< 0.5	<1	30%	Pass	
1.2.3.5-Tetrachlorobenzene	K22-No0015988	CP	mg/kg	< 0.5	< 0.5	<1	30%	Pass	
1.2.4-Trichlorobenzene	K22-No0015988	CP	mg/kg	< 0.5	< 0.5	<1	30%	Pass	
1.2.4.5-Tetrachlorobenzene	K22-No0015988	CP	mg/kg	< 0.5	< 0.5	<1	30%	Pass	
1.3-Dichlorobenzene	K22-No0015988	CP	mg/kg	< 0.5	< 0.5	<1	30%	Pass	
1.3.5-Trichlorobenzene	K22-No0015988	CP	mg/kg	< 0.5	< 0.5	<1	30%	Pass	
1.4-Dichlorobenzene	K22-No0015988	CP	mg/kg	< 0.5	< 0.5	<1	30%	Pass	
2-Chloronaphthalene	K22-No0015988	СР	mg/kg	< 0.5	< 0.5	<1	30%	Pass	
2-Chlorophenol	K22-No0015988	СР	mg/kg	< 0.5	< 0.5	<1	30%	Pass	
2-Methyl-4.6-dinitrophenol	K22-No0015988	СР	mg/kg	< 5	< 5	<1	30%	Pass	
2-Methylnaphthalene	K22-No0015988	CP	mg/kg	< 0.5	< 0.5	<1	30%	Pass	
2-Methylphenol (o-Cresol)	K22-No0015988	СР	mg/kg	< 0.2	< 0.2	<1	30%	Pass	
2-Naphthylamine	K22-No0015988	СР	mg/kg	< 0.5	< 0.5	<1	30%	Pass	
2-Nitroaniline	K22-No0015988	CP	mg/kg	< 0.5	< 0.5	<1	30%	Pass	
2-Nitrophenol	K22-No0015988	СР	mg/kg	< 1	< 1	<1	30%	Pass	
2-Picoline	K22-No0015988	CP	mg/kg	< 0.5	< 0.5	<1	30%	Pass	
2.3.4.6-Tetrachlorophenol	K22-No0015988	CP	mg/kg	< 5	< 5	<1	30%	Pass	
2.4-Dichlorophenol	K22-No0015988	CP	mg/kg	< 0.5	< 0.5	<1	30%	Pass	
2.4-Dimethylphenol	K22-No0015988	CP	mg/kg	< 0.5	< 0.5	<1	30%	Pass	
2.4-Dinitrophenol	K22-No0015988	CP	mg/kg	< 5	< 5	<1	30%	Pass	
2.4-Dinitrotoluene	K22-No0015988	CP	mg/kg	< 0.5	< 0.5	<1	30%	Pass	
2.4.5-Trichlorophenol	K22-No0015988	CP	mg/kg	< 1	< 1	<1	30%	Pass	
2.4.6-Trichlorophenol	K22-N00015988	CP	mg/kg	< 1	< 1	<1	30%	Pass	
2.4.6-Trichlorophenol	K22-N00015988	CP CP		1			30%	Pass	
2.6-Dichiorophenoi	K22-N00015988	CP CP	mg/kg	< 0.5	< 0.5 < 0.5	<1	30%	Pass	
			mg/kg	< 0.5	1	<1			
3&4-Methylphenol (m&p-Cresol)	K22-No0015988	CP	mg/kg	< 0.4	< 0.4	<1	30%	Pass	
3-Methylcholanthrene	K22-No0015988	CP	mg/kg	< 0.5	< 0.5	<1	30%	Pass	
3.3'-Dichlorobenzidine	K22-No0015988	CP	mg/kg	< 0.5	< 0.5	<1	30%	Pass	
4-Aminobiphenyl	K22-No0015988	CP	mg/kg	< 0.5	< 0.5	<1	30%	Pass	
4-Bromophenyl phenyl ether	K22-No0015988	CP	mg/kg	< 0.5	< 0.5	<1	30%	Pass	
4-Chloro-3-methylphenol	K22-No0015988	CP	mg/kg	< 1	< 1	<1	30%	Pass	
4-Chlorophenyl phenyl ether	K22-No0015988	CP	mg/kg	< 0.5	< 0.5	<1	30%	Pass	



Duplicate									
Semivolatile Organics				Result 1	Result 2	RPD			
4-Nitrophenol	K22-No0015988	CP	mg/kg	< 5	< 5	<1	30%	Pass	
4.4'-DDD	K22-No0015988	CP	mg/kg	< 0.01	< 0.01	<1	30%	Pass	
4.4'-DDE	K22-No0015988	CP	mg/kg	< 0.01	< 0.01	<1	30%	Pass	
4.4'-DDT	K22-No0015988	CP	mg/kg	< 0.01	< 0.01	<1	30%	Pass	
7.12-Dimethylbenz(a)anthracene	K22-No0015988	CP	mg/kg	< 0.5	< 0.5	<1	30%	Pass	
a-HCH	K22-No0015988	CP	mg/kg	< 0.01	< 0.01	<1	30%	Pass	
Acetophenone	K22-No0015988	CP	mg/kg	< 0.5	< 0.5	<1	30%	Pass	
Aldrin	K22-No0015988	CP	mg/kg	< 0.01	< 0.01	<1	30%	Pass	
Aniline	K22-No0015988	CP	mg/kg	< 0.5	< 0.5	<1	30%	Pass	
b-HCH	K22-No0015988	CP	mg/kg	< 0.01	< 0.01	<1	30%	Pass	
Benzyl chloride	K22-No0015988	CP	mg/kg	< 0.5	< 0.5	<1	30%	Pass	
Bis(2-chloroethoxy)methane	K22-No0015988	CP	mg/kg	< 0.5	< 0.5	<1	30%	Pass	
Bis(2-chloroisopropyl)ether	K22-No0015988	CP	mg/kg	< 0.5	< 0.5	<1	30%	Pass	
Bis(2-ethylhexyl)phthalate	K22-No0015988	CP	mg/kg	< 0.5	< 0.5	<1	30%	Pass	
Butyl benzyl phthalate	K22-No0015988	CP CP	mg/kg	< 0.5	< 0.5	<1	30%	Pass	
d-HCH	K22-No0015988	CP CP	mg/kg	< 0.01	< 0.01	<1	30%	Pass	
Di-n-butyl phthalate	K22-No0015988	CP	mg/kg	< 0.5	< 0.5	<1	30%	Pass	
Di-n-octyl phthalate	K22-No0015988	CP	mg/kg	< 0.5	< 0.5	<1	30%	Pass	
Dibenz(a.j)acridine	K22-No0015988	CP		< 0.5	< 0.5	<1	30%	Pass	
· •/		CP CP	mg/kg			<1	 	Pass	
Dibenzofuran Dieldrin	K22-No0015988	CP CP	mg/kg	< 0.5	< 0.5	<1	30%	Pass	
	K22-No0015988		mg/kg	< 0.01	< 0.01		30%		
Diethyl phthalate	K22-No0015988	CP	mg/kg	< 0.5	< 0.5	<1	30%	Pass	
Dimethyl phthalate	K22-No0015988	CP	mg/kg	< 0.5	< 0.5	<1	30%	Pass	
Dimethylaminoazobenzene	K22-No0015988	CP	mg/kg	< 0.5	< 0.5	<1	30%	Pass	
Diphenylamine	K22-No0015988	CP	mg/kg	< 0.5	< 0.5	<1	30%	Pass	
Endosulfan I	K22-No0015988	CP	mg/kg	< 0.01	< 0.01	<1	30%	Pass	
Endosulfan II	K22-No0015988	CP	mg/kg	< 0.01	< 0.01	<1	30%	Pass	
Endosulfan sulphate	K22-No0015988	CP	mg/kg	< 0.01	< 0.01	<1	30%	Pass	
Endrin	K22-No0015988	CP	mg/kg	< 0.01	< 0.01	<1	30%	Pass	
Endrin aldehyde	K22-No0015988	CP	mg/kg	< 0.01	< 0.01	<1	30%	Pass	
Endrin ketone	K22-No0015988	CP	mg/kg	< 0.01	< 0.01	<1	30%	Pass	
g-HCH (Lindane)	K22-No0015988	CP	mg/kg	< 0.01	< 0.01	<1	30%	Pass	
Heptachlor	K22-No0015988	CP	mg/kg	< 0.01	< 0.01	<1	30%	Pass	
Heptachlor epoxide	K22-No0015988	CP	mg/kg	< 0.01	< 0.01	<1	30%	Pass	
Hexachlorobenzene	K22-No0015988	CP	mg/kg	< 0.01	< 0.01	<1	30%	Pass	
Hexachlorobutadiene	K22-No0015988	CP	mg/kg	< 0.5	< 0.5	<1	30%	Pass	
Hexachlorocyclopentadiene	K22-No0015988	CP	mg/kg	< 0.5	< 0.5	<1	30%	Pass	
Hexachloroethane	K22-No0015988	CP	mg/kg	< 0.5	< 0.5	<1	30%	Pass	
Methoxychlor	K22-No0015988	CP	mg/kg	< 0.01	< 0.01	<1	30%	Pass	
N-Nitrosodibutylamine	K22-No0015988	CP	mg/kg	< 0.5	< 0.5	<1	30%	Pass	
N-Nitrosodipropylamine	K22-No0015988	CP	mg/kg	< 0.5	< 0.5	<1	30%	Pass	
N-Nitrosopiperidine	K22-No0015988	CP	mg/kg	< 0.5	< 0.5	<1	30%	Pass	
Nitrobenzene	K22-No0015988	CP	mg/kg	< 0.5	< 0.5	<1	30%	Pass	
Pentachlorobenzene	K22-No0015988	CP	mg/kg	< 0.5	< 0.5	<1	30%	Pass	
Pentachloronitrobenzene	K22-No0015988	CP	mg/kg	< 0.5	< 0.5	<1	30%	Pass	
Pentachlorophenol	K22-No0015988	CP	mg/kg	< 1	< 1	<1	30%	Pass	
Phenol	K22-No0015988	CP	mg/kg	< 0.5	< 0.5	<1	30%	Pass	
Pronamide	K22-No0015988	CP	mg/kg	< 0.5	< 0.5	<1	30%	Pass	
Trifluralin	K22-No0015988	CP	mg/kg	< 0.5	< 0.5	<1	30%	Pass	



Comments

Sample Integrity

Custody Seals Intact (if used) N/A Attempt to Chill was evident Yes Sample correctly preserved Yes Appropriate sample containers have been used Yes Sample containers for volatile analysis received with minimal headspace Yes Samples received within HoldingTime Yes Some samples have been subcontracted No

Qualifier Codes/Comments

Code Description

Please note:- These two PAH isomers closely co-elute using the most contemporary analytical methods and both the reported concentration (and the TEQ) apply specifically to the total of the two co-eluting PAHs

N07

Q15 The RPD reported passes Eurofins Environment Testing's QC - Acceptance Criteria as defined in the Internal Quality Control Review and Glossary page of this report.

Authorised by:

Karishma Patel Analytical Services Manager Michael Ritchie Senior Analyst-Organic

Michael Ritchie

Head of Semi Volatiles (Key Technical Personnel)

Final Report - this report replaces any previously issued Report

- Indicates Not Requested
- * Indicates IANZ accreditation does not cover the performance of this service

Measurement uncertainty of test data is available on request or please click here.

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Appendix E

Leachate within the landfill

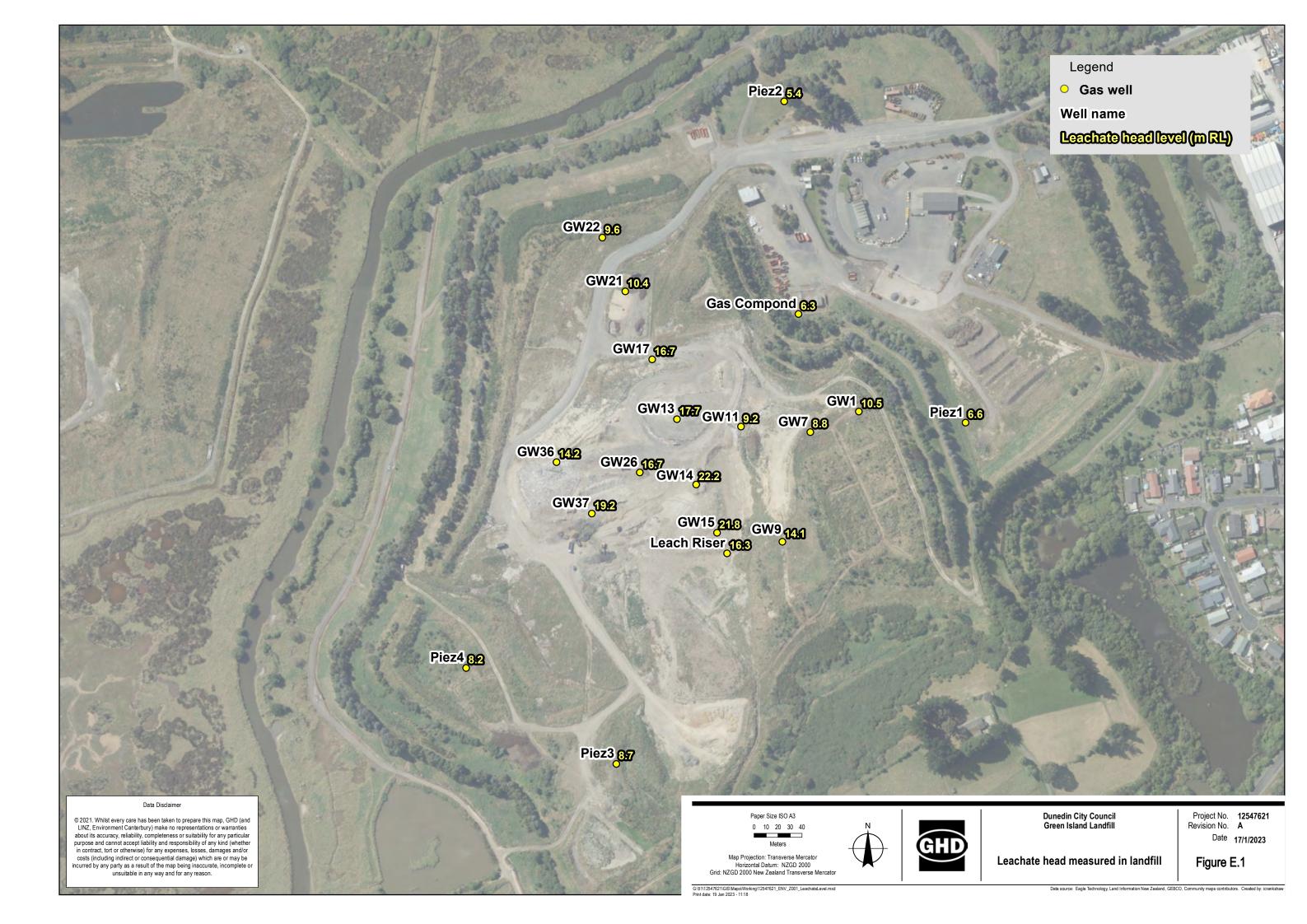
E-1 Leachate levels within landfill

Gas wells and groundwater monitoring wells within the landfill were dipped to measure the level of leachate within the landfill footprint (August 2022). These measurements are presented below in Table E.1. Figure E.1 shows the location of the wells. The leachate measurements were undertaken by a site contractor, GHD has not verified these measurements.

Table E.1 Leachate measurements

Gaswell and GW Well Dipping for Leachate & Base Levels - Aug 2022										
Areas and Wells	RL from Pix 4D	Dist from Ground	Top of Wellhead	Dip W/head to Leach	Dip W/head Bottom	Level of Leachate	Level of Base			
North / South Section										
GW22	17.4	0.900	18.3	8.750	10.320	9.6	8.0			
GW21	22.3	2.295	24.6	14.220	16.145	10.4	8.5			
GW17	23.2	2.050	25.3	8.570	18.825	16.7	6.4			
GW13	22.3	1.440	23.7	6.063	12.072	17.7	11.7			
GW14	24.0	1.580	25.6	3.415	13.365	22.2	12.2			
GW15	22.8	0.890	23.7	1.850	32.000	21.8	-8.3			
Leach Riser	23.0	2.280	25.3	8.970	9.925	16.3	15.4			
GW9	23.2	1.378	24.6	10.470	10.470	14.1	14.1			
West / East Section										
GW36	22.5	1.130	23.6	9.450	11.460	14.2	12.2			
GW37	20.7	1.310	22.0	2.835	5.923	19.2	16.1			
GW26	24.0	1.500	25.5	8.840	10.090	16.7	15.4			
GW11	23.2	0.845	24.0	14.873	26.000	9.2	-2.0			
GW7	22.5	1.310	23.8	15.000	20.295	8.8	3.5			
GW1	20.2	0.100	20.3	9.832	11.630	10.5	8.7			
Groundwater Wells										
Piez1	7.8	0.150	8.0	1.400	5.500	6.6	2.5			
Gas Compond	7.1	0.220	7.3	1.060	3.035	6.3	4.3			
Piez2	6.9	0.750	7.7	2.205	5.525	5.4	2.1			
Piez3	10.2	0.100	10.3	1.590	4.770	8.7	5.5			
Piez4	11.4	0.300	11.7	3.480	7.860	8.2	3.8			

^{*}GW15 was re-dipped by GHD, the well depth was measured at 26 m bgl



Appendix F

HELP modelling assessment

F-1 Introduction

This technical appendix outlines the water balance assessment of rainfall infiltration through the current and proposed cap at Green Island Landfill. The results of this HELP assessment has been incorporated into the Seep/W modelling (included as Appendix G), which has been prepared to inform an assessment of environmental effects (AEE) of the continued operation and subsequent closure of the landfill.

Rainfall infiltration through the landfill cap was assessed using the Hydrologic Evaluation of Landfill Performance (HELP) software (Berger and Schroeder, 2013). HELP 3.95D is a quasi-two-dimensional hydrologic model for conducting water balance analysis of landfills and cover systems. The model utilises weather, soil and landfill design data to account for the effects of surface storage, runoff, infiltration, evapotranspiration, soil moisture, lateral subsurface drainage, vertical drainage and leakage through soil and liners.

Four different areas of the landfill were modelled to assess leachate generation and leakage. The water balance estimated during the stages of landfill development were used to provide the whole of landfill water balance at different times through the proposed landfill operational lifecycle.

The results of the HELP modelling provide information to support an assessment of environmental effects for continued operation and subsequent closure of Green Island landfill.

F-1.1 Landfill profiles

Four different areas of the landfill, in different phases of landfill development, were considered in predicting the landfill water balance (Figure F.1)

- Area A Transfer station
 - 87,386 m² of final cap
- Area B Operational landfill capped area:
 - 50,379 m² of waste with final cap
- Area C Perimeter Bund capped area
 - 96,330 m² of waste overlain by a capping material
- Area D Operational landfill:
 - Existing conditions through to start of 2022
 - 97,380 m² of exposed open waste
 - Following installation of mid-landfill drainage layer
 - 97,380 m² of bare soil contoured towards drainage system
 - Future capped landfill
 - 97,380 m² of waste with final cap



Figure F.1 Landfill Areas A - D

The profile characteristics adopted are presented in Tables F.8 – F.11 (attached). Soil values of total pore volume, field capacity and wilting point were adopted from published USDA soil textures (Schroeder et al., 1994), however saturated hydraulic conductivity was modified for a number of the soil units. Justification for the adopted soil textures and saturated hydraulic conductivity is presented in Table F.1. Additional scenarios were also run for Area B and Area D where the hydraulic conductivity of the landfill cap (layer 3) was reduced to undertake sensitivity testing (Table F.2).

Table F.1 Justification for adopted soil texture and saturated hydraulic conductivity

Profile	Soil Unit	Adopted USDA Soil Texture ⁽¹⁾	USDA saturated hydraulic conductivity (cm/s)	Adopted saturated hydraulic conductivity (cm/s)	Justification
Area A Transfer Station - capped	Topsoil and cap (layers 1 & 2)	6 – Sandy Loam (slightly compacted)	1.9x10 ⁻⁴	5x10 ⁻⁴	Described as topsoil overlain by sandy silt or sandy gravel during GHD site investigation (2021).
Area B Operational landfill –	Topsoil and subsoil (layers 1 & 2)	14 – Silty Clay (slightly compacted)	2.5>	×10 ⁻⁵	No change to USDA value. Top soil and subsoil will be sourced from clean imported soils. Minor compaction and contouring assumed.
existing cap	Barrier Clay (layer 3)	28 – Silty Clay (moderately compacted)	1.2x10 ⁻⁶	1x10 ⁻⁶	On site borrow pit material largely described as sandy clayey silt. Laboratory testing results in the order of 1x10 ⁻⁹ m/s.
	Soil (layer 4)	28 – Silty Clay (moderately compacted)	1.2x10 ⁻⁶	1x10 ⁻⁴	Represents intermediate cover which will be present before placement of final cap.
Area C - Perimeter Bund	Cap (layer 1)	28 – Silty clay – moderately compacted	1.2x10 ⁻⁶	1x10 ⁻⁵	Adopted to represent perimeter bund capping material
Area D Operational landfill – open waste	Daily / intermediate cover (layer 1)	6 – Sandy Loam (slightly compacted)	1.9x10 ⁻⁴	5x10 ⁻⁴	Adopted to represent daily or intermediate cover for open waste
Area D Operational landfill – worst case drainage	Drainage layer (layers 1 and 2)	21 – Gravel	3x	10-1	No change to USDA value. Adopted to represent worst case for rainfall infiltration and collection of leachate from mid-landfill drainage layer.
Area D Operational landfill – worst case drainage and future cap	BSL beneath mid- landfill drainage (layer 3 in worst case profile and layer 7 in future capped profile)	28 – Silty Clay (moderately compacted)	1.2x10 ⁻⁶	1x10 ⁻⁵	Adopted to represent soil layer beneath mid-landfill drainage
Area D Operational landfill – future	Topsoil and subsoil (layers 1 & 2)	14 – Silty Clay (slightly compacted)	2.5x10 ⁻⁵		No change to USDA value. Top soil and subsoil will be sourced from clean imported soils. Minor compaction and contouring assumed.
cap	Barrier Clay (layer 3)	28 – Silty Clay (moderately compacted)	1.2x10 ⁻⁶	1x10 ⁻⁵	Cap is designed to meet specification of 1x10 ⁻⁷ m/s. Moderate compaction assumed.
	Soil (layer 4)	28 – Silty Clay (moderately compacted)	1.2x10 ⁻⁶	1x10 ⁻⁴	Represents intermediate cover which will be present before placement of final cap.

Profile	Soil Unit	Adopted USDA Soil Texture ⁽¹⁾	USDA saturated hydraulic conductivity (cm/s)	Adopted saturated hydraulic conductivity (cm/s)	Justification
All Landfill Profiles (landfill waste and	Municipal Waste	18 - Municipal Waste (312 kg/m³)	1x10 ⁻³	1x10 ⁻⁵	A large proportion of waste received (approx. 50%) is soils, which has reduced the hydraulic conductivity. Municipal waste at Green Island landfill reported to have density > 1,000 kg/m³.
natural ground)	Natural ground	28 – Silty Clay (moderately compacted)	1.2x10 ⁻⁶	5x10 ⁻⁷	UKEM and LKEM units reported horizontal hydraulic conductivity range of 1x10 ⁻⁶ – 1x10 ⁻⁷ m/s. Significant anisotropy assumed due to variable thin beds of sand, silty sand, sandy silt, silt, clayey silt and silty clay, and value calibrated to recorded leachate head.

¹⁾ Schroeder et al., 1994.

Table F.2 Scenario testing for landfill cap hydraulic conductivity

Landfill Area	Scenario	Hydraulic conductivity of capping layer (Layer 3) (cm/s)
Area B – Operational landfill – existing cap	1 – Base case (as per Table 1)	1x10 ⁻⁶
	2 – Reduced hydraulic conductivity by 1 order of magnitude	1x10 ⁻⁷
Area D – Operational landfill – future cap	1 – Base case (as per Table 1)	1x10 ⁻⁵
	2 – Reduced hydraulic conductivity by 1 order of magnitude	1x10 ⁻⁶
	3 – Reduced hydraulic conductivity by 2 orders of magnitude	1x10 ⁻⁷

²⁾ Berger and Schroeder, 2013.

F-1.2 Landfill base and drainage layers

HELP models include the functionality to apply the parameters of slope and flow path distance to lateral drainage layers (LDL) that overlay landfill liners or barrier soil layers (BSL) within the landfill profile. The existing Green Island landfill has no formal landfill liner or internal leachate collection system, however the underlying natural geology demonstrates low vertical permeability. Therefore, although a BSL was included within the landfill profiles to simulate the base of the landfill, the overlying layer of municipal waste was modelled as a vertical percolation layer (VPL), not an LDL. Lateral movement of leachate within the waste immediately above the base of the landfill is not expected to be occurring, particularly as the base of the landfill is inferred to be relatively flat.

The design of the operational area of the landfill (Area D) does however include a barrier soil layer and overlying leachate drainage collection system in the middle of the waste profile, which has been installed to capture leachate generated above this layer. The slope and average flow path distance to the leachate collection drains for the modelled LDL is presented in Table F.3. Sensitivity testing of the average flow path distance within the Area D future capped landfill profile indicated that this parameter has no influence on the results as no perching is expected to occur to generate horizontal flow towards the drains. A distance of 15 m was adopted for the Area D future capped profile. For the Area D worst case landfill drainage profile average flow path distances of both 15 m and 75 m were modelled, with the combined results assuming that 1/3 of Area D represents the average conditions in the southern section and 2/3 the average conditions in the northern section.

LDLs are also incorporated above BSLs representing the landfill cap in all landfill profiles which include capping layers (Table F.2). Due to significant differences in the slope and average flow path distance for the eastern and western sections of Area D, two models were used to assess the two different scenarios. The results for Area D (future capped) averaged the results from these two profiles, as each scenario is assumed to represent conditions across approximately 50% of Area D.

Table F.3 Landfill	area	and	slope	of	drainage	layer
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Landfill area and stage	Description of layer	Slope (%)	Average flow path distance to leachate collection / drain (m)		
Area D Operational landfill – worst case landfill drainage	Mid landfill drainage	4.5	15 – South		
Area D Operational landfill – future capped area	Mid-landfill drainage	4.5	75 – North		
Area A Transfer station – capped	Drainage above cap	2	100		
Area B Operational landfill – existing capped	Drainage above cap	4.5	100		
Area D Operational landfill – future capped	Drainage above cap	21 – West 3.8 – East	70 – West 175 – East		

F-1.2 Weather and rainfall runoff

Synthetic weather files were generated for a 50-year period using a weather generator model (WGEN), a stochastic model used to generate daily weather variables. Taulis and Milke (2005) from the University of Canterbury developed climate parameters for Dunedin, Otago, which were used within the WGEN to generate inputs for the parameters 'Solar Radiation' and 'Growing Season'. The NIWA weather station located at Musselburgh, Dunedin (Agent numbers 15752 and 5402) was used within WGEN to generate inputs for all other parameters, as described in Table F-4 (NIWA, 2022). Detailed input data are presented in Tables F.12 and F.13 (attached).

Table F.4 WGEN climate input data and sources

Parameter	Source and data period	Input format to WGEN
Precipitation	Daily rainfall NIWA station 5402: (May 1970 – August 1997) NIWA station 15752: (August 1997 – October 2022)	Monthly average
Temperature	Daily Tmax and Tmin NIWA station 5402: (May 1970 – August 1997) NIWA station 15752: (August 1997 – October 2022)	Average of daily Tmax and Tmin input as monthly average
Solar radiation	Taulis and Milke (2005) derived parameter for Dunedin, Otago (latitude -45.9)	Annual average solar radiation
Growing season	Taulis and Milke (2005) derived parameter for Dunedin, Otago (latitude -45.9).	'Start day' and 'End day' (Julian date) input to evapotranspiration parameter
Wind speed	Mean wind speed (annual): NIWA station 5402 (1982, 1988, 1989) NIWA station 15752 (1988 – 2021)	Annual average input to evapotranspiration parameter
Relative humidity	Monthly mean of 9am relative humidity: NIWA station 5402 (Jan 1972 – Aug 1997) NIWA station 15752 (Sep 1997 – Oct 2022)	Average per quarter input to evapotranspiration parameter

In developing the HELP climate data, the user defined parameters and values presented in Table F.5 were also included within the input data for evapotranspiration:

Table F.5 Evapotranspiration input data for WGEN

Landfill area and stage	Maximum leaf area index	Evaporative zone depth (cm)
Area A – Transfer station (capped)	1	20
Area B – Operational landfill (existing capped)	2	10
Area C – Perimeter bund (existing capped)	2	20
Area D – Operational landfill (existing open waste)	0	10
Area D – Operational landfill (worst case drainage)	0	10
Area D – Operational landfill (future capped)	3.5	15

Rainfall runoff within the HELP model is determined through application of a runoff curve number. The parameters utilised for determination of the curve number include soil texture, surface slope, compaction, surface vegetation and average surface flow path length. The adopted criteria for each modelled landfill profile are presented in Table F.13 (attached). As discussed in the previous section, the results for the Area D worst case drainage scenario assume the southern area represents average conditions across 1/3 of Area D and the northern area represents average conditions across 2/3 of Area D. The results for the Area D future capped scenario assume a 50% distribution of the average conditions in the east and west.

Table F.6 Parameters for determination of runoff curve number for each landfill area and development stage

Landfill area and stage	Soil Texture	Average Surface Flow Path Length (m)	Surface slope	Surface Vegetation	
Area A – Transfer station (capped)	Sandy loam – slightly compacted	100	2%	Poor grass	
Area B – Operational landfill (existing capped)	andfill (existing compacted		4.5%	Fair grass	
Area C – Perimeter bund (existing capped)	,,,		20%	Fair grass	
Area D – Operational landfill (existing open waste)	Sandy loam – slightly compacted	60	0.1%	Bare soil	
Area D – Operational landfill (worst case drainage)	Gravel	15 – South 75 – North	5%	Bare soil	
Area D – Operational landfill (future capped)	Silty clay – slightly compacted	70 – West 175 – East	21% – West 3.8% – East	Good grass	

F-2 Results

The results of the HELP modelling are presented in Tables F.14 and F.15 (attached). The reported results represent the annual average output values over the entire model period which incorporates the 50-year stochastic weather files. The results are therefore anticipated to increase or decrease in response to above or below average climatic conditions.

F-2.1 Infiltration

Infiltration as a percentage of total rainfall across the different landfill areas and scenarios are presented in Table F.7.

Table F.7 Infiltration as a percentage of total rainfall

Landfill area and stage	Scenario	Infiltration (%)
Area A – Transfer station (capped)	1 – Base case	32
Area B - Operational landfill (existing	1 – Base case	26
capped)	2 – Reduced hydraulic conductivity by 1 order of magnitude	5
Area C - Perimeter bund (existing capped)	1 – Base case	20
Area D - Operational landfill (existing open waste)	1 – Base case	32
Area D - Operational landfill (worst case drainage) (South)	1 – Base case	34
Area D - Operational landfill (worst case drainage) (North)	1 – Base case	47
	1 – Base case	33
Area D - Operational landfill (future capped) (West and East)	2 – Reduced hydraulic conductivity by 1 order of magnitude	28
	3 – Reduced hydraulic conductivity by 2 orders of magnitude	6

F-2.2 References

Berger, K. and Schroeder, P.R. 2013. The Hydrologic Evaluation of Landfill Performance (HELP) Model. User's Guide for HELP-D (Version 3.95D). 6th revised edition for version HELP 3.95D.

NIWA, 2022. The national climate database. https://cliflo.niwa.co.nz/ [Accessed November 2022]

Schroeder, P. R., Dozier, T.S., Zappi, P. A., McEnroe, B. M., Sjostrom, J.W., and Peyton, R. L. (1994). "The Hydrologic Evaluation of Landfill Performance (HELP) Model: Engineering Documentation for Version 3," EPA/600/R-94/168b, September 1994, U.S. Environmental Protection Agency Office of Research and Development, Washington, DC.Taulis, M.E. and Milke, M.W. Estimation of WGEN weather generation parameters in arid climates. Ecological Modelling 184: 177-191

Table F.8 Area A Transfer Station HELP Model soil profile

Scenario	Layer No.	Layer Description	Soil Texture Description*	Layer Type**	Layer Thickness (cm)	Saturated Hydraulic Conductivity (cm/s)	Total Pore Volume (Vol/Vol)	Field Capacity (Vol/Vol)	Wilting Point (Vol/Vol)
Transfer Station -	1	Topsoil	Sandy loam – slightly compacted	VPL	20	5.0E-04	0.452	0.19	0.085
Capped	2	Сар	Sandy loam – slightly compacted	BSL	50	5.0E-04	0.453	0.19	0.085
	3	Municipal waste	Municipal waste	VPL	700	1.0E-05	0.671	0.292	0.077
	4	Natural ground	Silty clay - moderately compacted	BSL	500	5.0E-07	0.452	0.411	0.311

^{*} U.S. Department of Agriculture (USDA) soil textural classification system reported in Schroeder et al., (1994).

Table F.9 Area B Operational landfill – existing capped HELP Model soil profile

Scenario	Layer No.	Layer Description	Soil Texture Description*	Layer Type**	Layer Thickness (cm)	Saturated Hydraulic Conductivity (cm/s)	Total Pore Volume (Vol/Vol)	Field Capacity (Vol/Vol)	Wilting Point (Vol/Vol)
Operational	1	Topsoil	Silty clay - slightly compacted	VPL	10	2.5E-05	0.479	0.371	0.251
landfill – capped area	2	Subsoil	Silty clay - slightly compacted	LDL	10	2.5E-05	0.479	0.371	0.251
sapped area	3	Clay cap	Silty clay – moderately compacted	BSL	150	1.0E-06	0.452	0.411	0.311
	4	Soil	Silty clay - slightly compacted	VPL	30	2.5E-05	0.479	0.371	0.251
	5	Municipal waste	Municipal waste	VPL	2000	1.0E-05	0.671	0.292	0.077
	6	Natural ground	Silty clay - moderately compacted	BSL	500	5.0E-07	0.452	0.411	0.311

^{*} U.S. Department of Agriculture (USDA) soil textural classification system reported in Schroeder et al., (1994).

^{**} VPL = Vertical Percolation Layer. BSL = Barrier Soil Layer.

^{**} VPL = Vertical Percolation Layer. LDL = Lateral Drainage Layer. BSL = Barrier Soil Layer.

Table F.10 Area C Perimeter Bund HELP Model soil profile

Scenario	Layer No.	Layer Description	Soil Texture Description*	Layer Type**	Layer Thickness (cm)	Saturated Hydraulic Conductivity (cm/s)	Total Pore Volume (Vol/Vol)	Field Capacity (Vol/Vol)	Wilting Point (Vol/Vol)
Operational landfill –	1	Сар	Silty clay – moderately compacted	VPL	500	1.0E-05	0.452	0.411	0.311
capped area	2	Municipal waste	Municipal Waste	VPL	500	1.0E-05	0.671	0.292	0.077
	3	Natural ground	Silty clay - moderately compacted	BSL	500	5.0E-07	0.452	0.411	0.311

^{*} U.S. Department of Agriculture (USDA) soil textural classification system reported in Schroeder et al., (1994).

^{**} VPL = Vertical Percolation Layer. BSL = Barrier Soil Layer.

Table F.11 Area D Operational landfill – open waste, worst case drainage and future capped HELP Model soil profiles

Scenario	Layer No.	Layer Description	Soil Texture Description*	Layer Type**	Layer Thickness (cm)	Saturated Hydraulic Conductivity (cm/s)	Total Pore Volume (Vol/Vol)	Field Capacity (Vol/Vol)	Wilting Point (Vol/Vol)
Operational landfill – open waste	1	Daily / intermediate cover	Sandy loam – slightly compacted	VPL	10	5.0E-04	0.452	0.19	0.085
	2	Municipal waste	Municipal waste	VPL	1000	1.0E-05	0.671	0.292	0.077
	3	Natural ground	Silty clay - moderately compacted	BSL	500	5.0E-07	0.452	0.411	0.311
Operational	1	Drainage layer	Gravel	VPL	10	3.0E-01	0.397	0.032	0.013
landfill – worst case	2	Drainage layer	Gravel	LDL	20	3.0E-01	0.397	0.032	0.013
drainage	3	Soil below drainage	Silty clay – Moderately compacted	BSL	500	1.0E-05	0.452	0.411	0.311
	4	Municipal waste	Municipal waste	VPL	1000	1.0E-05	0.671	0.292	0.077
	5	Natural ground	Silty clay - moderately compacted	BSL	500	5.0E-07	0.452	0.411	0.311
Operational	1	Topsoil	Silty clay - slightly compacted	VPL	15	2.5E-05	0.479	0.371	0.251
landfill – future	2	Subsoil	Silty clay - slightly compacted	LDL	20	2.5E-05	0.479	0.371	0.251
capping	3	Clay cap	Silty clay – moderately compacted	BSL	60	1.0E-05	0.452	0.411	0.311
	4	Intermediate cap	Silty clay - slightly compacted	VPL	30	2.5E-05	0.479	0.371	0.251
	5	Municipal Waste	Municipal waste	VPL	1000	1.0E-05	0.671	0.292	0.077
	6	Drainage layer	Municipal waste	LDL	200	1.0E-05	0.671	0.292	0.077
	7	Soil below drainage	Silty clay – Moderately compacted	BSL	500	1.0E-05	0.452	0.411	0.311
	8	Municipal waste	Municipal waste	VPL	1000	1.0E-05	0.671	0.292	0.077
	9	Natural ground	Silty clay - moderately compacted	BSL	500	5.0E-07	0.452	0.411	0.311

^{*} U.S. Department of Agriculture (USDA) soil textural classification system reported in Schroeder et al., (1994).

^{**} VPL = Vertical Percolation Layer. LDL = Lateral Drainage Layer. BSL = Barrier Soil Layer.

Table F.12 Average monthly values used in 50 – year synthetic weather file generation for daily precipitation and daily mean temperature (WGEN)

Month	Average monthly rainfall (mm) (1)	Average mean monthly temperature (C) (1)			
January	70.0	15.2			
February	61.0	15.2			
March	60.0	14.0			
April	57.4	12.0			
May	67.2	9.5			
June	65.2	7.4			
July	63.8	6.7			
August	57.9	7.8			
September	48.7	9.4			
October	64.0	10.9			
November	60.5	12.4			
December	74.0	14.1			
1) Data from Musselburgh, Dunedin. May 1970 – October 2022 (NIWA, 2022)					

Table F.13 Parameter values adopted in the solar radiation and evapotranspiration input data files

Parameter	Adopted values	Data source	
Solar radiation	12.2	Taulis and Milke (2005) value for Dunedin, Otago base	
Growing season (Julian date)	Start day: 274 End day: 171	on latitude of -45.9	
Wind speed	Yearly average: 13.3 km/hour	Musselburgh, Dunedin (1982 1988, 1989, 1998 – 2021) (NIWA, 2022)	
Average relative humidity	First quarter: 75.82% Second quarter: 78.47% Third quarter: 76.54% Fourth quarter: 71.5%	Musselburgh, Dunedin (January 1972 – October 2022) (NIWA, 2022)	

Table F.14 HELP Model Leachate Results for closed landfill areas

Landfill Area and Stage	Scenario	Average annual total runoff (mm/year) / m ²	Average annual evapotranspiration (mm/year) / m²	Average annual percolation to waste (mm/year) / m²
Area A Transfer Station - capped	1 – Base case	0.60	516.20	237.75
Area B	1 – Base case	84.93	475.23	193.88
Operational Landfill – existing capped area	2 – Reduced hydraulic conductivity by 1 order of magnitude	212.73	506.35	34.31
Area C Perimeter Bund	1 – Base case	94.32	493.25	153.01
Area D	1 – Base case	9.32	498.03	247.12
Operational Landfill – future capped area (West)	2 – Reduced hydraulic conductivity by 1 order of magnitude	40.96	497.65	211.67
	3 – Reduced hydraulic conductivity by 2 orders of magnitude	182.25	512.68	44.43
Area D	1 – Base case	6.94	497.95	249.66
Operational Landfill – future capped area (East)	2 – Reduced hydraulic conductivity by 1 order of magnitude	41.75	497.73	214.74
	3 – Reduced hydraulic conductivity by 2 orders of magnitude	196.09	512.5	44.83

Table F.15 HELP Model leachate results for operational landfill areas

Landfill Area and Stage	Scenario	Average annual total runoff (mm/year) / m ²	Average annual evapotranspiration (mm/year) / m²	Average annual percolation to waste (mm/year) / m²
Area D Operational Landfill – open waste	1 – Base case	4.19	471.41	240.75
Area D Operational landfill – worst case drainage (South)	1 – Base case	0.26	308.41	258.18
Area D Operational landfill – worst case drainage (North)	1 – Base case	0.19	300.11	356.86

Appendix G Seep/W modelling

G-1 Groundwater modelling (SEEP/W)

G-1.1 Model set-up

The groundwater assessment included 2D modelling using Geostudio 2021 SEEP/W finite element modelling software. Modelling was undertaken to estimate the seepage into the leachate collection drain and to simulate the leachate head within the landfill.

Two SEEP/W cross-sections were created to model the landfill. The location of the cross-section lines is shown in Figure G.1. The models were created based on a drone survey provided by DCC. Each model was initially run under steady-state conditions to simulate the interpreted baseline groundwater conditions. Model were calibrated to:

- Measured leachate head within the landfill (average level).
- Combined pump station flows (from leachate collection trench), under dry weather conditions.

Steady state model scenarios were then run to simulate future conditions at closure, in particular, capping of the landfill and installation of a leachate collection drain in the southern valley. The Line 2 future scenarios include additional filling as outlined in the design report.

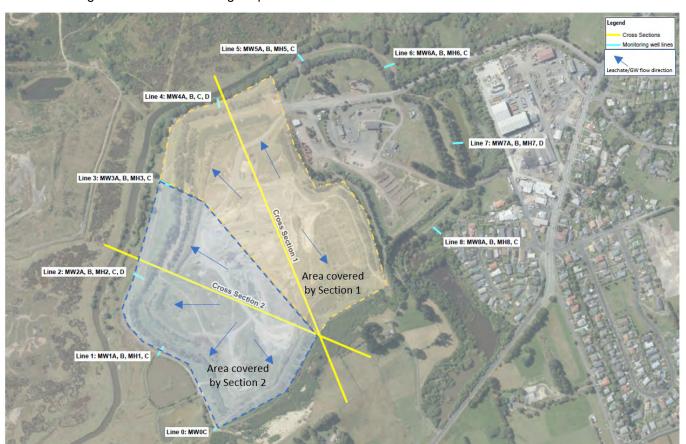


Figure G.1 Location of SEEP/W cross sections

The SEEP/W models adopt the recharge rates as calculated from the HELP modelling assessment (Appendix F), however the recharge rates have been scaled to reflect the actual area and volume of landfill covered by each cross section (Figure G.2 and G.3). The recharge applied reflects to the proportion of rainfall that infiltrates the proposed landfill cap, and has been applied to the layer below the landfill cap (except in un-capped areas where it has been applied to surface). Recharge rates and scaling factors are described in Section G.1.3.

Both model cross sections intercept the existing leachate trench and proposed southern valley trench. The estimated flow rates (per 1 m of trench) into these trenches are presented in the results. These results have been scaled up by the length of the trench.

To augment the leachate interception trench, the current landfill operation has progressively installed trenched leachate drains over the intermediate cover soils in the location of the southern progression of waste placement as well as in the northern sector of waste placed in 2019-2022. It is proposed that additional shallow (0.5 m) leachate drains will be installed in the fill material in the southern portion of the landfill – the area intersected by Line 2. However, given the heterogeneity of the fill materials it is considered that the effect of the drains may be localised. The steady-state model used in this assessment does not simulate variable recharge (from rain events) nor the fill heterogeneity and therefore may overestimate the effect of the drains in reducing leachate head within the fill. For this reason, the future modelling scenarios exclude the shallow leachate drains.

The effect of future sea level rise was considered in the future modelling scenarios. Sea level rise effects were modelled with a higher (0.5 m) river level boundary condition. For this scenario the road/bund level was also increased by 0.5 m as recommended in the *Design Report*.

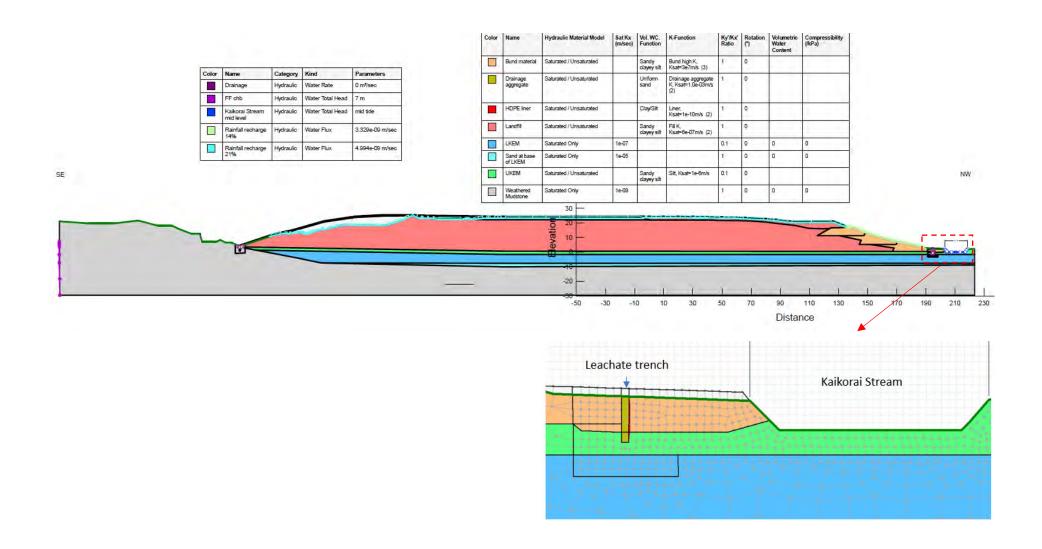


Figure G.2 Line 1 – SEEP/W model set up (current scenario)

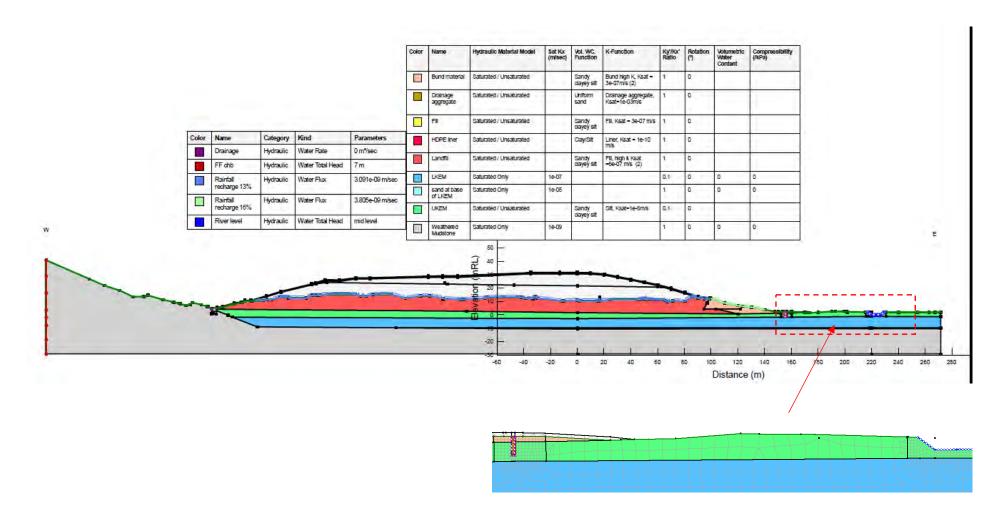


Figure G.3 Line 2 – SEEP/W model set up (current scenario)

G-1.2 Material Properties

Material properties used in the model scenarios considered are summarised in Table G.1.

Hydraulic conductivity values used in modelling were based on the results of hydraulic conductivity testing undertaken as part of the site investigations. In soil units where test results were not available, widely reported in the literature values were used to approximate hydraulic conductivity values.

It is likely that the waste material is highly heterogeneous, with some areas having a higher hydraulic conductivity than other areas depending on the type of waste materials and the degree of compaction undertaken. Hydraulic conductivities of municipal solid waste (MSW) reported in the literature vary between 1 x 10 ⁻³ and 1 x 10⁻⁹ m/s, although most values are in the range of 10⁻⁵ to 10 ⁻⁶ m/s (LANDSS⁶). It is understood that soil makes up a significant proportion of the waste material at GILF, which is likely to reduce the bulk hydraulic conductivity of the landfill material. The best calibration was achieved using a waste hydraulic conductivity of 6 x 10 ⁻⁷ m/s.

In general, for natural soils/materials, the hydraulic conductivity perpendicular to the soil layers is slower than the conductivity parallel to soil layers. Therefore, the underlying estuarine and marine sediments (UKEM/LKEM) were assigned a Ky/Kx ratio (anisotropy) of 0.1.

A section of the landfill has recently been capped (2022), however the base case model scenarios do not include the final capping material as the current leachate head level is unlikely to reflect these recent changes.

Material	Hydraulic material model	Volumetric water content function	Hydraulic conductivity (m/s)	K _y /K _x
Bund Material	Saturated/Unsaturated	Sandy clayey silt	3 x 10 ⁻⁷	1
Landfill material	Saturated/Unsaturated	Sandy clayey silt	6 x 10 ⁻⁷	1
Drainage Aggregate (trench)	Saturated/Unsaturated	Uniform sand	1 x 10 ⁻³	1
HDPE liner	Saturated/Unsaturated	Clay/silt	1 x 10 ⁻¹⁰	1
UKEM	Saturated/Unsaturated	Sandy clayey silt	1 x 10 ⁻⁶	0.1
LKEM	Saturated only	-	1 x 10 ⁻⁷	0.1
Sand at base of LKEM	Saturated only	-	1 x 10 ⁻⁵	1
Weathered Mudstone	Saturated only	-	1 x 10 ⁻⁹	1
Topsoil	Saturated/Unsaturated	Silt Loam	7 x 10 ⁻⁷	1
Cap (10 ⁻⁷)	Saturated/Unsaturated	Clay/Silt	1 x 10 ⁻⁷	1
Cap (10 ⁻⁸)	Saturated/Unsaturated	Clay/Silt	1 x 10 ⁻⁸	1
Cap (10 ⁻⁹)	Saturated/Unsaturated	Clay/Silt	1 x 10 ⁻⁹	1

G-1.3 Boundary Conditions

Far field constant head boundaries were set on the landward edge of the model to approximate inferred groundwater level. River level was inferred based on site observations (average water level when the estuary mouth is not blocked) (Table G.2).

⁶ Landfill (Aftercare) Decision Support System, University of Southampton https://landss.soton.ac.uk/

Table G.2 Boundary conditions used in SEEP/W modelling

Type	Value	Details
Constant head (Groundwater head)	7 m RL	Applied to landward (Southern) edge of model
Constant head with potential seepage face applied (Stream level)	Line 1 – 1.1 m RL Line 2 – 0.8 m RL	Approximate river level Applied on creek/stream beds to represent water level when the estuary mouth is not blocked*.
	Line 1 – 1.6 m RL Line 2 – 1.3m RL	Sea level rise scenario Applied on creek/stream beds to represent water level when the estuary mouth is not blocked*.
Seepage face	0 m ³ /s with potential seepage face applied	Applied to leachate collection trench to simulate pumping from leachate collection system

^{*}Higher river levels occur from time to time when the river/estuary mouth is blocked

Rainfall recharge was applied to both the landfill and bund surfaces. For the future capped landfill scenarios, the rainfall infiltration was applied to the underside of the landfill cap, as this value represented the rainfall going through the cap (taking into account rainfall and evapotranspiration). The amount of rainfall recharge was based on the outputs of the HELP modelling assessment. However, as the SEEP/W modelling assessment utilises two cross sections to represent a 3D landform, the recharge was scaled down to reflect the average infiltration rate over each landfill sub-area. Without scaling, the models would have overestimated the infiltration into the landfill material and leachate trench flow rates. However, this approach (averaging of rainfall recharge) may underestimate the leachate head level in the centre of the landfill. Scale factors are shown in Table G.3 with scaled infiltration rates included in Table G.4.

Table G.3 Recharge scale factors

Area		Area (ha)	Landfill section length (in model, m)	Length of trench covered by model section (m)	Area represented by model (ha)	Recharge scale factor
Area 1	Landfill	12.9	480	400	19.2	0.67
(Section Line 1)	Bund	1.4	50		2.0	0.7
Area 2	Landfill	11.0	420	620	26.0	0.42
(Section Line 2)	Bund	2.5	50		3.1	0.8

Table G.4 Scaled rainfall recharge applied in SEEP/W models

Section Line	Material / Surface	Scaled infiltration rate (% of annual rainfall)
Section Line 1	Uncapped	21%
	Cap (10 ⁻⁷ m/s)	21%
	Cap (10 ⁻⁸ m/s)	17.5%
	Cap (10 ⁻⁹ m/s)	3%
	Bund	14%
Section Line 2	Uncapped	13%
	Cap (10 ⁻⁷ m/s)	13%
	Cap (10 ⁻⁸ m/s)	12%
	Cap (10 ⁻⁹ m/s)	2.5%
	Bund	16%

G-1.4 Model assumptions and limitations

- It is assumed that hydraulic conductivities for the landfill and bund materials adopted in the model simulations
 are appropriate average values, given the likely heterogeneity of the fill materials.
- The field investigations have identified significant variation in the underlying marine and estuarine sedimentary deposits (UKEM/LKEM). Bulk hydraulic conductivities have been applied to simulate the measured pumping rates, however it is still possible that there may be localised deviation from the simulated results
- It is assumed that the underlying sedimentary deposits are anisotropic given the nature of the depositional environment.

G-1.5 Model results – existing conditions

Leachate Trench

The results of the modelling for the base case (current) simulations are summarised in Table G.5. The average combined flow rate to the leachate trench is approximately 1-2 L/s, this equates to a rate between 6 x 10^{-7} m³/s and 1.2 x 10^{-6} m³/s per metre of trench (based on a trench length of 1674 m). The results of the base case simulations are within this range, slightly higher flow rates are estimated for Section Line 1, likely due to the proximity to Kaikorai Stream.

As noted in section, 2.2.3, higher flows do occur periodically, associated with wet conditions and additional stormwater flows which are diverted to the leachate system.

Table G.5 Model results – flow into leachate collection trench

Section	Modelled flow rate into trench (m³/m of trench)	
Line 1	1.4 x 10 ⁻⁶	
Line 2	7.5 x 10 ⁻⁷	

The relative proportion of flows from each side of the trench was calculated from the models (Table G.6). Approximately 70% of the flow to the trench comes from the landfill (and underlying groundwater) and 30% from the direction of the stream. The presence of the HDPE liner in the trench limits flows from surface water, however the model indicates water flow from the stream side into the base of the trench.

Table G.6 Relative proportion of flows into leachate collection trench

Section Line	Proportion from landfill side	Proportion from stream side
Line 1	0.68	0.32
Line 2	0.73	0.27

The models were used to test the effect on surface and groundwater if the leachate trench was not operational (i.e pumps turned off). As the modelling has been undertaken on a steady state basis, the model assumes that the leachate trench has never been operational, when in reality it is likely to be short term failure only. Therefore, the leachate breakout shown in Figure G.4 is unlikely to be realised with a short term failure. Furthermore, it is considered unlikely that all pump stations would be out of action for an extended period given historical performance and additional mitigation measures that are proposed in the *Design Report*. However for the purposes of this assessment, the models can be used to estimate the rate of seepage into surface water based on the permeability of the sedimentary units. The model estimates a seepage rate between 4.2 x 10 ⁻⁷ and 4.7 x 10 ⁻⁷ m³/s per m of stream, this equates to a rate of ~0.5 L/s for the entire stream adjacent to the site (approximately length of 1 km).

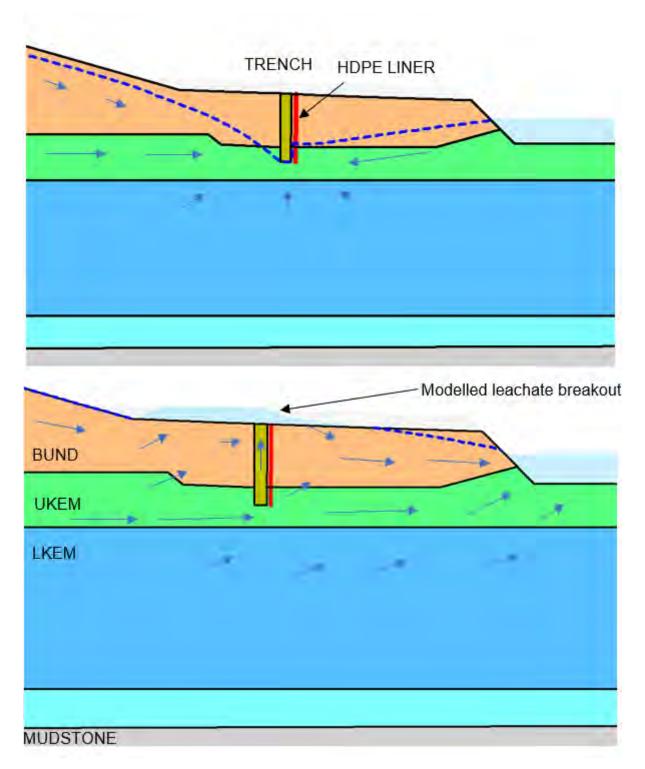


Figure G.4 Section Line 1 – leachate and groundwater flow paths with (top) and without trench (bottom) operational.

Leachate Head (in landfill)

The simulated water level (representing landfill leachate) in shown in Figure G.5. The modelled leachate level in the centre of the land fill is 16 m RL in Section Line 1. This is similar to the level recorded in gas wells GW17, 13 and the leach riser (16-17 m RL) (see Appendix D) but less than recorded in GW14 and GW 15 (22 m RL). As noted above, the rainfall recharge rates were scaled to represent the actual area represented by the model section. This approach may result in averaging or flattening of the modelled leachate head. However, in general the models predict leachate close to the surface in the bund materials. Observations made during a test pit

investigation of the bund materials (Tonkin and Taylor, 2020) suggest that the leachate level is close to the surface in some areas. It is understood that leachate breakout from the bund has occurred from time to time, but has been addressed by the installation of gravel drains in these areas.

In Line 2, the simulated leachate level in the centre of the section is approximately 10 m RL. As the leachate survey focussed on older parts of the landfill, there is limited information available for calibration in this area. The leachate level recorded in Piezo 3 and Piezo 4, located south of the section line, was 8.7 and 8.2 m RL respectively.

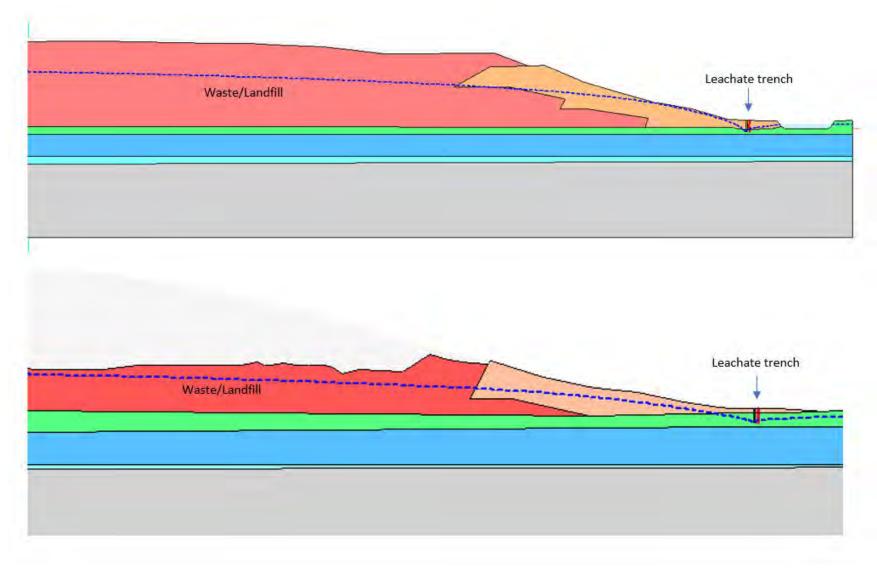


Figure G.5 Leachate head in landfill -Line 1 (top) and Line 2 (bottom) – current scenario

G-1.6 Future modelling scenarios

The base case SEEP/W models were modified to simulate conditions at closure (capping and leachate management). The future scenarios tested three different cap permeabilities (10⁻⁷ to 10⁻⁹ m/s). Additional scenarios were run to test the effect of trench failure and sea level rise effects. The model scenarios are listed in Table G.7. The future modelling scenarios include the installation of the proposed southern valley trench.

Table G.7 Future modelling scenarios

Scenario Number	Hydraulic conductivity of cap		Description
	Currently uncapped*	Capped areas (north/central, applies- Section Line 1 only)	
1A	1 x 10 ⁻⁷ m/s	1 x 10 ⁻⁸ m/s	Leachate trenches operational
1B	1 x 10 ⁻⁷ m/s	1 x 10 ⁻⁸ m/s	Trench" turned off"
1C	1 x 10 ⁻⁷ m/s	1 x 10 ⁻⁸ m/s	Trench operational, Sea level rise (+ 0.5 m) applied to river boundary.
2A	1 x 10 ⁻⁸ m/s (all areas)		Leachate trenches operational
2B	1 x 10 ⁻⁸ m/s (all areas)		Trench" turned off"
2C	1 x 10 ⁻⁸ m/s (all areas)		Trench operational, Sea level rise (+ 0.5 m) applied to river boundary.
3A	1 x 10 ⁻⁹ m/s (all areas)		Leachate trenches operational

^{*}Only applies to areas not currently capped. It is understood that the northern and central part of the landfill (Zone B in Figure G.1, Appendix G) has recently been capped with low permeability materials (achieving 10⁻⁸ 10⁻⁹ m/s in lab tests). For the purpose of this assessment, it has been assumed that the hydraulic conductivity is 10⁻⁸ of this new cap is at least 10⁻⁸ m/s.

Current and future scenario modelling results are summarised in Tables G.8 and G.9. Model flow rates (per metre) are scaled up by the length of trench each model section covers. This does not include the leachate trench flows from pump stations 5 to 8 (transfer station area) which contribute to the overall dry weather flow rate of 1-2 L/s. However, trench flow rates are expected to be lower than in the modelled sections, as the volume (and height) of filling is less.

Surface water interaction

The models were used to estimate the flow rate into or from (stream depletion) Kaikorai Stream. These results are shown Table G.9. In Section Line 1, the interaction with the stream is reversed depending on whether the leachate drain is "turned on".

For Section Line 2, the model predicts flows from groundwater into the stream in all model scenarios. The flow rate is more than an order of magnitude higher when the drain is turned off than when it is operating, showing that the leachate trench is effective in intercepting leachate/groundwater from the landfill. When the leachate trench is operating, flow vectors indicate the groundwater flow into the stream (~2 x 10 -8 m³/s per m, 0.01 L/s for a 600 m length) represents upward groundwater seepage from the underlying estuarine sediments. The stream is located ~70 m from the trench, therefore while the trench appears be effective at intercepting leachate from the landfill, it does not result in stream depletion and does not influence groundwater levels immediately adjacent to the stream.

Table G.8 SEEP/W model results – leachate trench and leachate head

Scenario	Section	Perimeter leachate drain flow rate		Southern Valley leachate drain flow rate		Modelling leachate
		m³/s per m of trench	L/s for length of trench modelled*	m ³ /s per m of trench	L/s for length of trench modelled*	head in fill (m RL)
Current	Line 1	1.4 x 10 ⁻⁶	0.5	-		16
	Line 2	7.5 x 10 ⁻⁷	0.5	-		10.2
1A	Line 1	1.3 x 10 ⁻⁶	0.5	7.4 x 10 ⁻⁷	0.1	15.4
	Line 2	7.0 x 10 ⁻⁷	0.4	5.0 x 10 ⁻⁷	0.2	9.8
1C	Line 1	1.4 x 10 ⁻⁶	0.6	7.4 x 10 ⁻⁷	0.1	15.4
	Line 2	7.3 x 10 ⁻⁷	0.5	5.0 x 10 ⁻⁷	0.2	9.9
2A	Line 1	1.3 x 10 ⁻⁶	0.5	6.4 x 10 ⁻⁷	0.1	14
	Line 2	6.6 x 10 ⁻⁷	0.4	4.5 x 10 ⁻⁷	0.1	9.3
2C	Line 1	1.4 x 10 ⁻⁶	0.6	6.4 x 10 ⁻⁷	0.1	14
	Line 2	7.0 x 10 ⁻⁷	0.4	4.5 x 10 ⁻⁷	0.1	9.3
3A	Line 1	6.2 x 10 ⁻⁷	0.3	7.8 x 10 ⁻⁸	0.01	4
	Line 2	3.0 x 10 ⁻⁷	0.2	7.2 x 10 ⁻⁸	0.02	3

Flow rate (per metre) multiplied by length of trench represented by each section:

Line 1 – Perimeter trench, 400 m, Southern Valley trench (150 m)

Line 2 – Perimeter trench, 620 m, Southern Valley trench (300 m)

Table G.9 SEEP/W model results – leachate trench and leachate head

Scenario	Section	Flow from strea	am (stream depletion)	Flow to stream	
		m³/s per m of trench	L/s for length of trench modelled*	m³/s per m of trench	L/s for length of trench modelled*
Current	Line 1	1.2 x 10 ⁻⁷	0.05	-	-
	Line 2	-	-	2.1 x 10 ⁻⁸	0.01
Current	Line 1	-	-	4.2 x 10 ⁻⁷	0.2
(trench off)	Line 2	-	-	4.7 x 10 ⁻⁷	0.3
1A	Line 1	1.2 x 10 ⁻⁷	0.05	-	-
	Line 2	-	-	1.6 x 10 ⁻⁸	0.01
1B	Line 1	-	-	5.4 x 10 ⁻⁷	0.2
(trench off)	Line 2	-	-	8.9 x 10 ⁻⁷	0.5
1C	Line 1	2.4 x 10 ⁻⁷	0.05	-	-
	Line 2	-	-	1.2 x 10 ⁻⁸	0.01
2A	Line 1	1.4 x 10 ⁻⁷	0.1	-	-
	Line 2	-	-	1.2 x 10 ⁻⁸	0.01
2B	Line 1	-	-	6.7 x 10 ⁻⁷	0.3
(trench off)	Line 2	-	-	8.3 x 10 ⁻⁷	0.5
2C	Line 1	2.5 x 10 ⁻⁷	0.1	-	-
	Line 2	-	-	1.7 x 10 ⁻⁸	0.01
3A	Line 1	2.5 x 10 ⁻⁷	0.1	-	-
	Line 2	-	-	3.9 x 10 ⁻⁸	0.02

Flow rate (per metre) multiplied by length of stream/trench represented by each section:

Line 1 – Perimeter trench, 400 m

Line 2 – Perimeter trench, 620 m

G-1.7 SEEP/W Modelling Summary

A summary of the key finding from the modelling assessment is provided below in Table G.10.

Table G.10 Modelling summary

Feature	Summary
Cap permeability	Decrease in leachate flow rate and leachate head level when a lower permeability cap is applied. This effect is most pronounced when a cap hydraulic conductivity (K) of 10 ⁻⁹ m/s is applied. The difference between leachate flow and head for the K 10 ⁻⁷ and 10 ⁻⁸ m/s scenarios is small
River side leachate trench	The trench is effective at drawing down leachate levels and intercepting flows to the stream. When the trench drainage boundary condition is "turned off" for modelling purposes the flow rate to stream is estimated to be between 0.5-0.8 L/s.
	Modelled flow rates are similar to current dry weather leachate trench pump rates.
Southern valley trench	Estimated flow rates to the southern trench are 0.2-0.3 L/s (0.03 L/s for low permeability cap scenario).
Surface water interaction	The modelling indicates that stream water (and shallow groundwater) is intercepted by the leachate trench. Stream depletion rates are estimated to be in the order of 0.1 L/s for Section 1. Stream depletion rates are higher in the sea level rise scenarios due to the higher water levels.
	However, the Section 2 model indicates flow (at a very low rate, 0.01 L/s) into the stream from shallow groundwater adjacent to the stream. This is much lower than the flow to the stream (0.3-0.5 L/s) if the trench was "turned off" showing that the trench is effective at intercepting leachate. The small inflow (0.01 L/s) to the stream under operating conditions is interpreted to represent groundwater seeping from the stream bed and adjacent river bank as due to the distance the operation of the leachate trench does not influence groundwater levels immediately adjacent to the stream.

Appendix B Bore logs RRPP wells

Project: Green Island Landfill RRPP Site Investigation Hole No. : BH01 : Dunedin City Council : 1 of 1 Sheet Site : Green Island Landfill : 10.95m Hole Length Job Number: 12547621 Scale @ A4 : 1:60 Commenced: 28/09/2021 Logged : HE Completed: 28/09/2021 Northing: 1399359.165 Easting: 4913129.948 System: NZTM2000 Processed : CJ RL: Datum: Ground level Checked : CG Sample **Geological Unit** Moisture condition 8 ** Estimated Consistancy / Relative density Defect Spacing (mm) Flush Return **Material Description** Water level Weathering TCR SCR RQD Depth (m) Number / Method Casing RL (m) Fill: GRAVEL, trace fine sand; dark grey, moist. 100mm Fill: Waste including metal, glass, wood, plastic fragments. Occasional layers of clay/silt. SNC Fill: Sandy SILT; brown to grey. Soft; moist. S 9.00 - 10.50 Sand is fine/micaecous. Rubber fragments to 100mm. 10 Slightly sandy SILT, trace clay; grey. Soft; moist. ·× 10.50 - 10.95 Sand is fine/micaceous. End of Hole @ 10.95m, Target Depth. **Ground Water Level** Orientation: Inclination: Vertical Notes and Comments: End of Hole @ 10.95m, Target Depth.

Monitoring well installation consisted of the following:

50 mm Blank pipe being installed between + 0.6 m a bgl - 2.1 m bgl Contractor: Speights Drilling Date Time 19/10/21 00:00 2.726 5.644 Equipment: Sonic Rig 50 mm Slotted pipe being installed between 2.1 - 5.1 m bgl. The stick up of the monitoring well is approximately 0.6 m agl. A

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metal casing was installed above ground.

Gravel pack was placed around the slotted portion of the pipe.

Bentonite was placed above the gravel to surface level

Project: Green Island Landfill RRPP Site Investigation Hole No. : BH02 Client : Dunedin City Council Sheet : 1 of 1 Site : Green Island Landfill : 10.95m Hole Length Job Number: 12547621 Scale @ A4 : 1:60 Commenced: 29/09/2021 Logged : HE Completed: 29/09/2021 Northing: 1399553.062 Processed : CJ Easting: 4912949.612 System: NZTM2000 RL: Datum: Ground Level Checked : CG Sample Geological Unit Moisture condition 8 Consistancy / Relative density Strength (MPa) Defect Spacing (mm) Flush Return **Material Description** Water level Estimated Depth (m) Number / Method Casing RL (m) SCR RQD FILL: Sandy SILT; trace clay with some gravel; grey, orange brown mottling. Stiff; moist, low plasticity [LANDFILL CAP]. FILL: Landfill waste, plastic, timber, organics; black. 4.35 - 4.40 Fine to medium SAND; grey 4.40 - 4.45 Sandy SILT, trace clay; grey, orange to brown mottling. Very Stiff; dry. Sand, fine. CORE LOSS FILL: Landfill waste CORE LOSS FILL: Landfill waste **CORE LOSS** Sandy SILT, trace clay; grey, mottled brown. Stiff; moist; medium plasticity [ALLUVIUM]. St × St М × × × × Slightly silty fine SAND; pale green to grey; moist to wet [ALLUVIUM]. M-W MD 10 -End of Hole @ 10.95m, Target Depth. **Ground Water Level** Inclination: Vertical Orientation: Notes and Comments: End of Hole @ 10.95m, Target Depth.

Monitoring well installation consisted of the following:

50 mm Blank pipe being installed between + 0.6 m a bgl - 2.9 m bgl Contractor: Speights Drilling Date Time 19/10/21 00:00 2.494 10.95 Equipment: Sonic Rig 50 mm Slotted pipe being installed between 2.9 - 5.9 m bgl. The stick up of the monitoring well is approximately 0.6 m agl. A

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metal casing was installed above ground.

Gravel pack was placed around the slotted portion of the pipe. Bentonite was placed above the gravel to surface level

Project: Green Island Landfill RRPP Site Investigation Hole No. : BH03 : Dunedin City Council Sheet : 1 of 1 Site : Green Island Landfill : 7.95m Hole Length Job Number: 12547621 Scale @ A4 : 1:60 Commenced: 28/09/2021 Completed: 28/09/2021 Logged : HE Easting: 4911134.71 Processed : CJ Northing: 1399523.518 System: NZTM2000 RL: Checked : CG Datum: Sample Geological Unit Moisture condition 8 [™]Estimated
[™]Strength (MPa) Consistancy / Relative density Defect Spacing (mm) Flush Return **Material Description** Water level TCR SCR RQD Depth (m) Number / Method Casing RL (m) FILL: Sandy fine to coarse GRAVEL, some cobbles; dark grey. Tightly packed. FILL: Landfill waste comprising of timber, plastics, metal. FILL: Sandy CLAY, trace gravel; Fragments of plastic, metal, timber, brick. Soft; moist. s FILL: Landfill waste comprising of metal, plastics, paper Slightly Sandy CLAY; grey. Mottled brown. Trace Organics. S М Soft; moist. Sandy CLAY; brown to grey. Trace organics. Soft; moist; End of Hole @ 7.95m, Target Depth. 10 -**Ground Water Level** Orientation: Inclination: Vertical Notes and Comments: End of Hole @ 7.95m, Target Depth.

Monitoring well installation consisted of the following:

50 mm Blank pipe being installed between + 0.6 m a bgl - 2.0 m bgl
50 mm Slotted pipe being installed between 2.0 - 5.0 m bgl.

The stick up of the monitoring well is approximately 0.6 m agl. A metal casing was installed above ground. Contractor: Speights Drilling Date Time 19/10/21 00:00 3.128 5.646 Equipment: Sonic Rig

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Gravel pack was placed around the slotted portion of the pipe. Bentonite was placed above the gravel to surface level

Project: Green Island Landfill RRPP Site Investigation Hole No. : BH04 Client : Dunedin City Council Sheet : 1 of 1 Site : Green Island Landfill : 10.95m Hole Length Job Number: 12547621 Scale @ A4 : 1:60 Commenced: 29/09/2021 Logged : HE Completed: 29/09/2021 Northing: 1399390.245 Processed : CJ Easting: 4913253.964 System: NZTM2000 Datum: Ground Level RL: Checked : CG Sample **Geological Unit** Moisture condition 8 Consistancy / Relative density Strength (MPa) Defect Spacing (mm) Flush Return **Material Description** Water level Weathering Estimated TCR SCR RQD Depth (m) Number / Method Casing RL (m) CAP: Sandy SILT; grey. Trace rootlets. FILL: Landfill waste comprising of plastic, wood, brick, basalt fragments. Trace gravel. Sandy SILT, trace clay; blue to grey. Orange to brown mottling. Soft; moist; medium plasticity [ALLUVIUM]. ٠× × ·×. Slightly sandy SILT, trace clay; grey to blue. Very soft; moist; medium plasticity. 10 -End of Hole @ 10.95m, Target Depth. **Ground Water Level** Orientation: Inclination: Vertical Notes and Comments:

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- End of Hole @ 10.95m, Target Depth.

 Monitoring well installation consisted of the following:

 50 mm Blank pipe being installed between + 0.6 m a bgl 3.0 m bgl

 50 mm Slotted pipe being installed between 3.0 6.0 m bgl.

 The stick up of the monitoring well is approximately 0.6 m agl. A metal casing was installed above ground.
- Gravel pack was placed around the slotted portion of the pipe. Bentonite was placed above the gravel to surface level

Contractor: Speights Drilling Date Time 19/10/21 00:00 3.721 10.95 Equipment: Sonic Rig



→ The Power of Commitment